<table>
<thead>
<tr>
<th>項目</th>
<th>内容</th>
</tr>
</thead>
<tbody>
<tr>
<td>Title</td>
<td>提案する手法に基づくモデル解析</td>
</tr>
<tr>
<td>Author(s)</td>
<td>Funamizu, Naoyuki; Takakuwa, Tetsuo</td>
</tr>
<tr>
<td>Citation</td>
<td>衛生工学シンポジウム論文集</td>
</tr>
<tr>
<td>Issue Date</td>
<td>1994-11-01</td>
</tr>
<tr>
<td>Doc URL</td>
<td><a href="http://hdl.handle.net/2115/7581">http://hdl.handle.net/2115/7581</a></td>
</tr>
<tr>
<td>Type</td>
<td>bulletin</td>
</tr>
<tr>
<td>Note</td>
<td>第2回衛生工学シンポジウム(平成6年11月10日(木・11日(金) 北海道大学学術交流会館) モデル解析</td>
</tr>
<tr>
<td>File Information</td>
<td>2-2-2_p42-47.pdf</td>
</tr>
</tbody>
</table>
Simulation of the operating conditions of the municipal wastewater treatment plant at low temperatures using a model that includes the IAWPRC activated sludge model

Naoyuki Funamizu and Tetsuo Takakuwa

Department of Sanitary and Environmental Engineering, Hokkaido University

INTRODUCTION

One of the most suitable plans regarding the sewage system is for its use to transport and melt snow in snowy regions. Since this would cause a drop in sewage temperature, adequate conditions of operating biological wastewater treatment plants at low temperatures should be simulated. To determine the influence of melting and transporting snow by sewer, plant performance should be measured by the effluent water quality, and the quality and quantity of sludge produced from a plant, because the plant treats both water and sludge. This simulation requires the modeling of all unit processes in the municipal wastewater treatment plant: water treatment processes including primary and final clarification and aeration basin and sludge treatment processes.

The IAWPRC activated sludge model (Henze et al., 1987) has been frequently used to predict the performance of water treatment processes, especially of biological reactions. Singrist and Tschui (1992) calibrated the activated sludge model using municipal wastewater treatment plants in Switzerland. It is recognized that the IAWPRC model has applicability as a prognostic tool (Pedersen and Sinkjaer, 1992). Lesouef et al. (1992) showed the on-site calibration techniques, verified the model on the full-scale plant, and proposed optimum reactor configurations for nitrogen removal. Therefore, the IAWPRC activated sludge model has been employed as the model of biological reactions in this work.

As for the model of unit processes other than biological ones, Dupont and Henze (1992) proposed the secondary clarifier model combined with the activated sludge model. It seems that few have attempted to model the entire water and sludge flow of a full-scale plant. The objectives of this paper are
- to verify the applicability of the IAWPRC model to the aeration tank of the plant at low temperatures using the pilot plant data,
- to show a simple method of modeling a sludge treatment process based on plant operation data,
- to draw operation maps of a full-scale plant in Sapporo at the presumed low temperature.

CALIBRATION OF THE IAWPRC MODEL USING THE PILOT PLANT DATA

Pilot plant experiments The pilot plant had a fully mixed aeration tank (1.2 m$^3$) and a secondary clarifier, and it treated the primary effluent (2.6 L/min) of the Sapporo Sosei treatment plant. Experiments were performed at 5, 6.5, 8 and 10°C. The measured variables were BOD, COD, nitrogen compound of influent and effluent of the plant, volatile and non-volatile suspended solids of mixed liquor and return sludge at steady state.

Calibration of the IAWPRC model The model was calibrated using experimental data from operation at 10°C. Only two parameters, the yield for the heterotrophic biomass $Y_H$ and the maximum specific growth rate for the autotrophic biomass $\mu_A$, were adjusted from the default values (Henze et al., 1987). $Y_H$ was 0.715 g cell COD formed (g COD oxidized)$^{-1}$ (default value 0.67). $\mu_A$ was 0.29 day$^{-1}$ (default 0.3). RUN1 in Table 1 shows the computed and observed results. Then, temperature coefficients in the reaction equation were estimated using the data from operation at 5°C. Reducing the temperature coefficient for the growth of autotrophic biomass from 1.103 (default) to 1.044 yielded the results in RUN2 in Table 1.

Verification of calibrated parameters using data from operation at other temperatures Comparisons of the simulated results with data from operation at other temperatures (RUN3 and RUN4 in Table 1) showed that the IAWPRC model could be used for expressing biological reactions in the aeration basin at low temperatures.
TABLE 1 COMPARISON OF COMPUTED RESULTS WITH PILOT PLANT DATA

<table>
<thead>
<tr>
<th></th>
<th>RUN1</th>
<th>RUN2</th>
<th>RUN3</th>
<th>RUN4</th>
</tr>
</thead>
<tbody>
<tr>
<td>Temp. °C</td>
<td>10.3</td>
<td>4.9</td>
<td>7.7</td>
<td>6.5</td>
</tr>
<tr>
<td>R_s1 (%)</td>
<td>30</td>
<td>30</td>
<td>33</td>
<td>31</td>
</tr>
<tr>
<td>R_s2 (%)</td>
<td>1.28</td>
<td>1.08</td>
<td>2.02</td>
<td>0.77</td>
</tr>
</tbody>
</table>

![Diagram of sludge and water flows in the investigated plant](image)

**Fig. 1.** Sludge and water flows in the investigated plant

**SIMULATION MODEL FOR THE FULL-SCALE DOMESTIC WASTEWATER TREATMENT PLANT**

**Description of the investigated treatment plant**

The plant (Fig.1) treats wastewater of about 93500 m³/day during dry weather in winter. This plant has two lanes in its water treatment system, and one sludge treatment system. Each water treatment lane consists of primary clarifiers, aeration basins and final clarifiers. Nitrification occurs partly in lane 1 and fully in lane 2 in winter. The sludge treatment system has gravity thickeners and pressure filtration units. 80% of cake from this system are incinerated. Supernatant of thickeners and filtrate is returned to the water treatment system. All water and sludge flows in Fig.1 were modeled in the simulation.

**Primary clarifier**

A Voshel & Sak type formula (Voshel and Sak, 1968) was used. This empirical model shows that solid removal efficiency r is directly proportional to a power function of the influent solid concentration X_in and inversely proportional to a power function of the overflow rate L_p. Regression analysis using plant data yielded the following relationships:

\[ r = 0.133 X_{in}^{0.212} L_p^{-0.0265} \quad (Lane 1), \quad r = 0.200 X_{in}^{0.290} L_p^{-0.0807} \quad (Lane 2) \]

Figure 2 shows that Eq.(1) can explain the solid removal efficiency of the primary clarifier in the investigated plant. The solid concentration of the sludge from the primary clarifier was estimated by the following flow volume and solid mass balance equation at steady state:
Fig. 2. Comparison of calculated results of solid removal efficiency with plant data

\[ Q_{in} = Q_e + Q_u \]
\[ Q_{in}X_{in} = Q_eX_e + Q_uX_u \]  
(2)

where \( Q_{in} \), \( Q_e \) and \( Q_u \) are flow rates of influent, effluent and sludge, respectively, and \( X_{in}, X_e \) and \( X_u \) are concentrations of suspended solids.

In Eq. (2), flow rate of sludge \( Q_u \) was fixed at 3% of influent flow rate in the simulation, which was the average value of the plant operation data.

Aeration Basin Biological reactions in the aeration basin were described by the IAWPRC model. Parameters in the model had values estimated by the pilot plant data. The flow scheme of the basin was approximate, with 4 complete mixing tanks in series.

Final Clarifier The final clarifier had two functions: clarification and thickening. The models of clarification proposed so far relate effluent concentration of suspended solids to both the flow rate and the concentration of suspended solids of the influent. A typical empirical model has the following equation (Chapman, 1984):

\[ X_e = a_1 + a_2X_{in} + a_3L_p \]  
(3)

The application of Eq. (3) to the plant data did not yield a close correlation. Since the plant data varied randomly, we set a constant value, 7.5 (g/m³), for concentration of suspended solids of the effluent in the simulation. For thickening, the balance equations of flow volume and solid mass at steady state were employed in the same manner as the model of primary clarifier.

Thickener The outputs of the thickener are supernatant and concentrated sludge. The solid recovery percentage, \( R_{TH, mass} = \frac{[concentrated solid mass]}{[input solid mass]} \), was adapted to measure of thickening performance. The mean value in winter, about 80%, was used in the simulation. The operational condition was specified by the withdrawal percentage \( R_{TH,Q} = \frac{[concentrated sludge flow rate]}{[input flow rate]} \). Plant data showed that \( R_{TH,Q} \) had a constant value of about 10%. The balance equations of flow and solid mass:

\[ Q_e = (1 - R_{TH,Q})Q_{in} \]
\[ X_eQ_e = (1 - R_{TH,max})X_{in}Q_{in} \]  
(4)

yielded the flow rate \( Q_e \) and solids mass concentration \( X_e \) in supernatant. The concentration of dissolved matter in supernatant was specified by the plant data.

Besides computation by these balance equations, the minimum required cross-sectional area \( A_{req} \) was calculated by limiting flux theory. By this area, one can examine the possibility of whether or not the desired concentrated sludge can be obtained by the thickening process in the simulation. Since this calculation requires a hindered settling velocity, the velocity and solids concentration relationship was specified by the settling test with the sludge from the pilot plant operating at 5°C. The analysis of the settling curve in Fig. 3 by Kynch’s (1952) theory yielded the hindered settling velocities. These
velocities were correlated by the logarithmic model, and the limiting flux $G_L$ was computed by the following equations (Baskin, Suidan, 1985):

$$V = V_0 \left( \frac{C}{C_0} \right)^n, \quad \frac{G_L}{V_0C_0} = (n-1) \left( \frac{n}{n-1} \right) \left( \frac{U}{V_0} \right) \left( \frac{C_0}{C} \right)^{n-1}$$

where $U$ is underflow velocity.

The settling test showed that the parameters in Eq.(5) were $n=4.0$, $V_0=5.0$ (cm/min), and $C_0=1.0$ (g/L), respectively. Solid mass and volume in withdrawal flow can be written as

$$A_{Req} G_L = X_{in} Q_{in} R_{TH,mass} \quad A_{Req} U = Q_{in} R_{TH,Q}$$

Substituting Eq.(5) in Eq.(6) and rearranging the resulting expressions yields

$$\frac{A_{Req}}{R_{TH,Q} Q_{in} V_0} = (n-1) \left( \frac{n}{n-1} \right) \left( \frac{X_{in}}{C_0} \right)^n \left( \frac{R_{TH,mass}}{R_{TH,Q}} \right)$$

Stokes law was assumed for the settling velocity and the temperature dependence of settling velocity was expressed by the viscosity $\mu$, that is, $V_0$ at $10^\circ C$, $V_0(t) = (\mu(10)/\mu(t))V_0(10)$.

**Pressure Filter**  
Coagulants in the plant are ferric chloride and lime. The average addition rates in winter were about 7% and 31%. There were few variations in the water content of cake and a constant value, 63%, was used in the simulation. The solid recovery percentages were nearly constant, about 90%. The heating value of cake was calculated by the experimental formula (Murakami et al., 1986);

$$H_C = (1-w/100)H_{lk} - 540w/100$$

$$H_{lk} = 0.93 \left( \frac{58.3X_{CAKE} - 193 - 353k_2 - 0.69k_1}{1 + k_2 - 0.03k_1} \right) - 33$$

where $H_C$ is heating value of cake (kcal/kg-cake), $w$ is water content of cake (%), $X_{CAKE}$ is organic matter content in cake (%), and $k_1$ and $k_2$ are respective addition rates of ferric chloride and lime.

This heating value of cake was the measure of burning characteristics in the simulation. Murakami et al. (1986) reported that the lower limit of heating value for combustion without fuel depended on the type of incinerator. In the simulation, 430kcal/kg-cake was used for the lower limit.

**Comparison of simulated results with plant data results**  
We simulated average performance in winter by steady state analysis and compared the results with plant data. Influent concentrations used in the simulation are summarized in Table 2. The comparison in Table 3 shows that the simulation model yielded the reasonable results.

**TABLE 2 INFLUENT CONCENTRATION USED IN SIMULATION (g/m³)**

<table>
<thead>
<tr>
<th>Readily biodegradable COD</th>
<th>Soluble inert COD</th>
<th>Slowly biodegradable COD</th>
<th>Particulate inert COD</th>
<th>Oxygen</th>
</tr>
</thead>
<tbody>
<tr>
<td>70</td>
<td>10</td>
<td>120</td>
<td>30</td>
<td>5</td>
</tr>
<tr>
<td>NH₄-N</td>
<td>NO₃-N</td>
<td>Soluble Org-N</td>
<td>Particulate Org-N</td>
<td>non-volatile SS</td>
</tr>
<tr>
<td>16.0</td>
<td>0.3</td>
<td>4</td>
<td>4</td>
<td>60</td>
</tr>
</tbody>
</table>

**PLANT OPERATION MAPS AT 4 AND 8°C**

It was expected that the melting and transporting snow by sewer would cause influent temperatures to fall to 4-8°C. The operation maps of the plant at 4 and 8°C were drawn by the simulation of the steady state. The operational variables were the sludge recycle rate ($R_s$), and the excess sludge withdrawal rate ($R_e$). The goal of operation was specified by the following four variables:
<table>
<thead>
<tr>
<th></th>
<th>LANE 1</th>
<th>LANE 2</th>
</tr>
</thead>
<tbody>
<tr>
<td>flow rate</td>
<td>42700 m³/day</td>
<td>50800 m³/day</td>
</tr>
<tr>
<td>Rₛ, Rₑ</td>
<td>26%, 1.3%</td>
<td>31%, 0.77%</td>
</tr>
<tr>
<td>BOD₃</td>
<td>3.4* 6.2</td>
<td>3.5 6.2</td>
</tr>
<tr>
<td>T-N</td>
<td>13.9 14.8</td>
<td>12.1 16.3</td>
</tr>
<tr>
<td>NH₄-N</td>
<td>6.3 5.4</td>
<td>0 0.6</td>
</tr>
<tr>
<td>NO₃-N</td>
<td>7.6 9.0</td>
<td>11.7 15.4</td>
</tr>
<tr>
<td>MLDO⁺</td>
<td>2.8 2.9</td>
<td>5.2 5.2</td>
</tr>
<tr>
<td>MLSS</td>
<td>1980 1660</td>
<td>2650 2450</td>
</tr>
<tr>
<td>MLVSS</td>
<td>72.0% 70.5%</td>
<td>71.0 71.3%</td>
</tr>
<tr>
<td>RSSS⁺</td>
<td>8050 7700</td>
<td>9560 10130</td>
</tr>
</tbody>
</table>

1. Rₑ = [sludge recycle flow rate]/[sewage flow rate into plant]  
2. Rₛ = [excess sludge flow rate]/[sewage flow rate into plant]  
3. BOD was measured after addition of ATU  
4. MLDO is DO in the aeration basin  
5. RSSS is suspended solids concentration in return sludge  
6. Unit not specified in table is g/m³

Fig. 4. Plant operation maps
We prepared four kinds of scenarios and searched the available operating conditions in each case:

S-1: Keep the present level of nitrate nitrogen concentration in effluent.
Neither thickener nor incineration plant has an excess capacity.

S-2: Keep the present level of nitrate nitrogen concentration in effluent.
Both thickener and incineration plant have 10% excess capacity.

S-3: Keep the present level of nitrate nitrogen concentration in effluent.
There is no limitation either in thickener or in incineration.

S-4: No limitation on nitrification.
Both thickener and incineration plant have 10% excess capacity.

In S-1, no available operating point was found at 4°C or at 8°C. In S-2, very few operating points were obtained only at 8°C. Figure 4 shows the operation maps of S-3 and S-4. In S-3, available operating points correspond to the small $R_e$, that is, long SRT operation. The operating conditions in S-4 at 8°C are in a wide range and include the present plant operating points. At 4°C in S-4, a large $R_e$ is required, and this might cause some troubles with the final clarifier. The difference between maps of S-3 and S-4 shows that it is impossible to keep nitrifying bacteria at the present level in the plant while preventing excess load for sludge treatment. This suggests that the enlargement of the sludge treatment system or that the addition of nitrifying-bacteria-holding apparatus to aeration basins is required.

CONCLUSIONS

A simulation model was developed of the full-scale municipal wastewater treatment plant at a steady state was developed. Processes included in this model are those of primary clarifier, aeration tank, final clarifier, thickener and pressure filter. The IAWPRC model was used for simulation of reactions in the aeration basin. This model was calibrated using the data of 5 and 10°C pilot plant experiments. The calibrated model was able to predict the results of other temperature experiments at the pilot plant. The full scale plant model yielded fairly accurate results on the sludge mass load of thickener, production rate and heating value of cake as well as on nitrification of the municipal wastewater plant in winter. Plant operation maps at 4 and 8°C were drawn. They showed that the enlargement of the sludge treatment system or the addition of nitrifying-bacteria-holding apparatus to aeration basins is required to maintain the present treatment level if the sewage system is used for melting and transporting snow.

ACKNOWLEDGMENT

The financial support provided by the research grant from the city of Sapporo is gratefully appreciated.

REFERENCES