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3 **ORIGINAL ARTICLE**

4

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6 **Recovery of forest-floor vegetation after a wildfire in a *Picea mariana***
7 **forest**

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1 **Abstract** We aimed to detect the trajectories of forest-floor vegetation recovery in a *Picea*
2 *mariana* forest after wildfire. Since fire severity in boreal forests is expected to increase
3 due to changes in climate, we investigated the effects of ground-surface burned severity, a
4 surrogate to the overall fire severity, on the revegetation. We annually monitored
5 vegetation less than 1.3 m high in 80 1 m × 1 m quadrats at Poker Flat Research Range
6 (65°12'N, 147°46'W, 650 m a.s.l.) near Fairbanks, interior Alaska, where a large wildfire
7 occurred in the summer of 2004, from 2005 to 2009. *Sphagnum* mosses were predominant
8 on the unburned ground surface. In total, 66% of the ground surface was burned
9 completely by the wildfire. Total plant cover increased from 48% in 2005 to 83% in 2009.
10 The increase was derived mostly by the vegetative reproduction of shrubs on the unburned
11 surface, and by the immigration of non-*Sphagnum* mosses and deciduous trees on the burned
12 surface. Deciduous trees, which had not been established before the wildfire, colonized
13 only on the burned surface and grew faster than *P. mariana*. Although species richness
14 decreased with increasing slope gradient, these deciduous trees established even on steep
15 slopes. The wildfire that completely burned the ground surface distorted the revegetation,
16 particularly, on steep slopes. Therefore, restoration of the *Sphagnum* surface was
17 prerequisite when the severe wildfire occurred, because *Sphagnum* cover seemed to be
18 difficult to return to predominance in the short term.

19

20 **Key words:** Burned ground surface · Deciduous trees · Mosses · Revegetation · Slope
21 gradient

22

1 Introduction

2 Lightning-caused wildfire is a key to maintain the regeneration of taiga dominated by *Picea*
3 *mariana* (P. Mill.) B.S.P. (black spruce) in boreal regions including interior Alaska
4 (Engelmark 1999). Lightning has led to crown fire, which moderately burned the forest
5 floor providing safe sites for *P. mariana* seedlings (Greene et al. 2004; Jayen et al. 2006).
6 *P. mariana* produces a semi-serotinous cone that releases most of its seeds after wildfire, and
7 the seedlings establish themselves steadily on incompletely-burned ground surface (Ilisson
8 and Chen 2009). In recent years, however, wildfires have been increasing fuel
9 consumption above and below the ground surface in the boreal forest zones of the Northern
10 Hemisphere, because of dry lightning (Kashischke and Turetsky 2006, Johnstone et al. 2011).
11 Climate projections suggest that these extreme disturbances will increase during this century
12 (Flannigan et al. 2000; Anisimov et al. 2007).

13 Species recovering after wildfire are often divided into two types: seeders that recover
14 by sexual reproduction and sprouters that recover by vegetative reproduction. Revegetation
15 patterns after wildfire differ with fire severity, due to changes in the contributions of seeders
16 and sprouters (Santana et al. 2012). When fire severity is low, i.e., the ground surface is
17 incompletely burned, sprouters contribute more to revegetation, and *vice versa* (Schimmel
18 and Granstrom 1996, Gurvich et al. 2005). Consequently, vegetation recovery is delayed or
19 altered when disturbance exceeds the threshold of the resilience of sprouters (Dale et al.
20 2001; Johnstone and Chapin 2006). Delayed recovery induces increase in active layer
21 depth in permafrost zones (Burn 1998; Tsuyuzaki et al. 2009), and accelerates positive
22 feedback on global warming (Kaplan and New 2006). Severe wildfire burning of ground
23 surfaces may accelerate establishment of species that did not exist pre-fire and reorganize the
24 flora.

25 Physical site characteristics, such as aspect, elevation and slope gradient, affect
26 regeneration patterns after disturbances (Chapin et al. 2006). Permafrost distribution is
27 related with the presence of *P. mariana*-dominated forests in discontinuous permafrost zones

1 (Davis 2001; Hollingsworth et al. 2006), and is decreased in post-fire vegetation due to
2 decreasing albedo and increasing soil temperature until the vegetation recovers well
3 (Chambers et al. 2005; Tsuyuzaki et al. 2009). Therefore, the paces and patterns of
4 revegetation are a key to sustain permafrost and to conserve the ecosystems. Since
5 large-scale disturbances, including burn-out wildfire, delay or alter the revegetation more
6 than small-scale ones such as crown fire (Dale et al. 2001; Rydgren et al. 2004), we
7 hypothesized that burned-out wildfire was associated with shifts in species composition by
8 removing the ground-surface plant cover and by species immigrating from the external
9 environments. We had a chance to obtain revegetation patterns after various-scale wildfires,
10 including both crown and burn-out fires, coded as “Boundary Fire” that occurred in interior
11 Alaska in the summer of 2004, and reported the first five years of monitoring of forest-floor
12 vegetation after the wildfire.

13

14 **Materials and methods**

15 *Study area*

16 The study site is located at Poker Flat Research Range, approximately 50 km north of
17 Fairbanks, Alaska. It is a scientific rocket launching facility owned by the University of
18 Alaska Fairbanks. The region is located in the discontinuous permafrost zone. Annual
19 precipitation averaged 297.4 mm at Fairbanks from 1971 to 2000, the maximum monthly
20 mean air temperature was 23.0°C in June, and the minimum was -25.0°C in January (ACRC
21 2007). The Boundary Fire occurred in the region including Poker Flat from mid-June to
22 late August in 2004, and burned ca 217,000 ha of forests with various severities (Betts and
23 Jones 2009; Johnstone et al. 2011). Before the wildfire, the tree layer was dominated by *P.*
24 *mariana* (Tsuyuzaki et al. 2009) and the ground surface was covered with thick mosses
25 dominated by *Sphagnum* spp. (*S. capillifolium* (Ehrh.) Hedw., *S. subsecundum* Nees, and
26 others) and with *Hylocomium splendens* (Hedw.) Schimp. Since *Sphagnum* spp. often
27 establishes on wetter sites than *Hylocomium* in pre-burned forests (Heijmans et al. 2004;

1 Nichole and Yves 2006), *Sphagnum* was likely to remain more abundant than *Hylocomium*
2 on the un-burned surface after the wildfire.

3 4 *Sampling*

5 Sixteen 10 m × 10 m plots were established on a north slope of different fire severity,
6 ranging from 0% to 100%, before leaf flushing began in the spring of 2005, to detect
7 relationships between burned area and revegetation (Tsuyuzaki et al. 2009). Five 1 m × 1
8 m quadrats were randomly set up in each plot, and the percentage of burned area was
9 estimated visually in each quadrat. Burned ground surface was divided into two types in
10 each quadrat; complete burning that removed the organic layer and exposed soil, and
11 incompletely burning that left blackened organic matter such as peat. The measurements of
12 burned areas were conducted over the whole burned area, including complete and
13 incomplete burning, and completely-burned area. Monitoring continued annually from
14 2005 to 2009.

15 Percent cover was recorded for each plant taxa in a layer less than 1.3 m high on each
16 quadrat every summer. The vascular plant species were identified by the first author, and
17 the vouchers have been stored in SAPS. The mosses were sampled, identified by M.
18 Higuchi, NMNS, and the vouchers have been kept in TNS. The lichens were sampled, and
19 the vouchers have been kept in GSES, HU. The layers more than 1.3 m high had few
20 vascular plants except *P. mariana*, the density of which averaged 24 stems per plot.

21 Photos were taken toward the canopy on each quadrat at 1.3 m above the ground surface
22 by a fish-eye lens in the summers of 2005 and 2008. The photos were used to evaluate
23 canopy openness, using a freeware Gap Light Analyzer ver. 2.0 (Frazer et al. 1999).
24 Longitude and latitude on each quadrat and on the center of each plot were measured by
25 differential GPS receivers (StarBox SSII-51CPN-19, Amtechs, Tokyo) with antenna
26 (GPS-701, NovAtel, Calgary). Location, aspect and slope gradient on each quadrat were
27 measured by a laser level meter. Based on these measurements, the elevation on each

1 quadrat was calculated. Thermometers were established from 2 cm to the surface of
2 permafrost of which maximum was 150 cm in three sites in August 2007 at three sites of
3 which fire severities differed between none and completely burned. Thaw depth was
4 inspected by temperature profiles recorded at 1 hour intervals.

5 6 *Data analysis*

7 Diversity and evenness in each quadrat were calculated by Shannon-Wiener indices.
8 Generalized linear mixed-effects models (GLMM) were used to detect significant
9 environmental factors on species richness, diversity, evenness, and total plant cover. Four
10 response variables—total plant cover, species richness, diversity, and evenness—were
11 examined in the models. The distributions were assumed as Poisson for richness because
12 of discontinuous function, gamma distribution for diversity because of continuous function
13 without zero, and binomial for evenness and cover because of the ratio. The examined
14 explanatory variables were burned ratio, canopy openness, elevation, slope, aspect, and years
15 after wildfire on each plot. Plot locations, i.e., longitude and latitude, were used as random
16 effect, to reduce the effects of pseudo-replicated sampling designs.

17 Non-metric multi-dimensional scaling (NMDS) was applied to investigate relationships
18 between environmental factors and plant cover on each species, using all taxa. NMDS is an
19 ordination technique well suited for data that are nonnormal, or occur along arbitrary or
20 discontinuous scales and is considered the most effective ordination method for ecological
21 community data (McCune and Grace 2002). NMDS does not assume a unimodal model of
22 species responses to the environment, and allows for the possibility that the community
23 variations are related to unmeasured environmental variables. Therefore, community
24 patterns are retained, regardless of what environmental variables were measured. In
25 addition, multi-response permutation procedures (MRPP) of 999 permutations were used to
26 examine significant differences between groups of sampling units (Reich et al. 2001).

27 Since the presence or absence *Sphagnum* mat greatly influence revegetation and/or

1 carbon cycle on and in the ground after wildfires in boreal ecosystems (Greene et al. 2004;
2 Shetler et al. 2008; Whinam et al. 2010), two similarity analyses were compared using
3 Jaccard similarity index to inspect the habitat preferences, in particular, to *Sphagnum* mat, of
4 species that showed high cover and/or frequency (hereafter, i.e., dominant species) in
5 relation to *Sphagnum*. In the first analysis, percentage similarities between *Sphagnum*
6 *fuscum* (Schimp.) Klinggr. and each dominant species or each lifeform were calculated. In
7 the second analysis, the similarities were calculated between non-*Sphagnum* cover and the
8 species or lifeform to investigate the preferences of species to ground surface not occupied
9 by *Sphagnum*. On the similarity between *Sphagnum* mat and moss lifeform, *Sphagnum*
10 species were excluded from the lifeform. The non-*Sphagnum* surface was mostly created
11 by burning in the first year (i.e., 2005). Therefore, non-*Sphagnum* cover was a surrogate
12 for the burned area in 2005. All statistical analyses were conducted with the statistical
13 program R (ver. 2.10.1) (R Development Core Team, 2010).

14

15 **Results**

16 *Initial vegetation patterns and the environments*

17 There were 29 vascular plant species (27 seed plants and 2 ferns), and over 13 non-vascular
18 plant species, i.e., mosses and lichens, recorded in the 80 quadrats. The two ferns were
19 *Equisetum silvaticum* L. and *Lycopodium annotinum* L. Of the mosses and lichens, *S.*
20 *fuscum*, which occupied 29.3% of the unburned surface in 2005. The vascular plants
21 consisted of 13 herbs (forbs, grasses and ferns), 12 shrubs, and 4 trees. Visual observation
22 confirmed that the shrubs survived throughout the wildfire and recovered mostly by
23 vegetative reproduction, i.e., sprouters. Of the four tree taxa, *P. mariana* was evergreen,
24 and the others (*Betula neoalaskana* Sarg., *Populus tremuloides* Michx., and *Salix* spp.) were
25 deciduous. *P. mariana* survived through the fire when they were rooted in *Sphagnum*.
26 Therefore, the cover was high in 2005 (0.6%) and was derived mostly from surviving
27 saplings. In contrast, the most of all broad-leaved trees were regenerated by seedlings.

1 Canopy openness ranged from 57% to 94%. The burned area averaged 66%, ranging
2 from 0% to 100%. *Sphagnum* cover decreased annually, probably due to the delayed
3 effects of burning, such as desiccation. Thaw depth increased with increasing burned
4 surface on plot level from 50 cm to 330 cm in August 2007. The total number of species in
5 all the quadrats ranged from 39 to 41 for five years, and did not differ largely, indicating that
6 the species composition was fixed soon after the wildfire. Plant cover was negatively
7 correlated to burned area (GLMM, $P < 0.01$).

8 Slope gradient was negatively correlated to richness ($P < 0.01$), showing that steep
9 slopes restricted species richness. Species diversity and evenness ranged from 1.24 to 1.61
10 and from 0.56 to 0.68, respectively, and were not correlated to any examined environmental
11 variables. Canopy openness and elevation were not related to these four parameters on
12 plant community structure ($P > 0.01$).

13 14 *Temporal changes in vegetation*

15 Total plant cover increased from $60\% \pm 50$ (mean with standard deviation) in 2005 to 108%
16 ± 38 in 2009 (GLMM, $P < 0.01$). Similarly, species richness ranging from 9.8 to 11.3 was
17 negatively correlated to burned area ($P < 0.01$) and positively correlated to year ($P < 0.01$).
18 A sedge, *Carex bigelowii* Torr., and a grass, *Calamagrostis canadensis* (Michx.) Beauv.,
19 gradually increased their cover from 0.9–1.7% to 3.0–7.7% over the five years. A forb,
20 *Epilobium angustifolium* L., had a peak of cover ($3.0\% \pm 9.0$) two years after the wildfire
21 and then gradually decreased to $1.9\% \pm 3.4$. Ferns were infrequent in the quadrats. All of
22 the dominant shrub species gradually increased their cover from 0.3–4.3% in 2005 to
23 0.9–11.3% in 2009. Shrubs, such as *Ledum groenlandicum* L., *Vaccinium vitis-idaea* L.,
24 *Vaccinium uliginosum* L., *Betula nana* L., and *Oxycoccus microcarpus* Turcz., had high
25 cover even soon after the wildfire.

26 All of the trees gradually increased in cover across time. The cover of *P. mariana*
27 increased slowly, and was 0.9% even in 2009. In contrast, *B. neoalaskana* and *P.*

1 *tremuloides* established from seeds, i.e., seeders, and thus the initial cover was low (less than
2 0.1%). Increase in cover was, however, faster for the two deciduous trees than for *P.*
3 *mariana*. In 2009, cover of *P. tremuloides* and *B. neoalaskana* reached 1.0% and 0.7%,
4 respectively. Of the mosses and lichens, *S. fuscum* was dominant but decreased in cover
5 annually to 21.1%. *Polytrichum commune* Hedw. and *Ceratodon purpureus* Hedw.
6 increased in cover annually reaching 6% and 15%, respectively, in 2009. *Ceratodon*
7 *purpureus* was established in 80% of quadrats in 2009. These two mosses, *P. commune* and
8 *C. purpureus* were uncommon on unburned surfaces.

9 Coefficients of determination on NMDS indicated that all examined variables
10 significantly explained the ordination patterns (test of random data permutations, $P < 0.01$).
11 As axis I explained 52% of variance, axis II 26%, and axis III 14%, the first two axes were
12 examined. The stress was 0.191. Of these, fire severity expressed by burned surface was
13 the prime factor determining vegetation structures (Fig. 1). Burned area and canopy
14 openness were both negatively correlated with axis I, showing that fire severity is a strong
15 factor in determining post-fire species composition. The other three site factors—elevation,
16 aspect and slope gradient—were related more to axis II, showing that burn severity was not
17 greatly related to the site characteristics. Slope gradient was related to axis II more than
18 aspect and elevation. Year had the weakest effects of the examined variables on vegetation
19 structures, showing that revegetation pathways were not unique. MRPP also supported
20 these results, i.e., these examined variables explained significantly different between the
21 sampled groups, including continuous variables, at $P < 0.01$ (A ranging from 0.022 to 0.718,
22 $\delta < 61.2$ that were less than expected $\delta = 62.58$, and $n = 400$).

23 Species scores on NMDS showed distinct patterns between the four life forms—herbs,
24 shrubs, trees, and mosses (Fig. 1). Shrubs had high scores on axis I close to *Sphagnum*
25 *fuscum*, while non-*Sphagnum* mosses, *P. commune* and *C. purpureus*, had low scores far
26 from *S. fuscum* on axis I. These patterns indicated that the shrubs established less with the
27 non-*Sphagnum* mosses. Deciduous trees scored low and clustered to each other on axis I

1 and axis II but *P. mariana* scored high on axis I, showing that the two deciduous trees
2 established least with *P. mariana*. The scores of two non-*Sphagnum* mosses came close to
3 those of deciduous trees but were slightly higher on axis I, showing that the mosses
4 established primarily on burned surface but less so on the unburned surface. Although
5 species richness decreased with increasing slope gradient, two deciduous trees (*P.*
6 *tremuloides* and *B. neoalaskana*) and a moss (*P. commune*) established more on steep slopes,
7 as shown by the low scores on axis II correlated to slope gradient.

8 High quadrat scores on axis I meant that quadrats were burned less, and *vice versa* (Fig.
9 2). The annual fluctuations of quadrat scores on NMDS became larger on more burned
10 areas indicating that vegetation structures changed faster on burned surfaces with increasing
11 fire severity. In addition, the scores on severely burned surfaces fluctuated along axis II
12 rather than axis I, showing that the vegetation changes did not go toward less-burned
13 vegetation. The major causes of the fluctuation patterns were derived from the colonization
14 of two deciduous trees (*P. tremuloides* and *B. neoalaskana*, a herb, *E. angustifolium*) and two
15 mosses (*P. commune* and *C. purpureus*), all of which established more on burned surfaces
16 (Fig. 1).

17 *Habitat preferences*

18 The patterns of yearly fluctuations on the two similarities were different between burned and
19 unburned surfaces (Fig. 3). The similarity of each lifeform were stable on unburned
20 *Sphagnum* surface for the five years, as compared with the similarity on burned surface.
21 These implied that the vegetation structures temporally changed more on burned surfaces
22 than on unburned surfaces. The similarity of trees increased gradually on both unburned
23 and burned surfaces, indicating that the trees increased their cover with time on both the
24 surfaces.
25

26 The patterns were not synchronized between the three herbs (Fig. 3). The similarity of
27 *E. angustifolium* to *Sphagnum* cover was extremely low and to non-*Sphagnum* cover was

1 high, showing that *E. angustifolium* established least on the unburned surface. In contrast,
2 the similarity indicated *C. canadensis* recovered rapidly not only on *Sphagnum* surface but
3 also on non-*Sphagnum* surface. *C. bigelowii* established on the *Sphagnum* surface more
4 than on the non-*Sphagnum* surface, and the annual changes in the two similarities were
5 small.

6 For all the dominant shrubs, the similarities to *Sphagnum* cover were high but to
7 non-*Sphagnum* cover were low, showing that shrubs established well with *Sphagnum* as
8 indicated by NMDS (Fig. 1).

9 *P. mariana* showed contrast patterns on the similarities on the *Sphagnum* and
10 non-*Sphagnum* surfaces to the two deciduous species (Fig. 3). *P. mariana* showed higher
11 similarity to the *Sphagnum* mat than any other tree species, and did not show low similarity
12 to the non-*Sphagnum* mat. Although *P. mariana* established more on unburned surfaces,
13 that did not mean *P. mariana* did not establish on burned surfaces. In contrast, two
14 deciduous trees established least with *Sphagnum* and increased annually the similarities to
15 non-*Sphagnum*, showing that these species gradually in cover on the non-*Sphagnum* surface.
16

17 **Discussion**

18 *Vegetation patterns along gradients of fire severity*

19 NMDS on quadrats (Fig. 1) indicated that revegetation patterns on the floor differed greatly
20 between burned and unburned surfaces since the recovery on unburned surfaces was
21 dependent on regeneration and that on burned surfaces was on colonization. Species
22 producing wind-dispersed seeds contribute more to revegetation after severer wildfire
23 (Johnson and Miyanishi 2007). Fire severity determines community composition derived
24 from differences in regeneration strategy and root depth among species (Hollingworth et al.
25 2013).

26 The similarities between ground-surface combustion and plants indicated as follows
27 (Fig. 3): *P. mariana* of the tree species produced semi-serotinous wind-dispersed seeds, but

1 established more on the unburned surface probably because of a habitat preference that is
2 adapted to crown fire. Deciduous trees needed high severity sites where most *Sphagnum*
3 carpets were removed for their establishments, while *P. mariana* trees colonized in both high
4 and low severity sites because they were either able to survive fire on thick *Sphagnum* and
5 germinated on burned and unburned surfaces. When post-fire residual organic matter is
6 more than 2 cm in thickness, vegetation composition diverges in *P. mariana* forest of eastern
7 Canada (Siegwart Collier and Mullik 2010).

8 Species producing wind-dispersed seeds or spores established steadily when suitable
9 habitats were provided by burning the *Sphagnum* surface; viz. all the deciduous trees
10 produced wind-dispersed seeds, and a perennial herb, *Epilobium angustifolium*, did, too.
11 Mosses produced spores that should be dispersed by wind. Deciduous trees do not recruit
12 seedlings in thick organic mats but outcompete with conifers on mineral soils after wildfire
13 (Johnstone and Chapin 2006). There should be no safe sites for deciduous trees before *P.*
14 *mariana* forests were burned.

15

16 *Vegetation patterns along gradients of topography*

17 Species richness decreased with increasing slope gradients. In addition, NMDS on species
18 showed that revegetation on steep slopes was altered by the establishment of deciduous trees
19 and non-*Sphagnum* mosses (Fig. 2). Wind-dispersed seeds often accumulate in rough
20 microtopography provided by live plants, litter and coarse-textured soil, all of which
21 function as seed traps (Koyama and Tsuyuzaki 2010). Deciduous trees disperse numerous
22 seeds to treeless areas after wildfire (Landhausser and Wein 1993). The wind-dispersed
23 seeds of broad-leaf trees should be captured by coarse and/or concave ground surfaces, even
24 on steep slopes. Revegetation after wildfire is determined not only by fire severity but also
25 pre-disturbance conditions, including topography, and is not interpreted by a vegetation
26 gradient related to time after disturbance (Rydgren et al. 2004).

27

1 *Lifeforms in relation to seeders vs sprouters*

2 All the deciduous trees were seeders at Poker Flat after the wildfire. In contrast, shrubs,
3 most of which were sprouters, increased in cover annually, depending on the unburned
4 surface. The clonal shrubs (e.g., *Vaccinium uliginosum* and *V. myrtilloides*) rarely produced
5 seedlings on the *Sphagnum* mat (Eriksson 1989; Moola and Mallik 1998). Perennial
6 grasses and sedges also reproduced vegetatively on unburned surface. Sprouters, such as
7 perennial sedges, grasses, and shrubs are common when wildfire is not severe in a
8 mixed-evergreen forest, Oregon (Donato et al. 2009). Revegetation is progressed by
9 species in soil bud-bank (sprouters) and seed-bank after weak fire that consumes
10 relatively-shallow moss layer in a boreal Swedish forest, while the revegetation is conducted
11 by seeders after deep-burning fire (Schimmel and Granstrom 1996).

12 Such differences in regeneration strategies between species or lifeforms determine the
13 plant community structures and should be used for predicting successional sere with
14 different fire severities.

15
16 *Temporal changes in vegetation with reference to habitat preferences*

17 *Sphagnum* gradually decreased in cover for five years and did not re-colonize on the burned
18 surface, due probably to post-fire stresses such as by desiccation. Albedo was reduced
19 when the plant cover burned and did not return to the pre-fire status until plant cover became
20 high (Tsuyuzaki et al. 2009). Low albedo leads to high ground-surface temperature and
21 deep thaw depth (Jorgenson et al. 2001). These changes induce drought stresses for plants,
22 particularly *Sphagnum* mosses, which require more water than vascular plants for the short
23 term. This is because they develop shallow belowground organs for water acquisition and
24 have no special organs for water transportation (Schouwenaars and Gosen 2007; Thompson
25 and Waddington 2008). *Sphagnum* is likely to recover slower than vascular plants,
26 particularly when ground surface is completely consumed by wildfire.

27 The increase in moss cover was mostly derived from the pioneer mosses, represented by

1 *Polytrichum strictum* and *Ceratodon purpureus*, on burned surfaces. Mineral soils exposed
2 after disturbances, including severe wildfires, were covered with the pioneer mosses, and the
3 vegetation was totally dissimilar to the pre-fire vegetation (Bernhardt et al. 2011). The
4 aboveground biomass and productivity of non-vascular plants are lower on drier sites after a
5 wildfire in interior Alaska, in particular, where *Ceratodon* and/or *Polytrichum* were
6 dominant (Mack et al. 2008). Deciduous trees established with *P. commune* and *C.*
7 *purpureus* on the burned surface. Since these two mosses established well on burned
8 surfaces, they should induce the alteration of tree regeneration. The revegetation did not go
9 towards *P. mariana* forest directly when wildfire burns out the ground surface.

10 Forest revegetation seems to be delayed by severe burn (Dale et al. 2001).
11 Furthermore, the complete removal of moss and organic matter promoted the colonization of
12 non-*Sphagnum* mosses and deciduous trees. In conclusion, seeders represented by
13 broad-leaved trees have a great role on revegetation after severe wildfire in *P. mariana*
14 forests that completely burns out not only *Sphagnum* mat but also organic layer.

15
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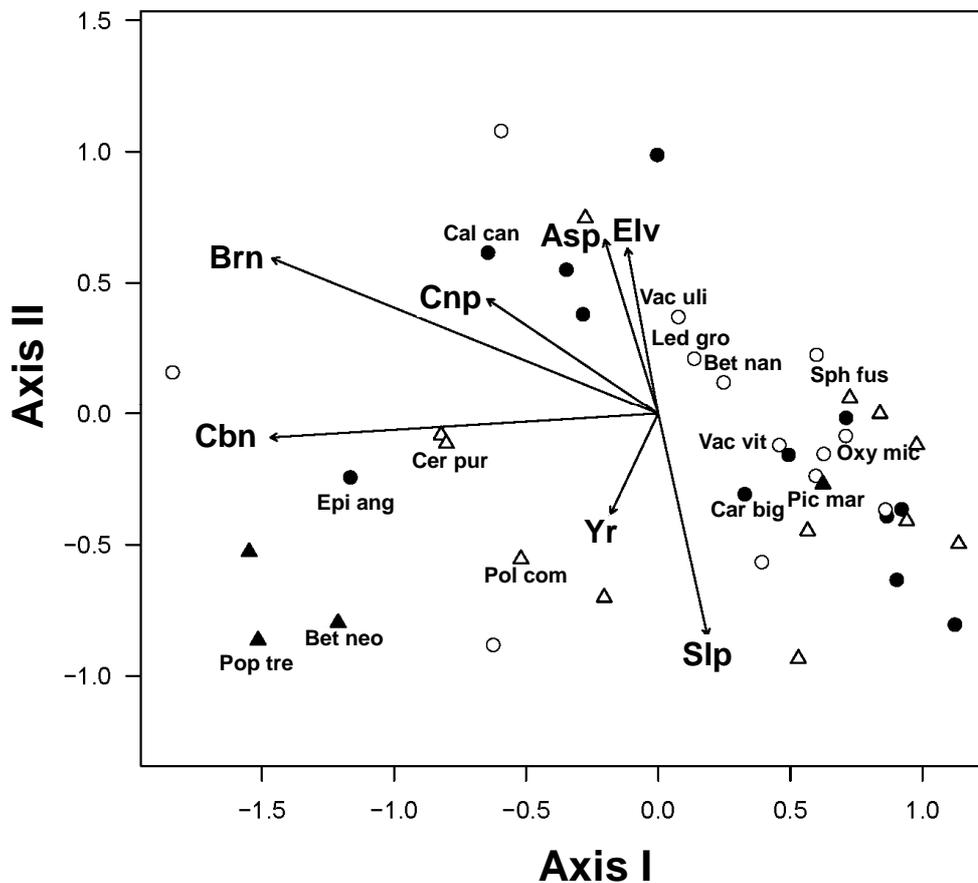
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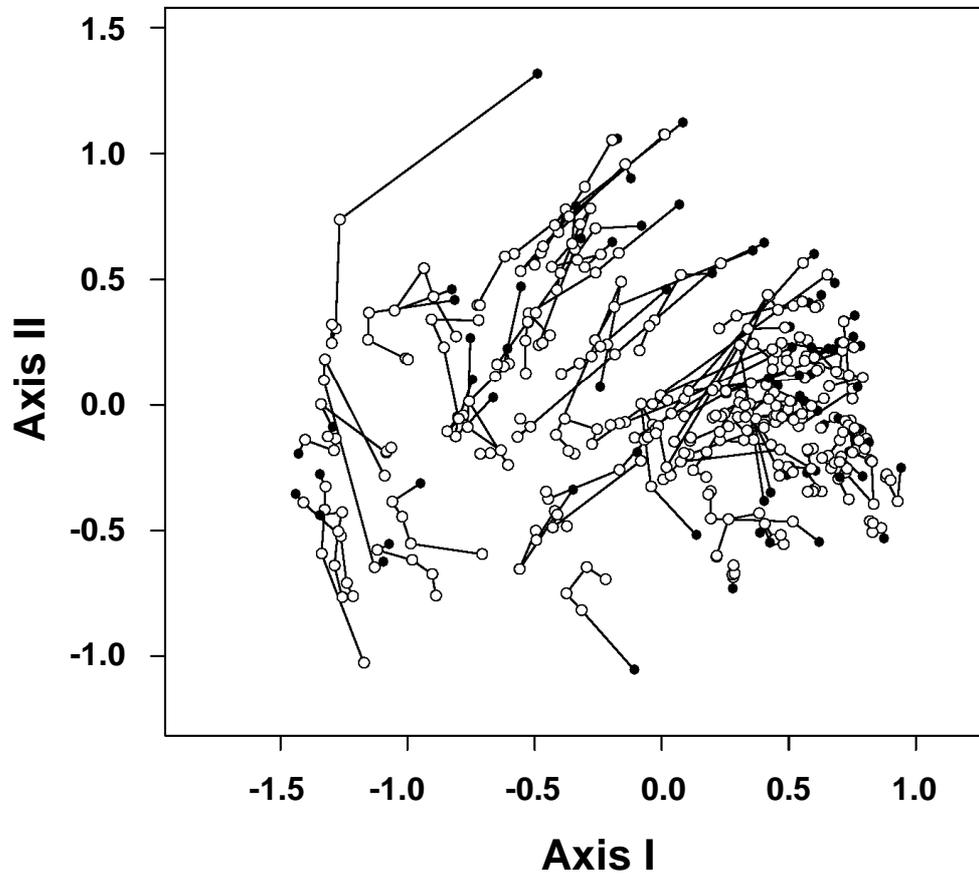
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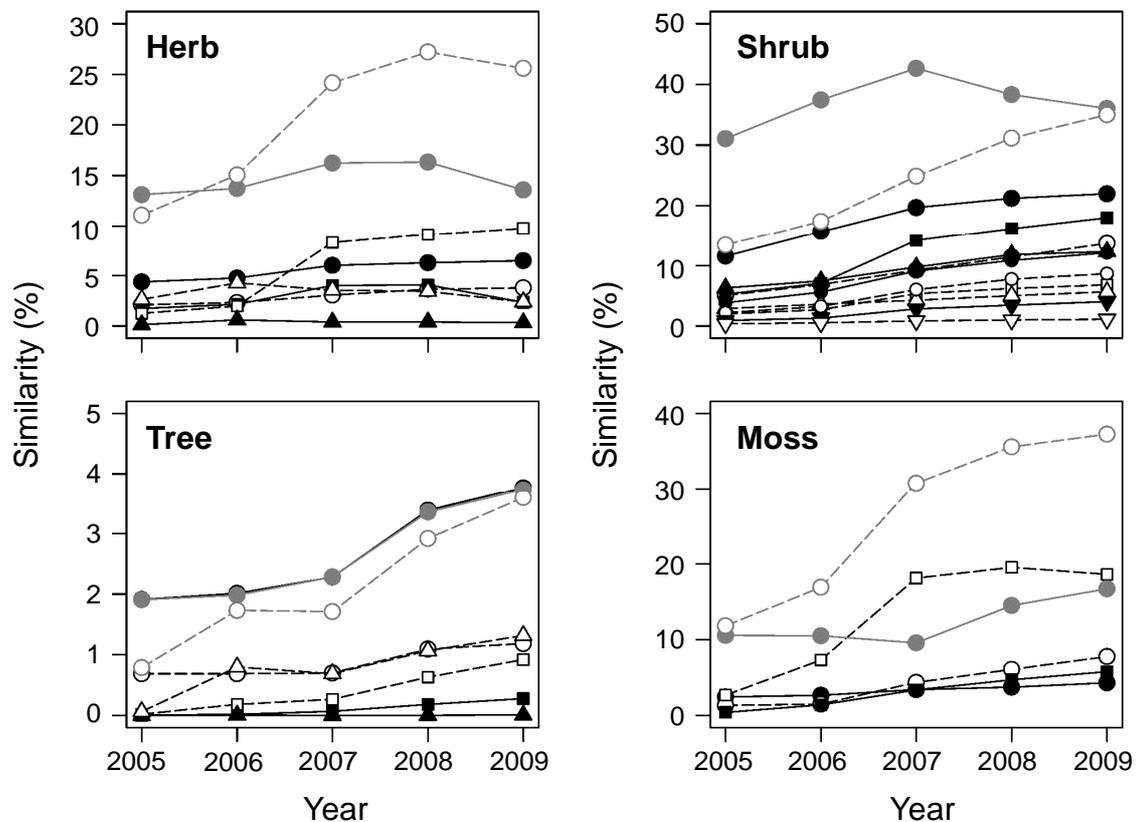
Fig. 1 NMDS ordination diagram on species scores in 400 quadrats (80 × 5 years) surveyed from 2005 to 2009 after the 2004 wildfire. Life forms: closed circles = herbs and ferns (13 taxa), open circles = shrubs (12), closed triangles = trees (4), and open triangles = mosses and lichens (13). Species codes explained in the text: Car big = *Carex bigelowii*, Cal can = *Calamagrostis canadensis*, Epi ang = *Epilobium angustifolium*, Led gro = *Ledum groenlandicum*, Vac vit = *Vaccinium vitis-idaea*, Vac uli = *Vaccinium uliginosum*, Bet nan = *Betula nana*, Oxy mic = *Oxycoccus microcarpus*, Pic mar = *Picea mariana*, Bet neo = *Betula neoalaskana*, Pop tre = *Populus tremuloides*, Pol for = *Polytrichum commune*, Sph fus = *Sphagnum fuscum*, and Cer pur = *Ceratodon purpureus*. Environmental factors: Yr = years after wildfire, Brn: burned area (%), Cbn: completely burned area (%) where the soils were exposed. Cnp = canopy openness (%), Elv = elevation (m), Slp = slope gradient ($^{\circ}$), and Asp = aspect ($^{\circ}$).

17



6 **Fig. 2** NMDS ordination diagram on quadrat scores in 400 quadrats (80×5 years)
7 surveyed for five years after the 2004 wildfire. The connected lines show the same
8 quadrats surveyed from 2005 to 2009. NMDS scores at the first survey, i.e., in 2005,
9 are shown by solid circles, and the others are by open circles. The arrows of
10 environmental factors are not shown.

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5 **Fig. 3** Percentage similarities between *Sphagnum fuscum* surface and species or
 6 lifeform (shown by solid lines with closed symbols), and between non-*Sphagnum*
 7 surface and species (by interrupted lines with open symbols). Gray symbols and lines
 8 indicate the lifeforms. Herb: ●/○ = *Carex bigelowii*, ■/□ = *Calamagrostis canadensis*,
 9 ▲/△ = *Epilobium angustifolium*, Shrub: ●/○ = *Ledum groenlandicum*, ■/□ = *Vaccinium*
 10 *vitis-idaea*, ▲/△ = *Vaccinium uliginosum*, ◆/◇ = *Betula nana*, ▼/▽ = *Oxycoccus*
 11 *microcarpus*, Tree: ●/○ = *Picea mariana*, ■/□ = *Betula neoalaskana*, ▲/△ = *Populus*
 12 *tremuloides*, Moss: ●/○ = *Polytrichum commune*, ■/□ = *Ceratodon purpureus*. Note
 13 that the scales of y-axis are different between life forms and that the similarities of *P.*
 14 *mariana* and tree on unburned surface are overlapped.

15