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5 **1 Prenatal organochlorine pesticide exposure and the disruption of steroids and**
6 **2 reproductive hormones in cord blood: The Hokkaido Study**

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38 Conflicts of interest: none

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123 41 **Abstract**
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125 42 Certain organochlorine pesticides (OCPs) are designated as persistent organic pollutants
126
127 43 and are regulated in many countries. The effects of OCPs on pediatric endocrinology are
128
129 44 a concern; however, only limited data exist from human studies on maternal OCP
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131 45 exposure and its effects on infants' hormone levels. This study was conducted as part of
132
133 46 the Hokkaido Study Sapporo Cohort, a prospective birth cohort study in Japan.
134
135 47 Participants included 514 women who enrolled at 23–35 weeks of gestation between
136
137 48 2002 and 2005; maternal blood samples were collected in late pregnancy, and 29 OCPs
138
139 49 were measured. Reproductive and steroid hormone levels in cord blood were also
140
141 50 determined. Characteristics of mothers and their infants were obtained from self-
142
143 51 administered questionnaires and medical records. Ultimately, 232 samples with both
144
145 52 OCP and hormone data were analyzed. Fifteen of 29 investigated OCPs were detected
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147 53 in over 80% of the samples, with *p,p'*-dichlorodiphenyldichloroethylene showing the
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149 54 highest concentration (median value: 619 pg/g-wet). The association between OCPs and
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151 55 sex hormone levels varied by sex. Linear regression models after sex stratification
152
153 56 showed that chlordanes, cis-hexachlorobenzene, heptachlor epoxide, Mirex, and
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155 57 toxaphenes in maternal blood were inversely associated with testosterone, cortisol,
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157 58 cortisone, sex hormone-binding globin, prolactin, and androstenedione-
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159 59 dehydroepiandrosterone (DHEA) and testosterone-androstenediones ratios among boys.
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161 60 Furthermore, these OCPs were positively correlated with DHEA, follicle stimulating
162
163 61 hormone (FSH), and adrenal androgen-glucocorticoid and FSH-inhibin B ratios among
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165 62 boys. In categorical quartile models, testosterone and DHEA were inversely and
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167 63 positively associated with OCPs, respectively. Estradiol-testosterone and adrenal
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169 64 androgen-glucocorticoid ratios tended to increase with increasing OCP concentrations
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182 65 in the higher quartile, while the testosterone-androstenedione ratio tended to decrease.
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184 66 Sex hormone-binding globulin and prolactin showed an inverse association with OCPs.
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186 67 Among girls, the linear regression model showed that only *p,p'*-
187
188 68 dichlorodiphenyltrichloroethane was inversely associated with the level of DHEA and
189
190 69 the adrenal androgen-glucocorticoid ratio, but was positively associated with cortisone
191
192 70 levels. However, no associations were observed using the quartile categorical model.
193
194 71 These results suggest that prenatal exposure to OCPs disrupt reproductive hormones of
195
196 72 fetuses in utero among boys, even at relatively low levels.
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201
202 74 **Key Words:** Organochlorine pesticides; reproductive hormones; steroidal hormones;
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204 75 prenatal exposure; cord blood; birth cohort
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207
208 77 **Abbreviations:**

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210 78 CI, confidence interval
211
212 79 CYP11A1, cytochrome P450 family 11 subfamily A member 1
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214 80 CYP17A1, cytochrome P450 family 17 subfamily A member 1
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216 81 CYP19A1, cytochrome P450 family 19 subfamily A member 1
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218 82 DDD, dichlorodiphenyldichloroethane
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220 83 DDE, dichlorodiphenyldichloroethylene
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222 84 DDT, dichlorodiphenyltrichloroethane
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224 85 HCB, hexachlorobenzene
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226 86 HCE, heptachlor epoxide
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228 87 HCH, hexachlorocyclohexane
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230 88 HSD17B1, hydroxysteroid 17-beta dehydrogenase 1
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- 89 HSD3B1, hydroxy-delta-5-steroid dehydrogenase, 3 beta- and steroid delta-isomerase 1
- 90 IRMA, immunoradiometric assay
- 91 DHEA, dehydroepiandrosterone
- 92 EIA, enzyme immunoassay
- 93 ELISA, enzyme-linked immunosorbent assay
- 94 FSH, follicle stimulation hormone
- 95 INSL3, insulin-like factor 3
- 96 LC-MSMS, liquid chromatography-tandem mass spectrometry
- 97 LH, luteinizing hormone
- 98 LSM, least square mean
- 99 OCP, organochlorine pesticides
- 100 SHBG, sex hormone-binding globulin
- 101 StAR, steroidogenic acute regulatory protein

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103 **1. Introduction**

104 Organochlorine pesticides (OCPs) are chlorinated hydrocarbons used extensively in the
105 1940s for agriculture and pesticide control, and are now designated as persistent organic
106 pollutants by the Stockholm Convention (<http://chm.pops.int>). Although the Stockholm
107 Convention has issued an exemption for the production and public health use of
108 dichlorodiphenyltrichloroethane (DDT) to control vector-borne diseases, most OCPs
109 were banned in the United States, Europe, and many other countries in the early 1970s
110 (WHO, 2012). The use of OCPs has been eliminated or restricted in Japan since the
111 1970's (Kanazawa et al. 2012). Although most OCPs have been prohibited for over 30
112 years, they are still detected in the environment and in human populations. According to
113 Japanese monitoring data, the levels of DDT and its metabolites in water and sediment
114 have decreased since 1990 and have consistently remained low since 2000; however,
115 they are still detectable (Ministry of Environment, Japan 2006). Heptachlor epoxide
116 (HCE), hexachlorocyclohexane (HCH), Mirex, Parlar-26, and Parlar-50 are also above
117 detectable levels in water and sediments, even though the latter three have never been
118 used in Japan (Ministry of Environment, Japan 2006).

119 The endocrine disrupting properties of OCPs are considered a health concern. In
120 previous cross-sectional studies among adults, heptachlor and *o,p'*-DDT concentrations
121 were associated with lower testosterone levels in men (Freire et al. 2014). In women,
122 hexachlorobenzene (HCB), *p,p'*-DDT, *p,p'*-dichlorodiphenyldichloroethane (DDD),
123 endosulfan, aldrin, and Mirex showed inverse associations with luteinizing hormone
124 (LH) and follicle stimulation hormone (FSH) while showing positive associations with
125 prolactin (Freire et al. 2014).

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126 Maternal exposure to OCPs may affect fetal hormone levels. Sex steroid
127 hormones including testosterone, progesterone, and estradiol exert their functions
128 predominantly in the gonads, and dehydroepiandrosterone (DHEA) and
129 androstenedione are activated to form androgens and estrogens that have important roles
130 in sex differentiation and maturation (Labrie et al. 2001). Cortisol and cortisone are
131 synthesized within the adrenal cortex, are involved in a wide range of physiological
132 processes, and are essential for regulating and/or modulating homeostasis in
133 metabolism, growth, neurodevelopment, and the immune system (Braun et al. 2013;
134 Reynolds 2010). LH and FSH play critical roles in the development and regulation of
135 numerous body functions via the hypothalamic-pituitary-gonadal (HPG) axis (Kuiri-
136 Hänninen et al. 2014). Inhibin B and insulin-like factor-3 (INSL3) are major products
137 secreted by the Leydig and Sertoli cells, respectively, and the establishment of sufficient
138 numbers of these cells is critical for the production of sperms in adulthood (Ivell et al.
139 2013; Orth and Boehm 1990). In response to gonadotropins, testosterone (via LH
140 signaling) and inhibin B together act to regulate the secretion of FSH; these constitute
141 the major negative feedback signals that maintain the physiological function of the HPG
142 axis (Carlson, 2009). However, only limited data exist regarding human studies on
143 prenatal exposure to OCPs and their effects on steroids and reproductive hormone levels
144 in offspring. There is only one study in France that found that prenatal α -endosulfan and
145 HCE increase estradiol and sex hormone-binding globulin (SHBG), whereas these same
146 agents reduce testosterone levels at birth (Warembourg et al. 2016).

147 We have previously reported that 21 of 29 tested OCPs were detected in
148 maternal blood acquired between 2002 and 2005 in Japan (Kanazawa et al. 2012). The
149 impact of relatively low levels of OCP exposure on hormones at birth has still not been

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418 150 well-investigated in epidemiological studies. In particular, the effects of OCPs other
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420 151 than DDTs are rarely investigated. Thus, we hypothesized that prenatal exposure to
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423 152 even relatively low levels of these agents may alter hormone levels in infants. To that
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425 153 end, the aim of this study was to examine the associations between prenatal OCP
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427 154 exposure and cord blood steroid and reproductive hormone levels.
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429 155

431 156 **2. Methods**

433 157 *2.1 Participants*

435 158 This investigation was based on the Sapporo Cohort of the Hokkaido Study on
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437 159 Environment and Children's Health. Details of this study, including the population, data
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439 160 collection, sampling of the biological specimens, and contents of the administered
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441 161 questionnaire, were described previously (Kishi et al. 2017; Kishi et al. 2013; Kishi et
442
443 162 al. 2011). Briefly, Japanese pregnant women who lived in Sapporo City or surrounding
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445 163 areas were recruited at 23–35 weeks of gestation between July 2002 and October 2005
446
447 164 at an obstetrics and gynecology hospital in Sapporo, Hokkaido, Japan. Among the 1796
448
449 165 eligible women approached, 25% were excluded because they were enrolled in the
450
451 166 Japanese Cord Blood Bank or planned to deliver at another hospital. Ultimately, 514
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453 167 pregnant women (28.6% of those approached) were enrolled in this study.
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459 169 *2.2 OCP measurement*

460 170 Maternal blood samples were obtained at the time of patients' hospital examinations
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462 171 following recruitment (n=296). If a blood sample could not be obtained during
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464 172 pregnancy because of maternal anemia, a sample was collected during post-partum
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466 173 hospitalization within a week after delivery (n=130). All samples were stored at -80°C
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477 174 until analysis. OCPs in whole blood were measured by gas chromatography/high-
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479 175 resolution mass spectrometry and gas chromatography/negative-ion chemical-ionization
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481 176 mass spectrometry at IDEA Consultants, Inc. (Shizuoka, Japan) The 29 OCPs evaluated
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483 177 in this study were 5 chlordanes (*cis*-chlordane, *trans*-chlordane, *cis*-nonachlor, *trans*-
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485 178 nonachlor, and oxychlordane), 6 DDTs (*o,p'*-DDT, *p,p'*-DDT, *o,p'*-DDE, *p,p'*-DDE,
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487 179 *o,p'*-DDD, and *p,p'*-DDD), 3 'drins' (aldrin, dieldrin, and endrin), 3 heptachlors
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489 180 (heptachlors, *cis*-HCE, and *trans*-HCE), HCB, 4 HCH isomers (α -HCH, β -HCH, γ -
490
491 181 HCH, and δ -HCH), Mirex, and 6 toxaphenes (Parlar-26, Parlar-41, Parlar-40, Parlar-44,
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493 182 Parlar-50, and Parlar-62). Details of the measurement methods have been described
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495 183 previously (Kanazawa et al. 2012).
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500 185 *2.3 Measurement of steroids and reproductive hormones*

501
502 186 The methods used to measure steroids and reproductive hormones were described
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504 187 previously (Araki et al. 2017; Araki et al. 2014; Goudarzi et al. 2016). Briefly, the
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506 188 concentrations of 7 steroid hormones including progesterone, estradiol, testosterone,
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508 189 DHEA, androstenedione, cortisol, and cortisone in cord blood were measured using
509
510 190 liquid chromatography-tandem mass spectrometry (LC-MSMS) (Yamashita et al.
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512 191 2007a; Yamashita et al. 2007b). An immunoradiometric assay (IRMA) was used to
513
514 192 measure the concentrations of LH, FSH, and prolactin (Spac-S LH Kit, Spac-S FSH Kit,
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516 193 and Spac-Prolactin Kit, respectively, TFB, Inc., Tokyo Japan). SHBG was also
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518 194 measured using IRMA-Count SHBG (Siemens, Berlin, Germany). Concentrations of
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520 195 inhibin B were measured by using an enzyme-linked immunosorbent assay (ELISA)
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522 196 (Inhibin B Gen ELISA, Beckman Coulter, Inc., CA, USA), while INSL3 was measured
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524 197 by using an enzyme immunoassay (EIA) (INSL3/RLF [human] EIA kit, Phoenix
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198 Pharmaceutical., Inc., CA, USA). All hormone measurements were conducted at Aska
199 Pharma Medical Co., Ltd (Kanagawa, Japan).

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201 *2.4 Questionnaires and medical records*

202 The participants completed a self-administered questionnaire that extracted information
203 on maternal age, education level, household income, maternal smoking and alcohol
204 consumption during the first trimester, and medical history. Information at the time of
205 delivery, including pre-pregnancy body mass index, pregnancy complications,
206 gestational age, infant sex, parity, congenital anomalies such as hypospadias and
207 cryptorchidism, and infant size was obtained from medical records (Kishi et al. 2013;
208 Kishi et al. 2011).

209

210 *2.5 Statistical analyses*

211 Of the 514 participants, 10 were excluded from the study owing to miscarriage,
212 stillbirth, relocation, or voluntary withdrawal prior to delivery. Among 426 maternal
213 blood samples, 379 were of sufficient quantity for OCP analysis, while hormone
214 measurements were obtained from 295 infant cord blood samples. Ultimately, 232
215 matched maternal serum and cord blood samples (for OCP and hormone levels
216 measurements, respectively) were included in the statistical analysis.

217 Associations between maternal OCP concentrations and infant steroid hormone
218 levels were examined for each OCP separately via linear regression analysis. In each
219 model, the OCP was the independent variable while the hormone was the dependent
220 variable. Initially, linear regression models for both sexes combined were constructed
221 with the interaction terms of sex \times OCP levels added in each model; this revealed

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595 222 significant differences between the sexes. Next, the models were applied following
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597 223 stratification by sex. OCP levels and the concentrations of steroid hormones were
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599 224 converted to a \log_{10} scale because they were not normally distributed. Two-sided *P*-
600
601 225 values <0.05 were considered statistically significant. Selected OCPs with *P*-values
602
603 226 <0.1 in linear regression models were then categorized into concentration quartiles to
604
605 227 examine dose-response relationships. The interquartile range for each OCP
606
607 228 concentration and the least squares means (LSM) of log-transformed hormone levels
608
609 229 were calculated and back-transformed. To calculate a *P*-value for the trend, linear
610
611 230 contrast coefficients of -3 , -1 , $+1$, and $+3$ were assigned to the first, second, third, and
612
613 231 fourth quartiles, respectively (Goudarzi et al., 2016; Itoh et al., 2016). *P*-values for trend
614
615 232 <0.05 were considered statistically significant. The OCP levels in the first quartile were
616
617 233 also compared to those in the second, third, and fourth quartiles using the Dunnett-Hsu
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619 234 method; *P*-values were adjusted using Bonferroni's correction ($P<0.0167$). When below
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621 235 their detection limits, the half values of these detection limits were used for both OCPs
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623 236 and hormones.

627 237 In addition to single hormone levels, we also examined the product-to-substrate
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629 238 ratios of hormones that are adjacent in the metabolic pathway to determine their enzyme
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631 239 activity indices (Hicks et al. 2014). For example, the estradiol-testosterone ratio
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633 240 represents the index of the cytochrome P450 family 19 subfamily A member 1
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635 241 (CYP19A1), better known as aromatase. The androstenedione-DHEA ratio represents
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637 242 the hydroxy-delta-5-steroid dehydrogenase, 3 beta- and steroid delta-isomerase 1
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639 243 (HSD3B1) index; the testosterone-androstenedione ratio represents the hydroxysteroid
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641 244 17-beta dehydrogenase 1 (HSD17B1) index; and the cortisone-cortisol ratio represents
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643 245 the HSD3B1 index. Increasing and decreasing ratios suggest the up- and
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654 246 downregulation of enzyme activity, respectively. Additionally, the adrenal androgen
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656 247 (the sum of DHEA and androstenedione)-glucocorticoid (sum of cortisol and cortisone)
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658 248 ratio was examined to determine the balance shift of adrenal androgen (C19-steroids)
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660 249 and glucocorticoid (C21-steroids) (Goudarzi et al. 2016). Similarly, testosterone-LH
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662 250 and FSH-inhibin B ratios were examined as indices of gonadal function. The inclusion
663
664 251 of covariates was examined based on biological considerations, and included maternal
665
666 252 age (continuous), parity (primipara or multipara), and gestational age (continuous). All
667
668 253 statistical analyses were performed using the Japanese version of IBM SPSS Statistics
669
670 254 19 (IBM Analytics, NC, USA) and the Japanese version of JMP Pro 12 (SAS Institute
671
672 255 Inc., NC, USA).

673 674 675 676 677 257 *2.6 Ethical approval*

678
679 258 This study was approved by the Institutional Ethical Board for Epidemiological Studies
680
681 259 at Hokkaido University Graduate School of Medicine and Hokkaido University Center
682
683 260 for Environmental and Health Sciences, in accordance with the principles of the
684
685 261 Declaration of Helsinki. All participants provided written informed consent.

686 687 688 689 690 263 **3. Results**

691
692 264 Table 1 shows the characteristics of the participants included in this study as well as
693
694 265 those of the original cohort. Compared to the original cohort, the mean birth weight and
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696 266 gestational age were slightly larger among the participants; the vaginal delivery rate
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698 267 among participants was 99.1%. One infant with cryptorchidism was included in the
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701 268 study.

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713 269 The concentrations of OCP in maternal blood samples are shown in Table 2.
714
715 270 There were 15 OCPs (3 chlordanes [*cis*-nonachlor, *trans*-nonachlor, and oxychlordane],
716 271 5 DDTs [*o,p'*-DDT, *p,p'*-DDT, *o,p'*-DDE, *p,p'*-DDE, *p,p'*-DDD], dieldrin, *cis*-HCE,
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718 272 HCB, β -HCH, Mirex, and 2 toxaphenes [Parlar-26 and Parlar-50]) that were above 80%
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720 273 of their detection limits; these OCPs were subjected to further analysis. The median
721
722 274 concentration of *p,p'*-DDE was the highest at 619.26 pg/g-wet, followed by β -HCH at
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724 275 154.31 pg/g-wet, and HCH at 103.99 pg/g-wet. The distributions of OCPs in the
725
726 276 original cohort are shown in Supplemental Table S1; *trans*-chlordane, *cis*-nonachlor,
727
728 277 and *trans*-nonachlor levels of participants in this study were slightly higher compared
729
730 278 with the levels in those who were excluded (Mann-Whitney U test, $P=0.06$, 0.039, and
731
732 279 0.040, respectively; data not shown). Associations between OCP levels and maternal
733
734 280 and child characteristics are shown in Supplemental Table S1. The levels of most OCPs
735
736 281 increased with the ages of the mothers. OCP concentrations were not significantly
737
738 282 associated with infant characteristics such as sex, birth weight, or gestational age.
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743 283 The steroid and reproductive hormone levels in infants are shown in Table 3.
744
745 284 Testosterone and DHEA levels were significantly higher and lower in boys than in girls,
746
747 285 respectively. The detection rate of LH, FSH, and inhibin B was below 30% in girls,
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749 286 among whom only 16 samples were analyzed for INSL3. Therefore, no further tests of
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751 287 LH, FSH, inhibin B, and INSL3 were conducted for girls. The distribution of all
752
753 288 measured hormone levels are shown in Supplemental Table S3; there were no
754
755 289 differences between subjects included and excluded from the study.
756
757

758 290 The associations between OCPs and steroid and reproductive hormone levels for
759
760 291 both sexes combined were examined (Supplemental Table S4). Sex-specific OCP
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762 292 interactions ($P<0.05$) between one or more hormones and *cis*-nonachlor, *trans*-
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772 293 nonachlor, *p,p'*-DDE, *o,p'*-DDT, dieldrin, and Mirex were observed, suggesting that the
773
774 294 effects of OCPs differ according to sex.

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777 295 Adjusted regression coefficients (β) and 95% confidence intervals (CIs) for the
778
779 296 association between a 10-fold increase of OCP and \log_{10} -transformed hormone levels as
780
781 297 determined by a linear regression model in boys and girls are shown in Tables 4 and 5,
782
783 298 respectively. Among infant boys, Mirex was inversely associated with testosterone, and
784
785 299 *p,p'*-DDE showed a positive association with the estradiol-testosterone ratio.

786
787 300 Oxychlorthane, *cis*-nonachlor, *trans*-nonachlor, dieldrin, *cis*-HCE, HCB, Mirex, Parlar-
788
789 301 26, and Parlar-50 were positively associated with DHEA. Oxychlorthane, *trans*-
790
791 302 nonachlor, *cis*-HCE, and Mirex were inversely associated with the androstenedione-
792
793 303 DHEA ratio. Oxychlorthane, *cis*-nonachlor, *trans*-nonachlor, dieldrin, *cis*-HCE, HCB,
794
795 304 Mirex, and Parlar-50 showed inverse associations with the testosterone-androstenedione
796
797 305 ratio. Both *trans*-nonachlor and Mirex were inversely associated with cortisol and
798
799 306 cortisone. Oxychlorthane, *cis*-nonachlor, *trans*-nonachlor, *cis*-HCE, Mirex, and Parlar-
800
801 307 50 were positively associated with the adrenal androgen-glucocorticoid ratio. β -HCH
802
803 308 was inversely associated with SHBG, and β -HCH, Mirex, Parlar-26, and Parlar-50 were
804
805 309 positively associated with FSH. *o,p'*-DDE, *p,p'*-DDE, *o,p'*-DDT, *p,p'*-DDT, dieldrin, β -
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807 310 HCH, Mirex, and Parlar-50 were all inversely associated with prolactin. Finally, *cis*-
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809 311 nonachlor, *p,p'*-DDE, *p,p'*-DDT, *cis*-HCE, HCB, β -HCH, Mirex, Parlar-26, and Parlar-
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811 312 50 were inversely associated with the FSH-inhibin B ratio. There was no statistically
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813 313 significant association between OCPs and progesterone, estradiol, androstenedione,
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815 314 FSH, inhibin B, or INSL3. Among girls, *p,p'*-DDD was inversely associated with
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817 315 DHEA and the adrenal androgen-glucocorticoid ratio, and was positively associated
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831 316 with cortisone. There was no association between OCPs and progesterone, estradiol,
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833 317 testosterone, androstenedione, cortisol, SHBG, or prolactin among girls.

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836 318 Figure 1 shows the relationships between hormones and OCPs in quartile models
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838 319 for boys (only in cases where $P < 0.05$ was observed). Testosterone showed a decreasing
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840 320 trend in relation to quartiles of Mirex, while LSM analysis of the estradiol-testosterone
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842 321 ratio showed an increasing trend of p,p' -DDE in relation to quartiles. DHEA showed
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844 322 increasing trends in relation to the quartiles of *cis*-Nonachlor, Dieldrin, Parlar-26, and
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846 323 Parlar-50. Moreover, the LSM method showed that the 4th quartile of *cis*-nonachlor was
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848 324 significantly increased compared to the 1st quartile *cis*-nonachlor. LSM analysis of the
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850 325 testosterone-androstenedione ratio showed decreasing trends in relation to *cis*-
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852 326 nonachlor, *trans*-nonachlor, Dieldrin, Mirex, and Parlar-50. Moreover, LSM analysis of
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854 327 the adrenal androgen-corticoid ratio showed an increasing trend of *cis*-nonachlor,
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856 328 Parlar-26, and Parlar-50, while SHBG showed a decreasing trend of β -HCH. Prolactin
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858 329 showed decreased trends of p,p' -DDE and o,p' -DDT. Statistically significant
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860 330 relationships between hormones and OCPs in the quartile models were not found among
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862 331 girls.

863 864 865 332 866 867 333 **4. Discussion**

868
869 334 In linear models, we found that relatively low levels of OCPs were inversely associated
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871 335 with testosterone, cortisol, cortisone, SHBG, and prolactin, but positively associated
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873 336 with DHEA in newborn boys after stratification by sex. Positive associations between
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875 337 OCPs and each of estradiol-testosterone and adrenal androgen-glucocorticoid, as well as
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877 338 an inverse association between OCPs and each of androstenedione-DHEA, testosterone-
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879 339 androstenedione, corticoid-cortisone, and the FSH-inhibin B ratio, were also observed
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340 among boys. In quartile models, testosterone was inversely associated with OCPs,
341 whereas DHEA was positively associated with them. The estradiol-testosterone and
342 adrenal androgen-glucocorticoid ratios tended to increase when OCP concentrations
343 were the higher quartile, while the testosterone-androstenedione ratio tended to
344 decrease. SHBG and prolactin showed inverse associations with OCPs. Among girls,
345 *p,p'*-DDD was inversely associated with DHEA levels and positively associated with
346 levels of cortisol; it was also inversely associated with the glucocorticoid-adrenal
347 androgen ratio. However, these associations were not observed in the quartile models,
348 which would have provided more credence to the findings of the linear models. Overall,
349 our data suggested that the natures of the associations between OCP and hormones
350 differ according to sex, and that clear associations were observed only among infant
351 boys.

352 The levels of OCPs in maternal blood have been measured in several studies
353 performed in various countries. In this study, the median *p,p'*-DDE value in maternal
354 whole blood was 619.26 pg/g-wet. In a study in Chiapas, Mexico, the median values of
355 DDT and DDE in maternal serum collected in 2002–2003 were 1.9 and 19.5 µg/L,
356 respectively (Longnecker et al. 2007). A study of Mexican-Americans in the US state of
357 California found that the maternal geometric mean values of *p,p'*-DDT, *o,p'*-DDT, and
358 *p,p'*-DDE were 22.0, 1.8, and 1436.9 ng/g-lipid, respectively (Eskenazi et al. 2006). A
359 study in China showed geometric means of *p,p'*-DDE, HCB, and β-HCH in maternal
360 serum, at 203.54 ng/g, 70.62 ng/g, and 67.67 ng/g, respectively (Guo et al. 2014). In
361 another recent study in South Africa, the median maternal *p,p'*-DDE level was 241.3
362 ng/g lipids (Bornman et al. 2016). A meta-analysis of 12 European cohorts found that
363 the median cord serum *p,p'*-DDE concentration was 527.9 ng/L (ranging from 49.8

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949 364 ng/L to 1208 ng/L in various studies) (Govarts et al. 2012). By using Govarts et al.'s
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951 365 conversion factor to calculate cord serum levels from maternal whole blood levels in
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953 366 our study (cord serum level = $0.36 \times$ maternal whole blood level), the estimated
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955 367 median level of *p,p'*-DDE in the cord serum of our study was 229.1 ng/L, which is
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957 368 approximately half the median value of European cohorts. Taken together, the OCP
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959 369 exposure levels in our cohort were relatively low in comparison.

962 370 Only 1 study, performed in France, examined prenatal OCP exposure and
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964 371 hormone levels at birth (Warembourg et al. 2016). The investigators found that higher
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966 372 levels of HCB and HCE were associated with reduced levels of testosterone and
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968 373 elevated levels of SHBG among boys. HCE was also associated with higher levels of
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970 374 estradiol and a lower testosterone-estradiol ratio among girls. There was no association
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972 375 between *p,p'*-DDE and any hormone levels. Their results were partly consistent with
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974 376 ours. In our study, OCPs including HCB and HCE showed inverse associations with
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976 377 testosterone; these were not significant except for Mirex among boys. Although we did
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978 378 not find any association between estradiol and OCPs, their levels tracked together in a
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980 379 manner also observed by Warembourg et al. (2016). Moreover, the positive association
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982 380 between the estradiol-testosterone ratio and OCPs, which was statistically significant for
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984 381 *p,p'*-DDE in this study, are consistent with the findings of Warembourg et al. (2016), as
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986 382 is the significant decrease in the aromatase index (testosterone/estradiol) with increasing
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988 383 HCE. On the other hand, SHBG had a significant inverse correlation with Mirex in our
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990 384 study; which was in direct contrast to Warembourg et al.'s findings; we are unable to
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992 385 explain this discrepancy, as the sampling period (2002–2006) and sample sizes (n=282
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994 386 for their study and n=232 for ours) were comparable. Additionally, the exposure levels
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996 387 of DDE, HCE, and HCB in our study and in that of Warembourg et al. were
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1008 388 comparable. However, the detection rates of HCB and HCE were higher in our study,
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1010 389 and Warembourg et al. did not measure chlordanes, Mirex, or toxaphenes. Because of
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1012 390 the low detection percentage, Warembourg et al. divided OCP levels into two or three
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1014 391 categories to examine the association of each with hormones; this could explain why
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1016 392 their results do not match ours more closely. We found no other studies that investigated
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1018 393 associations between OCPs and DHEA, androstenedione, cortisol, or cortisone. More
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1020 394 studies are therefore warranted to ascertain the association between OCPs and
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1022 395 steroidogenesis at birth.

1025 396 In this study, DDT, DDE, and DDD were not associated with steroid hormones
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1027 397 except for *p,p'*-DDE, which showed a positive association with the estradiol-
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1029 398 testosterone ratio in boys. The increased estradiol-testosterone ratio suggested increased
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1031 399 CYP19A1 enzyme activity. A previous animal study showed that *p,p'*-DDE induces
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1033 400 CYP19A1 in hepatic microsomal samples of adult male rats (You et al. 2001), which is
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1035 401 consistent with our results. In girls, *p,p'*-DDD was inversely associated with DHEA and
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1037 402 the adrenal androgen-glucocorticoid ratio, but was positively associated with cortisone
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1039 403 levels. However, these associations were not observed in the quartile models; therefore,
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1041 404 false positive associations are likely. Longnecker et al. (2007) conducted a study in
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1043 405 Mexico where maternal DDT levels are relatively high, and found no evidence that in
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1045 406 utero exposure to DDE was related to anogenital distance or penile dimensions at birth
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1047 407 among male infants. Additionally, Bornman et al. (2016) conducted a similar study in
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1049 408 South Africa and reported no associations between *p,p'*-DDT/-DDE or *o,p'*-DDT and
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1051 409 anogenital distance measurements at birth in either boys or girls. Yet another study in
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1053 410 Denmark found no association between prenatal *p,p'*-DDE levels and reproductive
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1055 411 hormones at approximately 20 years of age (Vested et al. 2014). Taken together, these
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1067 412 data indicate that DDTs may not alter infant steroid hormones at the levels detected in
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1069 413 these studies.

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1072 414 Among boys, *p,p'*-DDE and *o,p'*-DDT were inversely associated with prolactin
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1074 415 in quartile models. Although there have been no specific studies of the mechanisms of
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1076 416 OCP influence on prolactin, other investigations suggest that newborns with lower
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1078 417 levels of prolactin in their cord blood are more likely to have an increased risk of
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1080 418 respiratory distress syndrome than those with higher levels of prolactin (Parker et al.
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1082 419 1989; Padvi et al. 2017). Combined with data from the Canadian Health Measures
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1084 420 Survey that DDT exposure is inversely associated with lung function parameters in
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1086 421 adults (Ye et al. 2015), long term follow-up of children who were exposed to these
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1088 422 chemicals is needed to clarify the health-related consequences of such exposure.

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1090 423 Studies of OCPs other than DDTs are scarce. We found that *cis/trans*-
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1092 424 nonachlor, Dieldrin, Mirex, and toxaphenes (Parlar-26 and 50) were positively
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1094 425 associated with DHEA as well as with the adrenal androgen-glucocorticoid ratio, but
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1096 426 were inversely associated with testosterone, the testosterone-androstenedione ratio, and
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1098 427 prolactin in boys. Decreasing testosterone-androstenedione ratio trends suggests the
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1100 428 downregulation of HSB17B1; previous animal studies have shown that aldrin inhibits
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1102 429 HSB17B1 (Chatterjee et al. 1988), which is consistent with our findings. DHEA is the
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1104 430 main precursor of sex hormones and cortisol antagonists (Mastorakos and Ilias 2003).
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1106 431 The increasing adrenal androgen-glucocorticoid ratio suggests that chlordanes, Dieldrin,
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1108 432 Mirex, and toxaphenes may shift steroidogenesis towards androgenic hormones (C19-
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1110 433 steroids), although we observed no evidence of this. However, circulating hormones are
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1112 434 regulated by several cascade reactions. Translocation of cholesterol from the outer
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1114 435 mitochondrial membrane to the inner membranes is a critical step in steroidogenesis;
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1126 436 this process is enhanced by steroidogenic acute regulatory protein (StAR), which is
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1128 437 encoded by *STARD1*. Other key human steroidogenic genes are *CYP11A1*, *CYP17A1*,
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1130 438 *CYP11B1/CYP11B2*, *CYP21A2*, *SRD5A1*, and *SRD5A2*. Moreover, *SULT2A1* and
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1133 439 *SULT2B1* encode enzymes that convert DHEA to DHEA-sulfate and vice versa.
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1135 440 Although we have not measured methoxychlor, another synthetic organochlorine
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1137 441 insecticide and a derivative of DDT showed decreased expression of CYP19A1,
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1139 442 HSD17B1, *CYP17A1*-encoded cytochrome P450 family 17 subfamily A member 1,
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1141 443 HSD3B1, *CYP11A1*-encoded cytochrome P450 family 11 subfamily A member 1, and
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1143 444 StAR, but increased expression of *CYP11B1*-encoded cytochrome P450 family 1
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1145 445 subfamily B member 1 enzyme levels, *in vitro* (Basavarajappa et al. 2011; Vaithinathan
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1147 446 et al. 2008). Therefore, our overall findings do not rule out that many of the OCPs
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1149 447 measured in this study may alter steroid hormones by inhibiting such steroidogenesis
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1151 448 enzymes. Moreover, a previous study notably examined 15 organochlorines (OCs)
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1153 449 mixture similar to those found in the bladders of Arctic ringed seals. Exposing male rats
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1155 450 *in utero* to this OC mixture caused disruptions in the development of androgen-
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1157 451 dependent organs such as the testis, epididymis, seminal vesicle, and prostate (Anas et
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1159 452 al. 2005). They also investigated that OC mixture's direct inhibition of Leydig cell
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1161 453 steroidogenesis by disrupting cholesterol transport into the mitochondria via decreasing
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1163 454 StAR protein levels, and by converting cholesterol into pregnenolone by modulating
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1165 455 adrenodoxin reductase and CYP11A1 proteins (Enangue Njembele et al. 2014). We
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1167 456 speculate that different OCPs influence steroidogenesis similarly. Additional studies on
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1169 457 altered steroidogenic and metabolic enzymes with different OCPs are required to test
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1171 458 these hypotheses.
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1185 459 In our linear regression model, HCH, Mirex, and toxaphenes were positively
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1187 460 associated with FSH, as was the FSH-inhibin B ratio. Sertoli cells secrete inhibin B,
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1189 461 which was first identified by its ability to negatively regulate FSH (Carlson et al. 2009).
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1191 462 Increasing levels of FSH together with the FSH-inhibin B ratio point to a negative
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1193 463 feedback mechanism as well as the influence of OCPs on Sertoli cells. However, these
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1195 464 associations were not statistically significant in our quartile model; hence, the effects of
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1197 465 OCPs on Sertoli cells remain unclear.
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1200 466 Our study found a significant association between OCPs and hormones only
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1202 467 among boys. Although the reason for this remains unclear, one possible explanation for
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1204 468 this sex preference is that most OCPs have an affinity for hormone receptors including
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1206 469 androgen receptor, estrogen receptor, and/or aryl hydrocarbon receptor (Mnif et al.
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1208 470 2011; Kojima et al., 2004). Thus, OCPs may mimic or block the natural hormone's
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1210 471 action (as an agonist or antagonist, respectively), and can interfere with the synthesis,
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1212 472 transport, metabolism, and elimination of hormones (Mnif et al. 2011). Estrogen and
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1214 473 androgen receptors are primarily involved in sexual differentiation and reproduction
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1216 474 (Busillo et al. 2009). While estrogens and progestins are essential for normal female
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1218 475 development, androgens are involved in various aspects of male reproductive
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1220 476 physiology. One animal study showed that *p,p'*-DDE inhibits androgen binding to the
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1222 477 AR, androgen-induced transcriptional activity, and androgen action in developing
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1224 478 (Kelce W.R., et al., 1995). Androgen receptors are expressed in variety of tissues, but
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1226 479 their levels differed in human fetal tissues (Wilson and McPhaul, 1996). During fetal
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1228 480 genital development, the female phenotype is considered to be the baseline (or default)
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1230 481 condition, and the development of maleness requires additional secretions produced by
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1232 482 the testis (Carlson 2009). Thus, we speculate that the effect of anti-androgenic activities
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1245 483 of OCPs on males is more severe than the effect of their estrogenic activities on
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1247 484 females. The physiological differences in the roles of these hormones may explain the
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1249 485 observation of sex differences in OCP-hormone associations. One epidemiological
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1251 486 study in Mexico found that higher exposure to *p,p'*-DDE *in utero* shortened the anal
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1253 487 position index among boys but not among girls (Torres-Sanchez et al., 2008). The
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1255 488 authors explained that this was due to the putative androgen deficiency occurring due to
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1257 489 the reduction of transcriptional activity that occurs when AR is blocked by *p,p'*-DDE *in*
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1259 490 *utero*. Although the exposure levels in Torres-Sanchez et al.'s study were 10 times
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1261 491 higher than that in ours, our results were in line with those of their study. *In vitro*
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1263 492 reporter gene assays showed that not only *p,p'*-DDE, but also other DDTs, dieldrin, and
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1265 493 heptachlors showed AR antagonist effects (Kojima et al., 2004). Hence, we assume that
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1267 494 those OCPs studied herein with AR antagonist properties share similar modes of action
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1269 495 in terms of male-specific responses.

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1272 496 The strength of this study is that we examined a wide range of OCPs.
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1274 497 Investigating seven steroids enabled us to construct a clearer picture of the association
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1276 498 between steroidogenesis and OCPs levels. Measuring steroid hormones by LC-MSMS
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1278 499 is considered more accurate than other methods such as radioimmunoassays, which is
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1280 500 another strength of this study. However, there are several limitations as well. First, the
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1282 501 15 OCPs are moderately-to-highly correlated with each other (Spearman's rho: 0.230–
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1284 502 0.918, $P < 0.001$). Thus, we were unable to clarify the effect of individual OCPs on
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1286 503 hormones. Moreover, there may be other residual confounding factors. Previous studies
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1288 504 suggested that reproductive hormones in cord blood may be affected by factors such as
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1290 505 diurnal variation, duration of labor, placental weight, and the presence of pre-eclampsia
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1292 506 (Hollier et al. 2014; Keelan et al. 2012). A recent cross-sectional study among adults
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1303 507 suggested an association between OCP concentrations and total cholesterol levels
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1305 508 (Arrebola et al. 2014), which may also modify steroid hormones levels. Second,
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1308 509 multiple tests of 15 OCPs and 13 hormones may have found statistically significant
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1310 510 associations by chance. However, the associations between OCPs and hormones
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1312 511 exhibited consistent trends, which suggest that the results are robust and that OCP
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1314 512 exposure is likely to alter infants' hormones in utero. Finally, there is a possibility of
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1316 513 selection bias in this study, as only participants with available cord blood samples were
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1318 514 included in the analysis. Because the life and the safety of the mother and child were of
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1320 515 utmost priority, cord blood samples were seldom acquired during Caesarian sections.
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1322 516 Thus, two infants were delivered Caesarian sections and others included in this study
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1324 517 were delivered vaginally; they had longer gestational ages and heavier birth weights
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1326 518 than the infants who were excluded, which indicated that the analysis was biased
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1328 519 towards healthier infants. Therefore, the effects of OCPs may have been underestimated
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1331 520 in this study.

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1334 1335 522 **5. Conclusion**

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1337 523 We found that exposure to relatively low levels of OCPs such as *cis/trans*-nonachlor,
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1339 524 *p,p'*-DDE, *o,p'*-DDT, Dieldrin, β -HCH, Mirex, and toxaphenes *in utero* was
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1341 525 significantly associated with levels of hormones and their ratios in male fetuses. These
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1343 526 results suggest that OCPs that include but are not limited to DDTs ought to be
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1345 527 examined, as the existing data are limited. Disrupting the balance of steroid hormones
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1347 528 may cause adverse effects on reproductive growth, development, and other health
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1349 529 outcomes in later life. The clinical significance of these findings is unclear at present, as
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1351 530 it remains unknown whether these small hormonal alterations are of any future health
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531 consequences in these individuals. Therefore, further studies investigating the long-term
532 effects of OCP exposure are required.

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542 **References**

543 Araki, A., Mitsui, T., Miyashita, C., Nakajima, T., Naito, H., Ito, S., Sasaki, S., Cho, K.,
544 Ikeno, T., Nonomura, K., Kishi, R., 2014. Association between maternal exposure
545 to di(2-ethylhexyl) phthalate and reproductive hormone levels in fetal blood: the
546 hokkaido study on environment and children's health. PLoS One. 9, e109039.

547 Araki, A., Mitsui, T., Goudarzi, H., Nakajima, T., Miyashita, C. Itoh, S., Sasaki, S.,
548 Cho, K., Moriya, K., Shinohara, N., Nonomura, K., Kishi, R., 2017. Prenatal di(2-
549 ethylhexyl) phthalate exposure and disruption of adrenal androgens and
550 glucocorticoids levels in cord blood: The Hokkaido Study. Sci. Total Environ.
551 581–582, 297–304.

552 Arrebola, J.P., Ocana-Riola, R., Arrebola-Moreno, A.L., Fernandez-Rodríguez, M.,
553 Martin-Olmedo, P., Fernandez, M.F., Olea, N., 2014. Associations of accumulated
554 exposure to persistent organic pollutants with serum lipids and obesity in an adult
555 cohort from Southern Spain. Environ Pollut. 195, 9-15.

556 Anas, M.K.I., Guillemette, C., Ayotte, P., Pereg, D., Giguère, F., Bailey, J.L., 2005. In
557 Utero and Lactational Exposure to an Environmentally Relevant Organochlorine
558 Mixture Disrupts Reproductive Development and Function in Male Rats. Biol.
559 Reprod, 73, 14-426.

560 Basavarajappa, M.S., Craig, Z. R., Hernández-Ochoa, I., Paulose, T., Leslie, T.C.,
561 Flaus, J. A., 2011. Methoxychlor reduces estradiol levels by altering
562 steroidogenesis and metabolism in mouse antral follicles in vitro. Toxicol Appl
563 Pharmacol, 253, 161-169.

564 Bernard, L., Martinat, N., Lécureuil, C., Crépieux, P., Reiter, E., Tilloy-Ellul, A.,
565 Chevalier, S., Guillou, F., 2007. Dichlorodiphenyltrichloroethane impairs follicle-

1476
1477
1478
1479
1480 566 stimulating hormone receptor-mediated signaling in rat Sertoli cells. *Reprod.*
1481
1482
1483 567 *Toxicol.* 23, 158–164.
1484
1485 568 Bornman, M.S., Chevrier, J., Rauch, S., Crause, M., Obida, M., Sathyanarayana, S.,
1486
1487 569 Barr, D.B., Eskenazi, B., 2016. Dichlorodiphenyltrichloroethane exposure and
1488
1489 570 anogenital distance in the Venda Health Examination of Mothers, Babies and their
1490
1491 571 Environment (VHEMBE) birth cohort study, South Africa. *Andrology.* 4, 608–
1492
1493 572 615.
1494
1495 573 Braun, T., Challis, J.R., Newnham, J.P., Sloboda, D.M., 2013. Early-life glucocorticoid
1496
1497 574 exposure: the hypothalamic-pituitary-adrenal axis, placental function, and long-
1498
1499 575 term disease risk. *Endocr. Rev.* 34, 885–916.
1500
1501 576 Busillo, J. M., Rhen, T., Cidlowski, J.A., 2009. Chapter 5 Steroid Hormone Action, in:
1502
1503 577 Strauss III, J.F., Barbieri, R.L. (Eds.), *Yen and Jaffe’s Reproductive*
1504
1505 578 *Endocrinology, Physiology, Pathophysiology, and Clinical Management*, 6th
1506
1507 579 edition. Elsevier, Philadelphia, PA, USA, pp. 105-120.
1508
1509 580 Carlson, B.M., 2009. Fourth Edition, *Human Embrology and Developmental Biology*.
1510
1511 581 Elsevier, Philadelphia, PA, USA.
1512
1513 582 Chatterjee, S., Ray, A., Bagchi, P., Deb, C., 1988. Suppression of testicular
1514
1515 583 steroidogenesis in rats by the organochlorine insecticide Aldrin. *Environ. Pollut.*
1516
1517 584 51, 87–94.
1518
1519 585 Ministry of Environment, 2006. Available at:
1520
1521 586 https://www.env.go.jp/chemi/anzen/c_guidebook_h18/01.pdf [accessed 25 May
1522
1523 587 2017].
1524
1525
1526
1527
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1588
1589
1590
1591
1592
1593

588 Enangue Njembele, A.N., Bailey, J.L., Tremblay, J.J., 2014. In Vitro Exposure of
589 Leydig Cells to an Environmentally Relevant Mixture of Organochlorines
590 Represses Early Steps of Steroidogenesis¹. *Biol Reprod.*, 90, 1-10.

591 Eskenazi, B., Marks, A.R., Bradman, A., Fenster, L., Johnson, C., Barr, D.B., Jewell,
592 N.P., 2006. In utero exposure to dichlorodiphenyltrichloroethane (DDT) and
593 dichlorodiphenyldichloroethylene (DDE) and neurodevelopment among young
594 Mexican American children. *Pediatrics*. 118, 233–241.

595 Freire, C., Koifman, R.J., Sarcinelli, P.N., Rosa, A.C., Clapauch, R., Koifman, S., 2014.
596 Association between serum levels of organochlorine pesticides and sex hormones
597 in adults living in a heavily contaminated area in Brazil. *Int. J. Hyg. Environ.*
598 *Health*. 217, 370–378.

599 Goudarzi, H., Araki, A., Itoh, S., Sasaki, S., Miyashita, C., Mitsui, T., Nakazawa, H.,
600 Nonomura, K., Kishi, R., 2017. The association of prenatal exposure to
601 perfluorinated chemicals with glucocorticoid and androgenic hormones in cord
602 blood samples: The Hokkaido Study. *Environ. Health Perspect.* 125, 111–118.

603 Govarts, E., Nieuwenhuijsen, M., Schoeters, G., Ballester, F., Bloemen, K., de Boer,
604 M., Chevrier, C., Eggesbø, M., Guxens, M., Krämer, U., Legler, J., Martínez, D.,
605 Palkovicova, L., Patelarou, E., Ranft, U., Rautio, A., Petersen, M.S., Slama, R.,
606 Stigum, H., Toft, G., Trnovec, T., Vandentorren, S., Weihe, P., Kuperus, N.W.,
607 Wilhelm, M., Wittsiepe, J., Bonde, J.P.; Obelix., ENRIECO, 2012. Birth weight
608 and prenatal exposure to polychlorinated biphenyls (PCBs) and
609 dichlorodiphenyldichloroethylene (DDE): A meta-analysis within 12 European
610 birth cohorts. *Environ. Health Perspect.* 120, 162–170

1594
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1597
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1650
1651
1652

611 Guo, H., Jin, Y., Cheng, Y., Leaderer, B., Lin, S., Holford, T.R., Qiu, J., Zhang, Y., Shi,
612 K., Zhu, Y., Niu, J., Bassig, B.A., Xu, S., Zhang, B., Li, Y., Hu, X., Chen, Q.,
613 Zheng, T., 2014. Prenatal exposure to organochlorine pesticides and infant birth
614 weight in China. *Chemosphere*. 110, 1–7.

615 Haugen, T.B., Tefre, T., Malm, G., Jönsson, B.A., Rylander, L., Hagmar, L., Bjørsvik,
616 C., Henrichsen, T., Sæther, T., Figenschau, Y., Giwercman, A., 2011. Differences
617 in serum levels of CB-153 and p,p'-DDE, and reproductive parameters between
618 men living south and north in Norway. *Reprod. Toxicol.* 32, 261–267.

619 Hicks, R.A., Yee, J.K., Mao, C.S., Graham, S., Kharrazi, M., Lorey, F., Lee, W.P.,
620 2014. Precursor-to-product ratios reflect biochemical phenotype in congenital
621 adrenal hyperplasia. *Metabolomics*. 10, 123–131.

622 Hollier, L.P., Keelan, J.A., Hickey, M., Maybery, M.T., Whitehouse, A.J., 2014.
623 Measurement of androgen and estrogen concentrations in cord blood: accuracy,
624 biological interpretation, and applications to understanding human behavioral
625 development. *Front. Endocrinol.* 5, 64.

626 Ivell, R., Wade, J.D., Anand-Ivell, R., 2013. INSL3 as a biomarker of Leydig cell
627 functionality. *Biol. Reprod.* 88, 1–8.

628 Jusko, T.A., Klebanoff, M.A., Brock, J.W., Longnecker, M.P., 2012. In-utero exposure
629 to dichlorodiphenyltrichloroethane and cognitive development among infants and
630 school-aged children. *Epidemiology*. 23, 689–698.

631 Kanazawa, A., Miyasita, C., Okada, E., Kobayashi, S., Washino, N., Sasaki, S.,
632 Yoshioka, E., Mizutani, F., Chisaki, Y., Saijo, Y., Kishi, R., 2012. Blood
633 persistent organochlorine pesticides in pregnant women in relation to physical and

1653
1654
1655
1656
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1658
1659
1660
1661
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1710
1711

634 environmental variables in The Hokkaido Study on Environment and Children's
635 Health. *Sci. Total Environ.* 426, 73–82.

636 Keelan, J.A., Mattes, E., Tan, H., Dinan, A., Newnham, J.P., Whitehouse, A.J., Jacoby,
637 P., Hickey, M., 2012. Androgen concentrations in umbilical cord blood and their
638 association with maternal, fetal and obstetric factors. *PLoS One* 7, e42827

639 Kelce, W.R., Stone, C. R., Laws, S. C., Gray, L.E., Kemppainen J.A., Wislson, E.M.,
640 1995. Persistent DDT metabolite p,p'-DDE is a potent androgen receptor
641 antagonist. *Nature*, 375, 581-585.

642 Kishi, R., Araki, A., Minatoya, M., Hanaoka, T., Miyashita, C., Itoh, S., Kobayashi, S.,
643 Ait Bamai, Y., Yamazaki, K., Miura, R., Tamura, N., Ito, K., Goudarzi H.,
644 Members of The Hokkaido Study on Environment and Children's Health, 2017.
645 The Hokkaido Birth Cohort Study on Environment and Children's Health: cohort
646 profile—updated 2017. *Environ. Health Prev. Med.* 22, 46–62.

647 Kishi, R., Kobayashi, S., Ikeno, T., Araki, A., Miyashita, C., Itoh, S., Sasaki, S., Okada,
648 E., Kobayashi, S., Kashino, I., Itoh, K., Nakajima, S.; Members of the Hokkaido
649 Study on Environment and Children's Health, 2013. Ten years of progress in the
650 Hokkaido birth cohort study on environment and children's health: cohort
651 profile—updated 2013. *Environ. Health Prev. Med.* 18, 429–450.

652 Kishi, R., Sasaki, S., Yoshioka, E., Yuasa, M., Sata, F., Saijo, Y., Kurahashi, N.,
653 Tamaki, J., Endo, T., Sengoku, K., Nonomura, K., Minakami, H.; Hokkaido Study
654 on Environment and Children's Health, 2011. Cohort profile: the Hokkaido Study
655 on Environment and Children's Health in Japan. *Int. J. Epidemiol.* 40, 611–618.

1712
1713
1714
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1717
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1767
1768
1769
1770

656 Kojima, H., Katsura, E., Takeuchi, S., Niiyama, K., Kobayashi, K., 2004. Screening for
657 estrogen and androgen receptor activities in 200 pesticides by in vitro reporter
658 gene assays using Chinese hamster ovary cells. *112*, 524-531.

659 Kuiri-Hänninen, T., Sankilampi, U., Dunkel, L., 2014. Activation of the hypothalamic-
660 pituitary-gonadal axis in infancy: minipuberty. *Horm. Res. Paediatr.* 82, 73–80.

661 Labrie, F., Luu-The, V., Labrie, C., Simard, J., 2001. DHEA and its transformation into
662 androgens and estrogens in peripheral target tissues: intracrinology. *Front.*
663 *Neuroendocrinol.* 22, 185–212.

664 Longnecker, M.P., Gladen, B.C., Cupul-Uicab, L.A., Romano-Riquer, S.P., Weber, J.P.,
665 Chapin, R.E., Hernández-Avila, M., 2007. In utero exposure to the antiandrogen
666 1,1-dichloro-2,2-bis(p-chlorophenyl)ethylene (DDE) in relation to anogenital
667 distance in male newborns from Chiapas, México. *Am. J. Epidemiol.* 165, 1015–
668 1022.

669 Mastorakos G, Ilias I. 2003. Maternal and fetal hypothalamic-pituitary-adrenal axes
670 during pregnancy and postpartum. *Ann NY Acad Sci* 997:136–149.

671 Mnif, W., Ibn Hadj Hassin, A., Bouaziz, A., Bartegi, A., Thomas, O., Roig, B., 2011.
672 Effect of Endocrine Disruptor Pesticides: A Review. *Int. J. Environ. Res. Public*
673 *Health* 8, 2265-2303.

674 Padvi, N.V., Narkhede, H.R., Pawar A.P., Mhatre, P.N., 2017. Prolactin level in
675 umbilical cord blood of newborn and its relation to respiratory distress syndrome.
676 *Int. J. Reprod. Contracept. Obstet. Gynecol.* 6, 433-438.

677 Parker Jr., C.R., MadDonald, P.C., Guzick, D.S., Porter, J.C., Rosenfeld, C.R., Hauth,
678 J.C., 1989. Prolactin levels in umbilical cord blood of human infants: Relation to

1771
1772
1773
1774
1775 679 gestational age, maternal complications, and neonatal lung function. *Am J Obstet*
1776
1777
1778 680 *Gynecol.* 161, 7950-802.
1779
1780 681 Orth, J.M., Boehm, R., 1990. Functional coupling of neonatal rat Sertoli cells and
1781
1782 682 gonocytes in coculture. *Endocrinology.* 127, 2812–2820.
1783
1784 683 Reynolds, R.M., 2010. Corticosteroid-mediated programming and the pathogenesis of
1785
1786 684 obesity and diabetes. *J. Steroid. Biochem. Mol. Biol.* 122, 3–9.
1787
1788 685 Rignell-Hydbom, A., Rylander, L., Giwercman, A., Jönsson, B.A., Nilsson-Ehle, P.,
1789
1790 686 Hagmar, L., 2004. Exposure to CB-153 and p,p'-DDE and male reproductive
1791
1792 687 function. *Hum. Reprod.* 19, 2066–2075.
1793
1794 688 Strauss III, J. F., Chapter 4 The synthesis and Metabolism of Steroid Hormones, in:
1795
1796 689 Strauss III, J.F., Barbieri, R.L. (Eds.), *Yen and Jaffe's Reproductive*
1797
1798 690 *Endocrinology, Physiology, Pathophysiology, and Clinical Management*, 6th
1799
1800 691 edition. Elsevier, Philadelphia, PA, USA, pp. 105-120.
1801
1802 692 The United Nations Environment Programme (UNEP) and World Health Organization
1803
1804 693 (WHO), 2012. *State of the science of endocrine disrupting chemicals*. Edited by
1805
1806 694 Åke Bergman, Jerrold J. Heindel, Susan Jobling, Karen A. Kidd and R. Thomas
1807
1808 695 Zoeller. World Health Organization Press, Geneva. Available at
1809
1810 696 http://www.who.int/iris/bitstream/10665/78101/1/9789241505031_eng.pdf?ua=1
1811
1812 697 [accessed 25 May 2017].
1813
1814 698 The United Nations Environment Programme (UNEP). Available at
1815
1816 699 <http://chm.pops.int/default.aspx> [accessed 25 May 2017].
1817
1818 700 Torres-Sanchez, L., Zepeda, M., Cebrián, M.E., Belkind-Gerson, J., Garcia-Hernandez,
1819
1820 701 R.M., Belkind-Valdovinos, U., López-Carrillo, L., 2008.
1821
1822 702 *Dichlorodiphenyldichloroethylene Exposure during the First Trimester of*

1830
1831
1832
1833
1834 703 Pregnancy Alters the Anal Position in Male Infants. *Ann N Y Acad Sci*, 1140,
1835
1836 704 155-162.
1837
1838
1839 705 Vaithinathan, S., Saradha, B., Mathur, P.P., 2008. Transient inhibitory effect of
1840
1841 706 methoxychlor on testicular steroidogenesis in rat: an in vivo study. *Arch. Toxicol.*
1842
1843 707 82, 833–839.
1844
1845 708 Vested, A., Ramlau-Hansen, C.H., Olsen, S.F., Bonde, J.P., Støvring, H., Kristensen,
1846
1847 709 S.L., Halldorsson, T.I., Rantakokko, P., Kiviranta, H., Ernst, E.H., Toft, G., 2014.
1848
1849 710 In utero exposure to persistent organochlorine pollutants and reproductive health
1850
1851 711 in the human male. *Reproduction*. 148, 635–646.
1852
1853 712 Warembourg, C., Debost-Legrand, A., Bonvallot, N., Massart, C., Garlantézec, R.,
1854
1855 713 Monfort, C., Gaudreau, E., Chevrier, C., Cordier, S., 2016. Exposure of pregnant
1856
1857 714 women to persistent organic pollutants and cord sex hormone levels. *Hum.*
1858
1859 715 *Reprod.* 31, 190–198.
1860
1861 716 Wilson, C.M. McPhaul, M.J., 1996. A and B forms of the androgen receptor are
1862
1863 717 expressed in a variety of human tissues. *Mol Cell Endocrinol*, 120, 51-57.
1864
1865 718 Windham, G.C., Lee, D., Mitchell, P., Anderson, M., Petreas, M., Lasley, B., 2005.
1866
1867 719 Exposure to organochlorine compounds and effects on ovarian function.
1868
1869 720 *Epidemiology*. 16, 182–190.
1870
1871 721 Yamashita, K., Okuyama, M., Watanabe, Y., Honma, S., Kobayashi, S., Numazawa,
1872
1873 722 M., 2007a. Highly sensitive determination of estrone and estradiol in human
1874
1875 723 serum by liquid chromatography–electrospray ionization tandem mass
1876
1877 724 spectrometry. *Steroids*. 72, 819–827.
1878
1879 725 Yamashita, K., Takahashi, M., Tsukamoto, S., Numazawa, M., Okuyama, M., Honma,
1880
1881 726 S., 2007b. Use of novel picolinoyl derivatization for simultaneous quantification
1882
1883
1884
1885
1886
1887
1888

1889
1890
1891
1892
1893
1894
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1946
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727 of six corticosteroids by liquid chromatography-electrospray ionization tandem
728 mass spectrometry. *J. Chromatogr. A.* 1173, 120–128.

729 Ye, M., Beach, J., Martin, J.W., Senthilselvan, A., 2015. Association between Lung
730 Function in Adults and Plasma DDT and DDE Levels: Results from the Canadian
731 Health Measures Survey. *Environ. Health Perspect.* 123, 422-427.

732 You, L., Sar, M., Bartolucci, E., Ploch, S., Whitt, M., 2001. Induction of hepatic
733 aromatase by p,p'-DDE in adult male rats. *Mol. Cell Endocrinol.* 178, 207-214.

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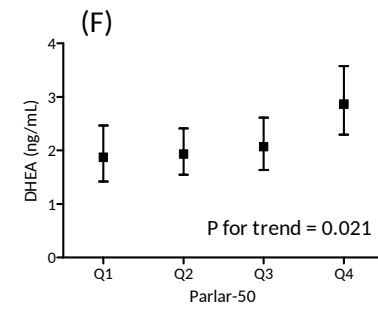
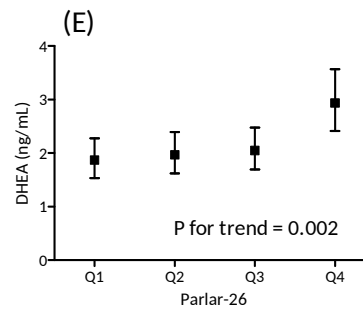
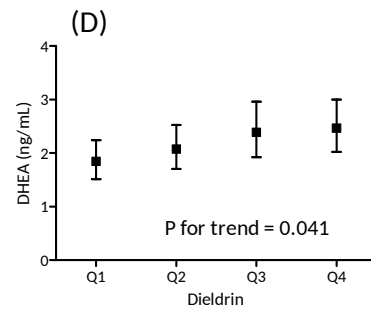
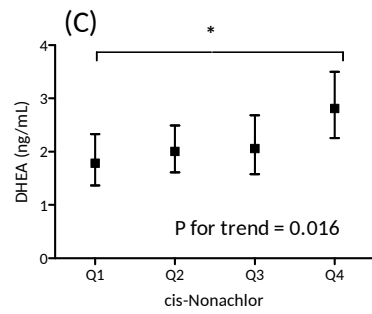
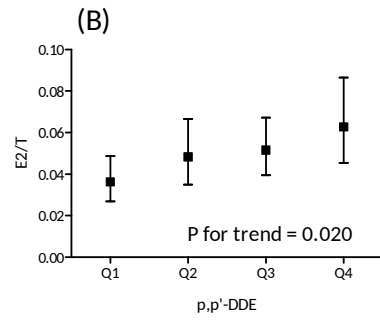
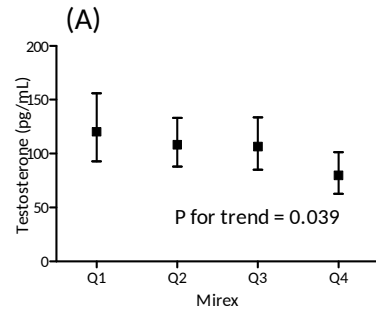
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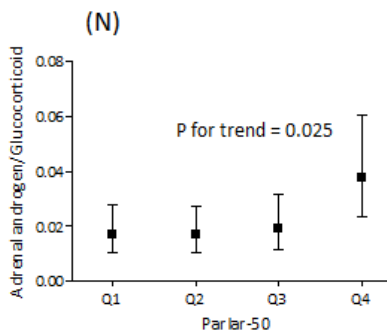
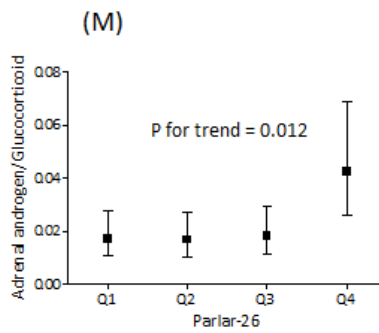
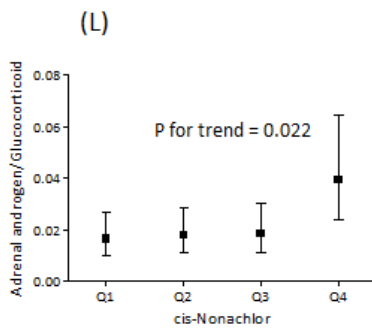
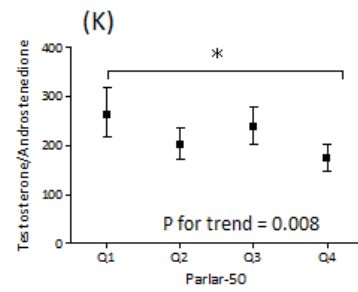
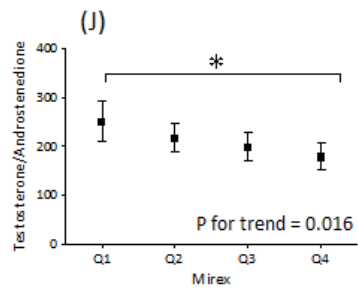
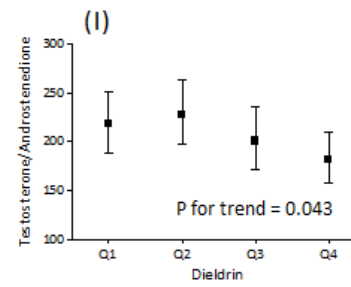
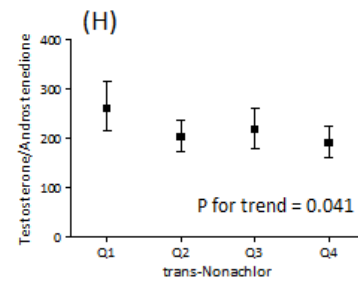
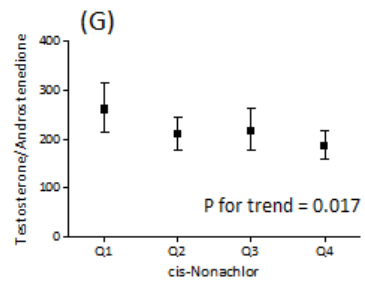
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737 Fig. 1. Least square means (LSMs) of hormone levels according to maternal
738 organochlorine pesticide (OCP) concentration quartiles in boys. The X-axes show the
739 OCP quartiles, while the Y-axes show each hormone level calculated using the LSM in
740 the boxes; the error bars are the 95% confidence intervals. The four OCP categories
741 were, *cis*-nonachlor: Quartile 1 (Q1) (≤ 7.07 pg/g-wet), Q2 (7.08–10.37 pg/g-wet), Q3
742 (10.38–15.06 pg/g-wet), and Q4 (≥ 15.07 pg/g-wet); *trans*-nonachlor: Q1 (≤ 52.25 pg/g-
743 wet), Q2 (52.26–75.60 pg/g-wet), Q3 (52.26–75.60 pg/g-wet), and Q4 (≥ 110.29 pg/g-
744 wet); *p,p'*-DDE: Q1 (≤ 0.99 pg/g-wet), Q2 (1.00–1.65 pg/g-wet), Q3 (1.66–2.54 pg/g-
745 wet), and Q4 (≥ 2.55 pg/g-wet); *o,p'*-DDT: Q1 (≤ 2.28 pg/g-wet), Q2 (2.29–3.36 pg/g-
746 wet), Q3 (3.37–4.66 pg/g-wet), and Q4 (≥ 4.67 pg/g-wet); Dieldrin: Q1 (≤ 12.17 pg/g-
747 wet), Q2 (12.17–16.68 pg/g-wet), Q3 (16.69–22.05 pg/g-wet), and Q4 (≥ 22.06 pg/g-
748 wet); β -hexachlorocyclohexane (β -HCH): Q1 (≤ 104.33 pg/g-wet), Q2 (104.34–154.31
749 pg/g-wet), Q3 (154.32–238.06 pg/g-wet), and Q4 (≥ 238.07 pg/g-wet); Mirex: Q1 (≤ 4.12
750 pg/g-wet), Q2 (4.13–6.04 pg/g-wet), Q3 (6.05–8.52 pg/g-wet), and Q4 (≥ 8.53 pg/g-
751 wet); Parlar-26: Q1 (≤ 2.84 pg/g-wet), Q2 (2.85–4.46 pg/g-wet), Q3 (4.47–7.11 pg/g-
752 wet), and Q4 (≥ 7.12 pg/g-wet); and Parlar-50: Q1 (≤ 4.31 pg/g-wet), Q2 (4.31–6.56
753 pg/g-wet), Q3 (6.56–9.83 pg/g-wet), and Q4 (≥ 9.83 pg/g-wet). (A) Testosterone
754 according to Mirex, (B) Estradiol-testosterone ratio (E2/T) according to *p,p'*-
755 dichlorodiphenyldichloroethylene (DDE), (C) dehydroepiandrosterone (DHEA)
756 according to *cis*-nonachlor, (D) DHEA according to Dieldrin, (E) DHEA according to
757 Parlar-26, (F) DHEA according to Parlar-50, (G) Testosterone-androstenedione ratio
758 (T/Adione) according to *cis*-Nonachlor, (H) T/Adione according to *trans*-Nonachlor, (I)

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759 T/Adione according to Dieldrin, (J) T/Adione according to Mirex, (K) T/Adione
760 according to Parlar-50, (L) Adrenal androgen-glucocorticoid ratio according to *cis*-
761 Nonachlor, (M) Adrenal androgen-glucocorticoid ratio according to Parlar-26, (N)
762 Adrenal androgen-glucocorticoid-ratio according to Parlar-50, (O) sex hormone-binding
763 globulin (SHBG) according to β -HCH, (P) Prolactin according to *p,p'*-DDE, (Q)
764 Prolactin according to *o,p'*-DDT. The first quartile is compared to the second, third, and
765 fourth quartile OCPs as calculated using the Dunnett-Hsu method; the statistical
766 significance of the *P* value was $*P<0.017$ based on Bonferroni's correction. LSMs were
767 adjusted for maternal age, parity, and gestational age.
768





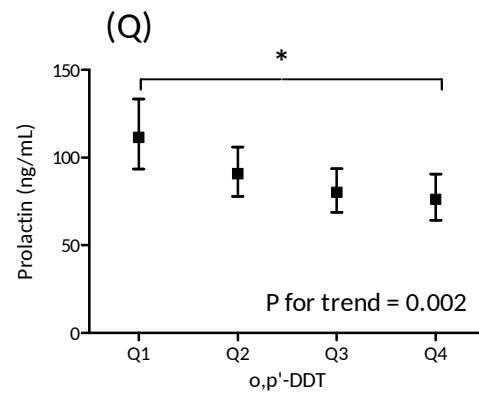
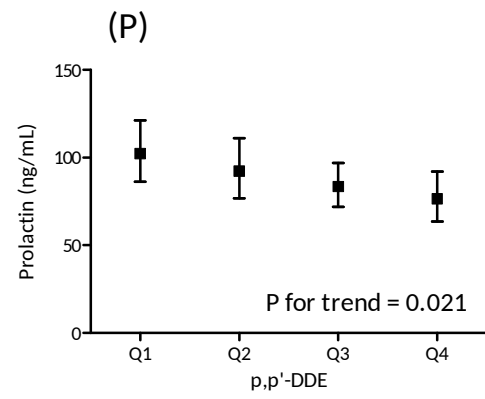
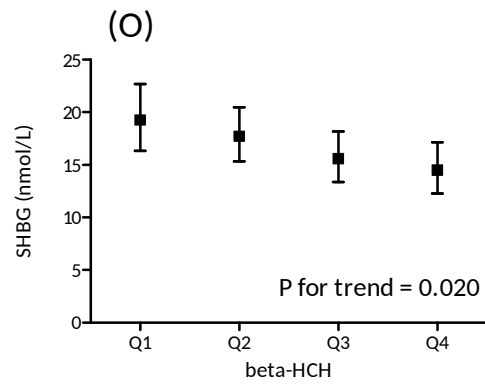


Fig. 1.

Table 1 Maternal and infant characteristics

			This study (n=232)			Original cohort (n=514)		
			No.	%	Mean ± SD	No.	%	Mean ± SD
Mother								
Age at delivery (years)			232		30.45 ± 4.81	510		30.4 ± 4.9
Pre-pregnancy BMI (kg/m ²)			232		21.03 ± 2.92	506		21.2 ± 3.2
Educational level (years)	≤12		100	43.1		225	44.3	
	>12		132	56.9		283	55.7	
Annual Household income (million yen per year)	<5		166	71.6		345	68.3	
	≥5		66	28.4		160	31.7	
Smoking during pregnancy	No		190	81.9		404	79.7	
	Yes		42	18.1		103	20.3	
Alcohol consumption during pregnancy	No		154	66.4		351	69.1	
	Yes		78	33.6		157	30.9	
Parity	Primiparous		120	51.7		240	47.7	
	Multiparous		112	48.3		263	52.2	
Type of delivery	Vaginal		230	99.1		397	78.8	
	Caesarian section		2	0.9		107	21.2	
Blood sampling timing	During pregnancy		159	68.5		296	69.5	
	After delivery		73	31.5		130	30.5	
Infant								
Sex	Boys		106	45.7		246	48.1	
	Girls		126	54.3		265	51.9	
Birth weight (grams)			232		3130.5 ± 332.5	511		3025.6 ± 420.7
Gestational Age (weeks)			232		39.3 ± 1.1	511		38.9 ± 1.5
cryptorchidism			1	0.9		1	0.4	
hypospadias			0	0.0		0	0.0	

Table 2 Concentrations of organochlorine pesticides in maternal blood

Persistent organochlorine pesticides	Detection limit (pg/g-wet)	>DL (%)	Minimum	Percentile			Maximum
				25th	50th	75th	
Aldrin	1.00	0.4				<DL	12.83
Chlordanes							
oxychlordane	0.90	100.0	7.93	28.87	40.04	57.32	250.94
<i>cis</i> -Chlordane	0.70	62.1		<DL	1.15	2.29	17.53
<i>trans</i> -Chlordane	0.50	49.6			<DL	0.84	3.79
<i>cis</i> -Nonachlor	0.40	100.0	1.63	7.07	10.37	15.07	37.58
<i>trans</i> -Nonachlor	0.50	100.0	13.45	52.09	75.60	110.54	513.52
DDTs							
<i>o,p'</i> -DDD	0.50	14.2				<DL	1.16
<i>p,p'</i> -DDD	0.40	88.8	<DL	0.98	1.65	2.54	9.04
<i>o,p'</i> -DDE	0.40	86.6	<DL	0.72	1.25	1.78	4.60
<i>p,p'</i> -DDE	0.60	100.0	99.52	409.79	619.26	968.05	2686.23
<i>o,p'</i> -DDT	0.60	96.6	<DL	2.28	3.36	4.67	17.15
<i>p,p'</i> -DDT	0.40	100.0	2.38	16.22	23.17	33.94	104.76
Dieldrin	0.80	100.0	4.11	12.16	16.68	22.21	71.52
Endrin	1.00	0.0					<DL
Heptachlors							
Heptachlor	0.80	0.9				<DL	1.14
<i>trans</i> -HCE	1.00	0.0					<DL
<i>cis</i> -HCE	0.40	100.0	6.17	18.81	26.25	37.45	200.53
HCB	0.90	100.0	34.94	83.04	103.99	131.61	245.48
HCHs							
α-HCH	0.70	68.5		<DL	0.91	1.31	3.10
β-HCH	0.60	100.0	19.95	104.25	154.31	238.45	717.67
γ-HCH	0.90	57.3		<DL	1.09	1.73	100.92
δ-HCH	0.70	1.3				<DL	1.11
Mirex	0.50	100.0	0.88	4.11	6.04	8.53	30.11
Toxaphenes							
Parlar-26	1.00	97.0	<DL	2.84	4.46	7.13	20.82
Parlar-41	0.70	28.4			<DL	0.73	1.96
Parlar-40	2.00	0.9				<DL	2.43
Parlar-44	2.00	2.2				<DL	2.77
Parlar-50	2.00	96.1	<DL	4.30	6.56	9.86	29.29
Parlar-62	6.00	0.0					<DL

DL, detection limit; DDD, dichlorodiphenyldichloroethane; DDE, dichlorodiphenyldichloroethylene; DDT, dichlorodiphenyltrichloroethane; HCB, hexachlorobenzene; HCE, heptachlor epoxide; HCH, hexachlorocyclohexane.

Table 3 Distribution of steroids and reproductive hormones

	DL	n	>DL (%)	Total (n=232)			Boys (n=106)					Girls (n=126)					P-value
				25%	Med	75%	n	>DL (%)	25%	Med	75%	n	>DL (%)	25%	Med	75%	
Steroid hormones																	
Progesterone (ng/mL)	0.01	232	100.0	177.381	219.036	278.019	106	100.0	184.087	224.121	280.633	126	100.0	169.619	216.065	276.129	0.460
Testosterone (pg/mL)	0.01	232	100.0	59.750	82.050	114.025	106	100.0	76.788	101.875	132.050	126	100.0	50.975	68.675	92.400	<0.001
Estradiol (ng/mL)	0.01	232	99.6	3.220	4.546	7.053	106	99.1	3.248	4.458	7.311	126	100.0	3.140	4.645	6.672	0.393
DHEA (ng/mL)	0.01	232	100.0	1.763	2.160	2.970	106	100.0	1.590	2.090	2.830	126	100.0	1.898	2.270	3.140	0.040
Androstenedione (ng/mL)	0.01	232	99.6	0.360	0.460	0.588	106	99.1	0.380	0.470	0.613	126	100.0	0.350	0.450	0.580	0.414
Cortisol (ng/mL)	0.25	232	98.7	22.808	41.485	65.258	106	99.6	22.415	41.065	66.213	126	98.4	24.725	42.455	63.953	0.773
Cortisone (ng/mL)	0.10	232	96.1	76.393	95.145	124.655	106	98.1	72.825	94.175	122.963	126	94.4	77.150	95.650	125.123	0.807
Steroid Hormone Binding Globulin (nmol/L)	1.10	232	99.6	13.500	15.700	18.875	106	100.0	13.575	16.200	19.300	126	99.2	13.300	15.450	18.425	0.239
Lutealizing Hormone (mIU/mL)	0.50	226	17.3			<DL	103	36.9		<DL	0.870	123	0.8			<DL	<0.001
Follicle Stimulating Hormone (mIU/mL)	0.50	225	22.2			<DL	103	48.5		<DL	0.660	122	0.0			<DL	<0.001
Inhibin B (pg/mL)	11	232	61.6	<DL	23.500	45.325	106	98.1	34.200	44.350	61.300	126	31.0		<DL	13.900	<0.001
Insulin-like factor 3 (ng/mL)	0.01	119	100.0	0.230	0.280	0.330	103	100.0	0.250	0.290	0.340	16	100.0	0.180	0.185	0.235	<0.001
Prolactin (ng/mL)	1.0	226	99.6	64.600	87.600	118.250	103	100.0	63.400	87.200	119.000	123	99.2	64.600	87.800	116.000	0.921

P values were calculated by Mann-Whitney U test;

DHEA, dehydroepiandrosterone; DL, detection limit

Table 4 Associations between OCPs exposure and steroid and reproductive hormone levels among boys

	Progesterone			Estradiol			Testosterone			Estradiol/Testosterone			DEHA			Androstenedione			Androstenedione/DHEA								
	β	95%CI		β	95%CI		β	95%CI		β	95%CI		β	95%CI		β	95%CI		β	95%CI							
Oxychlorthane	-0.090	-0.416	0.236	0.007	-0.359	0.372	-0.176	-0.421	0.068	0.183	-0.140	0.505	0.243	0.028	0.459	*	0.006	-0.214	0.226	-0.238	-0.470	-0.005	*				
cis-Nonachlor	-0.149	-0.424	0.125	-0.005	-0.314	0.304	-0.095	-0.303	0.113	0.090	-0.184	0.363	0.247	0.066	0.427	**	0.070	-0.116	0.255	-0.177	-0.375	0.021	†				
trans-Nonachlor	-0.162	-0.436	0.112	-0.027	-0.335	0.282	-0.161	-0.367	0.045	0.134	-0.138	0.407	0.255	0.076	0.435	**	0.038	-0.147	0.224	-0.217	-0.413	-0.021	*				
p,p'-DDD	-0.097	-0.279	0.085	-0.151	-0.354	0.051	-0.065	-0.203	0.073	-0.086	-0.267	0.095	-0.023	-0.147	0.101		-0.016	-0.139	0.108	0.007	-0.126	0.141					
o,p'-DDE	-0.046	-0.234	0.143	0.050	-0.161	0.260	-0.029	-0.171	0.114	0.078	-0.108	0.265	0.080	-0.046	0.207		0.013	-0.114	0.140	-0.067	-0.204	0.069					
p,p'-DDE	0.034	-0.218	0.286	0.192	-0.088	0.472	-0.105	-0.295	0.084	0.297	0.054	0.541	*	0.108	-0.061	0.278		-0.041	-0.211	0.128	-0.150	-0.331	0.032				
o,p'-DDT	-0.056	-0.248	0.136	0.048	-0.167	0.263	-0.035	-0.180	0.110	0.083	-0.107	0.273		0.118	-0.010	0.246	†	0.010	-0.119	0.140	-0.108	-0.246	0.031				
p,p'-DDT	-0.077	-0.339	0.184	0.031	-0.262	0.324	-0.065	-0.263	0.132	0.096	-0.163	0.356		0.151	-0.024	0.325	†	0.016	-0.161	0.192	-0.135	-0.324	0.054				
Dieldrin	-0.178	-0.482	0.125	0.069	-0.273	0.410	-0.023	-0.254	0.208	0.092	-0.211	0.394		0.267	0.067	0.467	**	0.155	-0.048	0.359	-0.112	-0.333	0.109				
cis-HCE	-0.090	-0.416	0.236	0.007	-0.359	0.372	-0.176	-0.421	0.068	0.183	-0.140	0.505		0.243	0.028	0.459	*	0.006	-0.214	0.226	-0.238	-0.470	-0.005	*			
HCB	0.025	-0.438	0.489	0.114	-0.404	0.633	-0.139	-0.489	0.210	0.254	-0.203	0.711		0.311	0.004	0.619	*	0.058	-0.254	0.370	-0.253	-0.586	0.081				
β -HCH	0.118	-0.157	0.392	0.168	-0.139	0.474	-0.091	-0.299	0.117	0.259	-0.010	0.528	†	0.136	-0.049	0.320		-0.004	-0.190	0.181	-0.140	-0.339	0.059				
Mirex	-0.108	-0.418	0.203	-0.007	-0.355	0.342	-0.262	-0.492	-0.032	*	0.256	-0.050	0.561		0.213	0.007	0.420	*	-0.061	-0.270	0.149	-0.274	-0.494	-0.054	*		
Parlar-26	-0.124	-0.338	0.091	0.124	-0.116	0.364	0.002	-0.161	0.165		0.122	-0.090	0.335		0.235	0.097	0.373	**	0.100	-0.044	0.244	-0.135	-0.290	0.020	†		
Parlar-50	-0.099	-0.315	0.116	0.072	-0.170	0.313	-0.020	-0.183	0.144		0.091	-0.123	0.305		0.236	0.097	0.374	**	0.103	-0.041	0.248	-0.132	-0.288	0.023	†		
	Testosterone/Androstenedione			Cortisol			Cortisone			Cortisone/Cortisol			Adrenal androgen/Glucocorticoid			SHBG			Prolactin								
	β	95%CI		β	95%CI		β	95%CI		β	95%CI		β	95%CI		β	95%CI		β	95%CI							
Oxychlorthane	-0.182	-0.339	-0.024	*	-0.363	-0.763	0.038	-0.398	-0.858	0.062	†	-0.035	-0.316	0.246	0.571	0.042	1.100	*	-0.082	-0.245	0.081	-0.077	-0.261	0.107			
cis-Nonachlor	-0.165	-0.297	-0.032	*	-0.218	-0.559	0.124	-0.310	-0.700	0.079		-0.092	-0.330	0.145	0.487	0.040	0.935	*	-0.040	-0.178	0.098	-0.150	-0.303	0.003	†		
trans-Nonachlor	-0.199	-0.330	-0.069	**	-0.372	-0.708	-0.036	*	-0.426	-0.810	-0.041	*	-0.054	-0.291	0.184	0.609	0.168	1.050	**	-0.027	-0.165	0.111	-0.075	-0.230	0.080		
p,p'-DDD	-0.049	-0.140	0.041		0.061	-0.166	0.289	0.014	-0.248	0.275		-0.048	-0.205	0.110	-0.049	-0.352	0.255		0.055	-0.036	0.146	0.021	-0.082	0.124			
o,p'-DDE	-0.042	-0.134	0.051		-0.108	-0.342	0.126	-0.148	-0.415	0.119		-0.040	-0.202	0.122	0.200	-0.109	0.510		-0.030	-0.124	0.064	-0.120	-0.224	-0.015	*		
p,p'-DDE	-0.064	-0.188	0.060		-0.290	-0.600	0.019	†	-0.249	-0.606	0.108		0.041	-0.176	0.258	0.354	-0.059	0.766	†	-0.080	-0.206	0.045	-0.226	-0.361	-0.092	**	
o,p'-DDT	-0.045	-0.140	0.049		-0.172	-0.410	0.065		-0.153	-0.426	0.120		0.019	-0.146	0.185	0.260	-0.055	0.574		-0.017	-0.114	0.079	-0.172	-0.275	-0.069	**	
p,p'-DDT	-0.081	-0.210	0.048		-0.073	-0.400	0.253		-0.192	-0.565	0.180		-0.119	-0.344	0.105	0.276	-0.155	0.707		-0.019	-0.150	0.112	-0.155	-0.299	-0.011	*	
Dieldrin	-0.178	-0.325	-0.032	*	0.028	-0.352	0.409		-0.207	-0.641	0.227		-0.235	-0.494	0.023	†	0.367	-0.134	0.867		-0.049	-0.202	0.103	-0.209	-0.377	-0.040	*
cis-HCE	-0.182	-0.339	-0.024	*	-0.363	-0.763	0.038	†	-0.398	-0.858	0.062	†	-0.035	-0.316	0.246		0.571	0.042	1.100	*	-0.082	-0.245	0.081	-0.077	-0.261	0.107	
HCB	-0.198	-0.424	0.028	†	-0.255	-0.830	0.321		-0.185	-0.845	0.476		0.070	-0.329	0.469	0.474	-0.289	1.236		-0.167	-0.397	0.062	-0.239	-0.495	0.017	†	
β -HCH	-0.087	-0.222	0.048		-0.153	-0.495	0.189		-0.076	-0.469	0.317		0.077	-0.160	0.314	0.221	-0.234	0.676		-0.147	-0.282	-0.012	*	-0.182	-0.332	-0.031	*
Mirex	-0.202	-0.350	-0.053	**	-0.588	-0.959	-0.218	**	-0.572	-1.002	-0.142	*	0.016	-0.252	0.285	0.744	0.249	1.239	**	-0.083	-0.238	0.073	-0.180	-0.352	-0.008	*	
Parlar-26	-0.098	-0.203	0.007	†	-0.174	-0.441	0.092		-0.289	-0.592	0.013	†	-0.115	-0.299	0.069	0.457	0.111	0.803	*	-0.017	-0.125	0.091	-0.107	-0.229	0.015	†	
Parlar-50	-0.123	-0.227	-0.019	*	-0.156	-0.424	0.111		-0.231	-0.537	0.075		-0.075	-0.260	0.111	0.418	0.069	0.767	*	-0.048	-0.156	0.060	-0.131	-0.252	-0.010	*	
	LH			FSH			Inhibin B			INSL3			T/LH			FSH/InhibinB											
	β	95%CI		β	95%CI		β	95%CI		β	95%CI		β	95%CI		β	95%CI										
Oxychlorthane	0.037	-0.305	0.380		0.097	-0.143	0.337		-0.093	-0.301	0.115		-0.064	-0.190	0.063	-0.224	-0.684	0.236	0.190	-0.119	0.499						
cis-Nonachlor	0.087	-0.202	0.376		0.199	0.000	0.399	†	-0.140	-0.315	0.034		-0.015	-0.123	0.092	-0.185	-0.574	0.203	0.341	0.086	0.595	**					
trans-Nonachlor	0.077	-0.211	0.366		0.121	-0.080	0.323		-0.115	-0.290	0.060		-0.061	-0.167	0.045	-0.245	-0.632	0.141	0.236	-0.023	0.494	†					
p,p'-DDD	-0.045	-0.236	0.146		0.057	-0.077	0.191		-0.050	-0.167	0.067		-0.044	-0.114	0.027	-0.023	-0.281	0.235	0.105	-0.068	0.277						
o,p'-DDE	-0.042	-0.240	0.156		-0.012	-0.151	0.127		-0.090	-0.209	0.029		0.003	-0.070	0.077	0.019	-0.248	0.287	0.086	-0.093	0.265						
p,p'-DDE	0.084	-0.179	0.347		0.145	-0.038	0.328		-0.127	-0.286	0.033		-0.069	-0.165	0.028	-0.184	-0.538	0.169	0.284	0.052	0.517	*					
o,p'-DDT	0.047	-0.154	0.248		-0.005	-0.146	0.137		-0.077	-0.199	0.045		-0.003	-0.078	0.071	-0.075	-0.346	0.196	0.080	-0.102	0.262						
p,p'-DDT	-0.014	-0.288	0.259		0.159	-0.031	0.349		-0.156	-0.320	0.009	†	-0.014	-0.116	0.087	-0.061	-0.430	0.308	0.307	0.066	0.548	*					
Dieldrin	-0.040	-0.362	0.282		0.071	-0.155	0.297		-0.183	-0.375	0.009	†	-0.048	-0.167	0.071	0.011	-0.424	0.446	0.246	-0.043	0.535	†					
cis-HCE	0.037	-0.305	0.380		0.097	-0.143	0.337		-0.093	-0.301	0.115		-0.064	-0.190	0.063	-0.224	-0.684	0.236	0.190	-0.119	0.499						
HCB	-0.116	-0.598	0.367		0.213	-0.124	0.549		-0.290	-0.581	0.000	†	-0.041	-0.220	0.138	-0.032	-0.683	0.620	0.496	0.069	0.924	*					
β -HCH	0.030	-0.257	0.317		0.203	0.006	0.401	*	-0.130	-0.304	0.045		-0.028	-0.134	0.078	-0.130	-0.516	0.256	0.324	0.072	0.577	*					
Mirex	0.137	-0.188	0.461		0.229	0.004	0.453	*	-0.069	-0.267	0.130		-0.061	-0.181	0.060	-0.401	-0.833	0.031	0.299	0.009	0.589	*					
Parlar-26	0.075	-0.154	0.304		0.164	0.006	0.323	*	-0.099	-0.235	0.038		-0.058	-0.142	0.027	-0.073	-0.383	0.237	0.267	0.065	0.469	*					
Parlar-50	0.033	-0.197	0.262		0.173	0.015	0.331	*	-0.124	-0.260	0.012	†	-0.028	-0.113	0.057	-0.056	-0.365	0.254	0.297	0.097	0.497	**					

Adjusted regression coefficients (β) and 95% confidence intervals (CIs) for the association between a 10-fold increase of OCP and log10-transformed hormone levels as determined by a linear regression model

Each OCP was introduced into the model separately and adjusted for maternal age, parity, and gestational age.

*P < 0.05, **P < 0.01, †P < 0.1

Table 5 Associations between OCPs exposure and steroid and reproductive hormone levels among girls

	Progesterone			Estradiol			Testosterone			Estradiol/Testosterone			DEHA			Androstenedione						
	β	95%CI		β	95%CI		β	95%CI		β	95%CI		β	95%CI		β	95%CI					
Oxychlorane	0.136	-0.149	0.420	0.047	-0.175	0.270	0.050	-0.219	0.320	-0.003	-0.222	0.216	-0.087	-0.377	0.203	0.032	-0.189	0.254				
cis-Nonachlor	0.194	-0.069	0.456	0.019	-0.188	0.225	0.035	-0.215	0.284	-0.016	-0.220	0.187	-0.136	-0.404	0.132	0.006	-0.199	0.211				
trans-Nonachlor	0.192	-0.065	0.449	0.035	-0.167	0.237	0.065	-0.179	0.310	-0.030	-0.229	0.169	-0.163	-0.425	0.099	0.038	-0.163	0.239				
p,p'-DDD	0.099	-0.046	0.244	-0.059	-0.172	0.054	-0.041	-0.178	0.096	-0.018	-0.130	0.094	-0.160	-0.305	-0.014	*	-0.069	-0.181	0.043			
o,p'-DDE	-0.042	-0.212	0.127	-0.082	-0.214	0.049	-0.038	-0.198	0.122	-0.044	-0.174	0.086	0.000	-0.173	0.173	-0.066	-0.197	0.065				
p,p'-DDE	0.018	-0.201	0.237	-0.116	-0.285	0.054	0.035	-0.172	0.241	-0.151	-0.317	0.015	†	-0.156	-0.377	0.065	-0.077	-0.246	0.092			
o,p'-DDT	-0.071	-0.282	0.140	-0.109	-0.272	0.055	-0.024	-0.222	0.175	-0.085	-0.246	0.076	0.013	-0.202	0.227	-0.072	-0.235	0.091				
p,p'-DDT	0.042	-0.205	0.289	-0.063	-0.256	0.129	0.075	-0.158	0.308	-0.138	-0.327	0.050	-0.102	-0.353	0.149	-0.041	-0.233	0.150				
Dieldrin	-0.012	-0.310	0.287	-0.028	-0.261	0.204	-0.092	-0.373	0.188	0.064	-0.165	0.293	0.068	-0.235	0.371	-0.204	-0.432	0.024	†			
cis-HCE	0.136	-0.149	0.420	0.047	-0.175	0.270	0.050	-0.219	0.320	-0.003	-0.222	0.216	-0.087	-0.377	0.203	0.032	-0.189	0.254				
HCB	0.216	-0.174	0.606	-0.052	-0.357	0.253	0.026	-0.344	0.395	-0.077	-0.378	0.223	-0.052	-0.450	0.346	0.051	-0.252	0.354				
β -HCH	-0.032	-0.286	0.222	-0.050	-0.248	0.147	0.057	-0.182	0.296	-0.107	-0.301	0.087	-0.016	-0.274	0.242	-0.006	-0.203	0.190				
Mirex	0.218	-0.068	0.503	0.043	-0.182	0.267	-0.030	-0.302	0.241	0.073	-0.148	0.294	-0.166	-0.457	0.126	0.025	-0.198	0.248				
Parlar-26	0.031	-0.167	0.229	-0.006	-0.160	0.149	-0.012	-0.198	0.175	0.006	-0.146	0.158	0.022	-0.180	0.223	-0.029	-0.182	0.124				
Parlar-50	0.046	-0.177	0.270	-0.012	-0.186	0.162	-0.002	-0.213	0.209	-0.010	-0.182	0.162	0.021	-0.207	0.248	-0.015	-0.189	0.158				
	Androstenedione/DHEA			Testosterone/Androstenedione			Cortisol			Cortisone			Cortisone/Cortisol			Adrenal androgen/Glucocorticoid						
	β	95%CI		β	95%CI		β	95%CI		β	95%CI		β	95%CI		β	95%CI					
Oxychlorane	0.119	-0.217	0.455	0.018	-0.140	0.175	0.214	-0.265	0.692	0.374	-0.302	1.050	0.160	-0.145	0.465	-0.361	-1.139	0.416				
cis-Nonachlor	0.142	-0.169	0.453	0.029	-0.117	0.175	0.229	-0.215	0.672	0.463	-0.161	1.087	0.234	-0.046	0.515	-0.463	-1.181	0.255				
trans-Nonachlor	0.201	-0.103	0.505	0.027	-0.116	0.170	0.276	-0.157	0.709	0.517	-0.092	1.126	†	0.241	-0.033	0.516	†	-0.533	-1.234	0.168		
p,p'-DDD	0.091	-0.080	0.262	0.028	-0.052	0.108	0.208	-0.034	0.450	†	0.358	0.017	0.699	*	0.150	-0.004	0.304	†	-0.412	-0.803	-0.021	*
o,p'-DDE	-0.066	-0.266	0.134	0.027	-0.066	0.121	-0.176	-0.460	0.108	-0.030	-0.434	0.374	0.146	-0.034	0.327	0.088	-0.376	0.551				
p,p'-DDE	0.079	-0.179	0.337	0.112	-0.007	0.231	†	0.106	-0.261	0.474	0.285	-0.233	0.803	0.179	-0.054	0.411	-0.333	-0.928	0.261			
o,p'-DDT	-0.084	-0.333	0.164	0.048	-0.068	0.164	-0.269	-0.620	0.083	-0.058	-0.560	0.443	0.210	-0.013	0.433	†	0.143	-0.432	0.718			
p,p'-DDT	0.061	-0.231	0.353	0.116	-0.019	0.251	†	-0.045	-0.461	0.371	0.207	-0.380	0.795	0.253	-0.009	0.514	†	-0.167	-0.842	0.508		
Dieldrin	-0.272	-0.621	0.076	0.112	-0.051	0.275	-0.203	-0.704	0.297	-0.274	-0.982	0.434	-0.071	-0.391	0.249	0.256	-0.557	1.069				
cis-HCE	0.119	-0.217	0.455	0.018	-0.140	0.175	0.214	-0.265	0.692	0.374	-0.302	1.050	0.160	-0.145	0.465	-0.361	-1.139	0.416				
HCB	0.102	-0.359	0.564	-0.025	-0.241	0.191	0.317	-0.339	0.973	0.499	-0.428	1.425	0.182	-0.237	0.600	-0.420	-1.486	0.647				
β -HCH	0.010	-0.290	0.309	0.063	-0.077	0.202	0.004	-0.422	0.431	0.012	-0.592	0.615	0.007	-0.265	0.280	-0.014	-0.707	0.679				
Mirex	0.191	-0.148	0.529	-0.055	-0.214	0.103	0.229	-0.254	0.712	0.510	-0.170	1.189	0.280	-0.025	0.585	†	-0.525	-1.306	0.257			
Parlar-26	-0.051	-0.285	0.183	0.017	-0.092	0.127	0.044	-0.289	0.377	0.091	-0.380	0.562	0.047	-0.165	0.260	-0.040	-0.581	0.501				
Parlar-50	-0.036	-0.300	0.228	0.014	-0.110	0.137	0.028	-0.348	0.404	0.123	-0.408	0.655	0.095	-0.144	0.334	-0.050	-0.661	0.560				
	SHBG			Prolactin																		
	β	95%CI		β	95%CI																	
Oxychlorane	0.060	-0.118	0.238	0.092	-0.149	0.332																
cis-Nonachlor	0.057	-0.108	0.221	0.076	-0.147	0.298																
trans-Nonachlor	0.059	-0.102	0.220	0.085	-0.134	0.303																
p,p'-DDD	0.064	-0.026	0.155	0.120	0.000	0.241	†															
o,p'-DDE	-0.004	-0.109	0.102	-0.048	-0.191	0.094																
p,p'-DDE	0.074	-0.061	0.210	0.092	-0.091	0.275																
o,p'-DDT	-0.002	-0.133	0.130	-0.009	-0.186	0.169																
p,p'-DDT	0.034	-0.120	0.188	0.029	-0.178	0.237																
Dieldrin	0.014	-0.172	0.200	-0.199	-0.453	0.055																
cis-HCE	0.060	-0.118	0.238	0.092	-0.149	0.332																
HCB	0.116	-0.127	0.359	0.128	-0.202	0.458																
β -HCH	0.125	-0.031	0.282	0.009	-0.207	0.225																
Mirex	0.021	-0.159	0.200	0.120	-0.121	0.362																
Parlar-26	0.037	-0.086	0.160	-0.022	-0.188	0.145																
Parlar-50	0.030	-0.110	0.169	-0.002	-0.190	0.186																

Adjusted regression coefficients (β) and 95% confidence intervals (CIs) for the association between a 10-fold increase of OCP and log10-transformed hormone levels as determined by a linear regression model. Each OCP was introduced into the model separately and adjusted for maternal age, parity, and gestational age.

*P < 0.05, **P < 0.01, † P < 0.1

Supplemental Table S1 Concentrations of organochlorine pesticides in maternal blood all measured of original cohort (n=379)

Persistent organochlorine pesticides	Detection limit (pg/g-wet)	>DL (%)	Minimum	Percentile			Maximum
				25th	50th	75th	
Aldrin	1.00	0.3				<DL	12.83
Chlordanes							
oxychlordane	0.90	100.0	7.93	27.05	39.67	56.02	250.94
cis-Chlordane	0.70	58.8		<DL	1.15	2.29	17.53
trans-Chlordane	0.50	45.4			<DL	0.71	3.79
cis-Nonachlor	0.40	100.0	1.63	6.76	9.97	14.36	38.07
trans-Nonachlor	0.50	100.0	13.14	49.70	71.52	107.59	513.52
DDTs							
<i>o,p'</i> -DDD	0.50	12.4				<DL	1.16
<i>p,p'</i> -DDD	0.40	89.7	<DL	0.94	1.48	2.29	9.04
<i>o,p'</i> -DDE	0.40	85.0	<DL	0.75	1.27	1.82	6.20
<i>p,p'</i> -DDE	0.60	100.0	99.52	401.53	650.99	1011.48	4575.67
<i>o,p'</i> -DDT	0.60	7.6	<DL	2.27	3.48	4.86	17.15
<i>p,p'</i> -DDT	0.40	100.0	2.38	16.63	23.16	33.99	121.52
Dieldrin	0.80	100.0	4.11	12.08	16.42	22.62	71.52
Endrin	1.00	0.0					<DL
Heptachlors							
Heptachlor	0.80	0.5				<DL	1.14
trans-HCE	1.00	0.0					<DL
cis-HCE	0.40	100.0	6.17	18.78	26.44	37.28	200.53
HCB	0.90	100.0	34.94	80.24	101.65	130.06	245.48
HCHs							
α -HCH	0.70	69.1		<DL	0.90	1.32	3.89
β -HCH	0.60	100.0	19.95	105.05	154.45	244.76	1667.12
γ -HCH	0.90	59.1		<DL	1.05	1.63	100.92
δ -HCH	0.70	0.8				<DL	1.11
Mirex	0.50	100.0	0.88	4.07	5.95	8.26	34.97
Toxaphenes							
Parlar-26	1.00	97.1	<DL	2.87	4.39	6.65	20.82
Parlar-41	0.70	26.9			<DL	0.72	1.96
Parlar-40	2.00	0.5				<DL	2.43
Parlar-44	2.00	1.6				<DL	2.84
Parlar-50	2.00	96.0	<DL	4.36	6.52	9.68	29.29
Parlar-62	6.00	0.0					<DL

DL, detection limit; DDD, dichlorodiphenyldichloroethane; DDE, dichlorodiphenyldichloroethylene; DDT, dichlorodiphenyltrichloroethane; HCB, hexachlorobenzene; HCE, heptachlor epoxide; HCH, hexachlorocyclohexane.

Supplemental table S2 Maternal and infant characteristics and concentrations of OCPs

characteristics	n	o,p'-DDT	p,p'-DDT	o,p'-DDE	p,p'-DDE	p,p'-DDD	cis-Nonachlor	trans-Nonachlor	oxychlorane	
Mother		ρ	ρ	ρ	ρ	ρ	ρ	ρ	ρ	
Age at delivery (years)		0.009	0.102	0.030	0.270**	0.137*	0.267**	0.311**	0.350**	
Pre-pregnancy BMI (kg/m ²)		0.062	0.118	0.119	0.003	0.052	0.089	-0.011	-0.093	
		Med (min-max)	Med (min-max)	Med (min-max)	Med (min-max)	Med (min-max)	Med (min-max)	Med (min-max)	Med (min-max)	
Educational level	≤12 years	100	3.01 (2.09, 4.47)	21.59 (15.41, 32.84)	1.16 (0.70, 1.71)	565.05 (376.81, 942.71)	1.62 (0.88, 2.44)	9.88 (7.22, 14.64)	72.73 (52.69, 108.21)	39.83 (27.1, 56.02)
	>12 years	132	3.54 (2.42, 4.99)	25.32 (16.37, 34.77)	1.33 (0.96, 1.86)	634.71 (422.04, 976.14)	1.69 (1.03, 2.58)	10.78 (6.85, 15.13)	78.85 (51.80, 118.08)	40.22 (28.87, 57.78)
Annual Household income	<5 million yen per year	166	3.27 (2.22, 4.66)	22.35 (16.38, 33.65)	1.26 (0.71, 1.74)	619.26 (411.26, 942.60)	1.62 (0.99, .53)	9.93 (6.71, 14.35)	70.06 (50.04, 111.09)	39.07 (27.61, 56.02)
	≥5 million yen per year	66	3.64 (2.41, 5.06)	24.18 (15.54, 38.90)	1.23 (0.77, 2.05)	633.52 (392.24, 1302.85)	1.68 (0.98, 2.58)	11.91 (8.21, 15.98)	86.66 (62.34, 110.04)	45.83 (30.02, 66.07)
Smoking during pregnancy	No	190	3.46 (2.37, 4.66)	23.74 (16.38, 33.93)	1.26 (0.73, 1.77)	617.59 (409.71, 971.83)	1.61 (0.98, 2.57)	10.70 (7.17, 15.02)	76.63 (54.06, 112.12)	40.31 (29.25, 57.75)
	Yes	42	3.03 (1.79, 5.03)	21.09 (14.17, 34.22)	1.17 (0.70, 1.84)	625.95 (409.34, 944.10)	1.79 (1.07, 2.39)	8.94 (6.01, 15.10)	69.19 (48.19, 109.46)	38.64 (23.48, 51.97)
Alcohol consumption during pregnancy	No	154	3.10 (2.21, 4.67)	22.62 (15.54, 33.44)	1.22 (0.69, 1.72)	60.5.48 (412.66, 971.83)	1.62 (0.99, 2.55)	10.14 (6.90, 14.76)	74.14 (51.75, 112.12)	40.13 (28.81, 58.13)
	Yes	78	3.86 (2.56, 4.98)	24.14 (17.27, 33.98)	1.27 (0.90, 1.87)	648.50 (406.30, 976.56)	1.67 (0.87, 2.41)	11.04 (7.29, 15.25)	78.85 (52.92, 108.95)	40.04 (28.87, 56.02)
Parity	0	120	3.39 (2.38, 4.67)	24.38 (17.17, 33.92)	1.35 (0.81, 1.83)	652.69 (453.78, 1002.45)	1.64 (1.00, 2.52)	11.52 (7.50, 15.13)	87.74 (56.35, 117.19)	42.99 (32.00, 60.65)
	≥1	112	3.24 (2.22, 4.78)	21.98 (14.92, 34.69)	1.13 (0.69, 1.71)	564.40 (336.51, 938.47)	1.65 (0.91, 2.62)	9.5 (6.37, 14.06)	70.05 (45.60, 97.48)	38.97 (26.35, 54.44)
Blood sampling period	During pregnancy	159	3.42 (2.25, 4.72)	22.49 (15.57, 33.95)	1.22 (0.73, 1.83)	626.50 (412.96, 941.47)	1.51 (0.99, 2.28)	10.01 (7.06, 14.87)	70.38 (21.93, 105.61)	39.19 (27.65, 55.64)
	After delivery	73	3.36 (2.28, 4.48)	23.77 (19.33, 34.17)	1.35 (0.71, 1.73)	614.51 (339.28, 1008.37)	1.76 (0.87, 2.84)	11.15 (6.86, 16.79)	87.82 (51.95, 124.70)	44.66 (29.31, 60.98)
Type of delivery	Vaginal	230	3.36 (2.27, 4.68)	22.84 (16.14, 33.93)	1.25 (0.72, 1.77)	619.26 (410.14, 964.80)	1.62 (0.98, 2.53)	10.33 (7.03, 15.04)	75.08 (51.89, 110.01)	39.87 (28.87, 57.31)
	Caesarian section	2	3.70 (3.03, 4.36)	36.58 (23.77, 49.38)	1.40 (0.91, 1.88)	688.72 (304.51, 1072.92)	2.87 (2.23, 3.5)	18.84 (12.12, 25.55)	119.01 (91.04, 146.97)	55.41 (40.2, 70.62)
Infant		ρ	ρ	ρ	ρ	ρ	ρ	ρ	ρ	
Birth weight		0.022	-0.025	0.005	0.002	-0.087	-0.057	-0.057	-0.061	
Gestational Age		0.066	0.037	0.039	0.117	-0.013	0.025	0.035	0.045	
		Med (min-max)	Med (min-max)	Med (min-max)	Med (min-max)	Med (min-max)	Med (min-max)	Med (min-max)	Med (min-max)	
Sex	Male	106	3.39 (2.42, 4.42)	23.90 (16.56, 33.88)	1.18 (0.72, 1.66)	641.93 (405.32, 934.78)	1.70 (1.04, 2.35)	10.29 (7.11, 15.11)	75.82 (53.44, 118.34)	40.22 (29.26, 57.54)
	Female	126	3.36 (2.20, 4.97)	21.96 (15.78, 34.45)	1.30 (0.72, 1.87)	611.15 (410.00, 998.41)	1.60 (0.93, 2.66)	10.42 (6.80, 15.04)	74.66 (51.75, 107.06)	39.87 (28.34, 57.15)

Supplemental table S1 Maternal and infant characteristics and concentrations of OCPs (continued)

Characteristics	n	Dieldrin	cis-HCE	HCB	β-HCH	Mirex	Parlar-26	Parlar-50	
Mother		ρ	ρ	ρ	ρ	ρ	ρ	ρ	
Age at delivery (years)		0.108	0.254**	0.119*	0.464**	0.513**	0.176**	0.181**	
Pre-pregnancy BMI (kg/m ²)		0.221**	0.177**	0.101	0.147*	-0.074	0.240**	0.240**	
		Med (min-max)	Med (min-max)	Med (min-max)	Med (min-max)	Med (min-max)	Med (min-max)	Med (min-max)	
Educational level	≤12 years	100	16.55 (12.52, 22.84)	26.07 (18.56, 40.28)	101.71 (80.26, 130.96)	154.79 (96.08, 262.30)	5.99 (4.08, 8.34)	4.85 (3.17, 7.55)	7.02 (4.61, 10.28)
	>12 years	132	16.80 (12.11, 21.59)	26.91 (18.96, 37.13)	106.79 (85.26, 167.25)	154.31 (109.21, 220.69)	6.11 (4.31, 8.59)	4.05 (2.56, 6.68)	6.50 (3.88, 9.48)
Annual Household income	<5 million yen per year	166	15.93 (12.02, 20.99)	26.07 (18.44, 35.00)	102.17 (80.36, 129.52)	153.64 (100.87, 209.23)	5.76 (3.88, 7.79)	4.32 (2.69, 6.82)	6.48 (4.25, 9.48)
	≥5 million yen per year	66	17.89 (12.60, 24.41)	26.55 (19.17, 43.07)	111.51 (85.53, 145.40)	165.66 (113.97, 279.53)	7.44 (4.82, 11.64)	5.22 (3.14, 7.61)	7.69 (4.33, 11.64)
Smoking during pregnancy	No	190	16.86 (12.35, 22.47)	26.79 (19.14, 37.90)	103.82 (83.69, 131.84)	158.46 (103.79, 241.30)	6.16 (4.51, 8.61)	4.50 (2.84, 6.60)	6.82 (4.32, 9.74)
	Yes	42	15.51 (11.41, 21.31)	24.39 (16.80, 35.42)	104.48 (74.82, 132.31)	150.87 (103.84, 184.91)	5.69 (3.73, 7.82)	4.23 (2.73, 8.24)	5.88 (3.75, 11.43)
Alcohol consumption during pregnancy	No	154	16.80 (12.14, 22.82)	27.18 (18.91, 39.08)	105.46 (81.20, 133.02)	162.80 (104.40, 244.79)	6.04 (4.11, 8.38)	4.19 (2.82, 6.82)	6.40 (4.28, 9.58)
	Yes	78	16.35 (12.42, 21.66)	24.82 (18.42, 35.69)	99.34 (86.16, 130.72)	146.89 (102.56, 202.87)	6.09 (4.11, 8.61)	4.54 (2.91, 7.58)	7.32 (4.44, 10.56)
Parity	0	120	17.16 (12.52, 21.71)	26.79 (19.16, 36.85)	109.93 (91.03, 131.94)	165.19 (114.52, 271.53)	5.98 (4.14, 8.81)	5.23 (2.96, 7.20)	7.54 (4.54, 9.99)
	≥1	112	16.35 (12.05, 22.66)	25.17 (18.19, 40.63)	95.36 (72.71, 129.15)	143.66 (90.00, 211.57)	6.19 (4.11, 8.47)	4.02 (2.72, 6.68)	5.94 (3.85, 9.66)
Blood sampling period	During pregnancy	159	17.59 (12.39, 22.87)	26.44 (18.91, 38.12)	104.06 (85.45, 130.56)	154.13 (108.08, 218.60)	5.88 (4.07, 7.86)	4.50 (2.93, 7.08)	7.16 (4.25, 9.72)
	After delivery	73	15.51 (12.06, 21.32)	25.10 (18.50, 36.78)	103.73 (75.77, 138.73)	158.62 (94.33, 254.44)	6.76 (4.54, 10.19)	4.33 (2.56, 7.49)	6.42 (4.45, 10.72)
Type of delivery	Vaginal	230	16.61 (12.16, 22.44)	26.15 (18.76, 37.42)	103.99 (82.85, 131.26)	154.31 (104.40, 237.67)	6.04 (4.11, 8.51)	4.41 (2.84, 7.09)	6.55 (4.28, 9.79)
	Caesarian section	2	18.66 (16.82, 20.50)	36.20 (28.26, 44.14)	134.17 (85.55, 182.79)	154.45 (68.50, 240.39)	13.99 (4.84, 23.14)	7.67 (5.34, 10.00)	10.15 (7.63, 12.67)
Infant		ρ	ρ	ρ	ρ	ρ	ρ	ρ	
Birth weight		-0.049	-0.101	-0.066	-0.128	-0.030	-0.024	-0.022	
Gestational Age		-0.035	-0.082	0.052	0.061	0.027	0.024	-0.018	
		Med (min-max)	Med (min-max)	Med (min-max)	Med (min-max)	Med (min-max)	Med (min-max)	Med (min-max)	
Sex	Male	106	15.92 (12.07, 22.55)	26.39 (19.22, 40.80)	101.76 (85.53, 130.85)	153.65 (107.63, 237.67)	6.04 (4.53, 8.94)	4.50 (2.91, 6.86)	6.64 (4.33, 10.27)
	Female	126	17.07 (12.51, 21.96)	26.15 (18.49, 35.42)	104.60 (78.58, 133.01)	155.04 (102.56, 240.78)	6.04 (4.03, 8.41)	4.24 (2.73, 7.36)	6.50 (4.25, 9.74)

Supplemental Table S3 Distribution of steroids and reproductive hormones among all measured of original cohort (n=295)

	DL	n	>DL (%)	Total (n=295)			Boys (n=106)					Girls (n=126)					P-value	
				25%	Med	75%	n	>DL (%)	25%	Med	75%	n	>DL (%)	25%	Med	75%		
Steroid hormones																		
Progesterone (ng/mL)	0.01	295	100	175.5	218.8	278.8	135	100	183.8	225.9	285.6	160	100	167.6	209.0	275.6	0.184	
Testosterone (pg/mL)	0.01	295	100	59.8	83.9	114.1	135	100	76.5	98.9	126.3	160	100	52.1	70.1	96.8	<0.001	
Estradiol (ng/mL)	0.01	295	99.7	3.29	4.70	7.10	135	99.3	3.33	4.86	7.42	160	100	3.16	4.68	6.61	0.227	
DHEA (ng/mL)	0.01	295	100	1.77	2.19	2.99	135	100	1.59	2.08	2.76	160	100	1.91	2.34	3.22		
Androstenedione (ng/mL)	0.01	295	99.3	0.36	0.46	0.58	135	98.5	0.38	0.47	0.61	160	99.4	0.35	0.45	0.58		
Cortisol (ng/mL)	0.25	295	97.6	22.7	38.9	63.6	135	98.5	22.5	38.3	65.3	160	96.9	22.8	39.2	62.8		
Cortisone (ng/mL)	0.10	295	94.9	69.9	93.9	123.0	135	97.0	70.5	95.3	123.4	160	93.1	69.7	93.0	123.0		
Steroid Hormone Binding Globulin (nmol/L)	1.10	295	100	13.3	15.8	19.0	135	100	13.9	16.5	19.3	160	100	13.0	15.5	18.7	0.079	
Lutealizing Hormone (mIU/mL)	0.50	288	16.3			<DL	132	34.8			<DL	0.84	156	0.6			<DL	<0.001
Follicle Stimulating Hormone (mIU/mL)	0.50	287	20.9			<DL	132	45.5			<DL	0.66	155	0			<DL	<0.001
Inhibin B (pg/mL)	11	295	59.7	5.5	23.2	44.6	135	98.5	33.9	44.0	58.3	160	26.9			<DL	12.4	<0.001
Insulin-like factor 3 (ng/mL)	0.01	157	100	0.23	0.27	0.32	132	100	0.25	0.29	0.34	25	100	0.17	0.18	0.23	<0.001	
Prolactin (ng/mL)	1.0	289	99.7	63.1	85.8	116.0	132	100	65.4	85.2	115.0	157	99.6	61.4	86.0	118.0	0.986	

P values were calculated by Mann-Whitney U test;

DHEA, dehydroepiandrosterone; DL, detection limit

Supplemental Table S4 Associations between OCPs exposure and steroid and reproductive hormone levels and sex interaction

	Progesterone				Estradiol				Testosterone				Estradiol/Testosterone				DEHA				Androstenedione			
	β	95%CI	$p^{a)}$	$p^{b)}$	β	95%CI	$p^{a)}$	$p^{b)}$	β	95%CI	$p^{a)}$	$p^{b)}$	β	95%CI	$p^{a)}$	$p^{b)}$	β	95%CI	$p^{a)}$	$p^{b)}$	β	95%CI	$p^{a)}$	$p^{b)}$
Oxychlorane	0.024	-0.189	0.238		0.025	-0.178	0.229		-0.065	-0.249	0.118		0.091	-0.099	0.281		0.067	-0.119	0.253		0.019	-0.137	0.175	
cis-Nonachlor	0.022	-0.166	0.210	*	0.010	-0.170	0.190		-0.030	-0.192	0.132		0.040	-0.128	0.208		0.055	-0.108	0.218		0.039	-0.098	0.176	
trans-Nonachlor	0.014	-0.172	0.200	*	0.005	-0.173	0.184		-0.049	-0.209	0.112		0.054	-0.112	0.221		0.046	-0.116	0.208		0.038	-0.098	0.175	
p,p'-DDD	0.004	-0.111	0.119	†	-0.102	-0.211	0.008	†	-0.054	-0.153	0.045		-0.047	-0.150	0.055		-0.091	-0.191	0.009	†	-0.043	-0.126	0.041	
o,p'-DDE	-0.044	-0.169	0.082		-0.017	-0.136	0.102		-0.034	-0.142	0.074		0.017	-0.094	0.128		0.039	-0.070	0.149		-0.027	-0.118	0.064	
p,p'-DDE	0.017	-0.149	0.182		0.027	-0.130	0.183		-0.031	-0.172	0.110		0.058	-0.087	0.202	*	-0.029	-0.172	0.114		-0.060	-0.179	0.060	
o,p'-DDT	-0.069	-0.210	0.073		-0.035	-0.169	0.099		-0.024	-0.146	0.097		-0.011	-0.136	0.114		0.067	-0.056	0.190		-0.031	-0.134	0.071	
p,p'-DDT	-0.020	-0.198	0.158		-0.016	-0.186	0.154		0.007	-0.146	0.160		-0.023	-0.181	0.136		0.024	-0.130	0.179		-0.012	-0.141	0.118	
Dieldrin	-0.093	-0.304	0.118		0.022	-0.179	0.224		-0.059	-0.240	0.122		0.081	-0.106	0.269		0.164	-0.019	0.347	†	-0.022	-0.174	0.130	*
cis-HCE	-0.040	-0.238	0.158	†	-0.009	-0.198	0.181		-0.053	-0.224	0.118		0.044	-0.133	0.221		0.098	-0.074	0.271		-0.028	-0.173	0.117	
HCB	0.122	-0.178	0.422		0.033	-0.253	0.319		-0.060	-0.317	0.198		0.093	-0.173	0.359		0.126	-0.135	0.387		0.053	-0.165	0.271	
β -HCH	0.035	-0.151	0.220		0.050	-0.126	0.227		-0.014	-0.173	0.145		0.065	-0.099	0.229		0.056	-0.106	0.217		-0.004	-0.139	0.130	
Mirex	0.061	-0.146	0.268	†	0.027	-0.171	0.226		-0.154	-0.332	0.024	†	0.182	-0.003	0.366	†	0.018	-0.163	0.200		-0.019	-0.171	0.133	
Parlar-26	-0.046	-0.190	0.099		0.061	-0.077	0.198		-0.004	-0.128	0.121		0.065	-0.064	0.193		0.127	0.003	0.252	*	0.036	-0.069	0.140	
Parlar-50	-0.027	-0.181	0.127		0.032	-0.114	0.179		-0.010	-0.142	0.123		0.042	-0.095	0.179		0.127	-0.006	0.260	†	0.044	-0.067	0.156	
	Androstenedione/DHEA				Testosterone/Androstenedione				Cortisol				Cortisone				Cortisone/Cortisol				Adrenal androgen/ Glucocorticoid			
Oxychlorane	-0.051	-0.263	0.161		-0.083	-0.195	0.029		-0.084	-0.402	0.234		-0.014	-0.436	0.408	†	0.071	-0.138	0.281		0.100	-0.386	0.585	†
cis-Nonachlor	-0.018	-0.205	0.168		-0.067	-0.166	0.032	†	-0.009	-0.290	0.272		0.066	-0.306	0.438	*	0.076	-0.108	0.260	†	0.022	-0.405	0.450	*
trans-Nonachlor	-0.009	-0.194	0.175	*	-0.086	-0.184	0.011	†	-0.057	-0.334	0.220	*	0.040	-0.328	0.407	*	0.097	-0.085	0.280	†	0.044	-0.378	0.466	*
p,p'-DDD	0.050	-0.065	0.164		-0.013	-0.074	0.048		0.133	-0.038	0.305		0.186	-0.040	0.413		0.053	-0.059	0.165	†	-0.230	-0.490	0.031	†
o,p'-DDE	-0.066	-0.190	0.058		-0.008	-0.074	0.058		-0.139	-0.325	0.048		-0.088	-0.336	0.161		0.052	-0.071	0.174		0.142	-0.143	0.428	
p,p'-DDE	-0.032	-0.195	0.131		0.029	-0.057	0.115	†	-0.092	-0.337	0.153		0.017	-0.308	0.343		0.107	-0.054	0.269		0.006	-0.368	0.380	†
o,p'-DDT	-0.097	-0.237	0.042		0.006	-0.068	0.081		-0.215	-0.424	-0.006	*	-0.104	-0.384	0.176		0.110	-0.028	0.248		0.199	-0.123	0.520	
p,p'-DDT	-0.037	-0.213	0.140		0.019	-0.075	0.112	*	-0.064	-0.331	0.202		0.003	-0.350	0.356		0.067	-0.106	0.240	*	0.058	-0.348	0.465	
Dieldrin	-0.189	-0.397	0.019	†	-0.035	-0.145	0.075	*	-0.094	-0.409	0.221		-0.243	-0.661	0.176		-0.149	-0.356	0.057		0.312	-0.168	0.793	
cis-HCE	-0.189	-0.326	0.068		-0.024	-0.127	0.080	*	0.001	-0.297	0.298		-0.092	-0.486	0.302		-0.094	-0.289	0.101		0.139	-0.314	0.592	
HCB	-0.076	-0.373	0.221		-0.112	-0.269	0.046		0.014	-0.433	0.461		0.146	-0.448	0.740		0.132	-0.162	0.426		0.036	-0.647	0.719	
β -HCH	-0.063	-0.246	0.121		-0.009	-0.106	0.089		-0.070	-0.347	0.206		-0.035	-0.402	0.333		0.035	-0.147	0.217		0.101	-0.321	0.524	
Mirex	-0.037	-0.243	0.169		-0.134	-0.243	-0.026	*	-0.207	-0.515	0.100	*	-0.047	-0.456	0.362	*	0.161	-0.042	0.364		0.122	-0.349	0.593	*
Parlar-26	-0.092	-0.235	0.051		-0.039	-0.115	0.037		-0.070	-0.286	0.145		-0.100	-0.386	0.186		-0.031	-0.172	0.111		0.209	-0.119	0.536	
Parlar-50	-0.084	-0.237	0.068		-0.053	-0.134	0.028		-0.070	-0.300	0.159		-0.058	-0.362	0.247		0.012	-0.139	0.163		0.186	-0.163	0.536	
	SHBG				Prolactin																			
Oxychlorane	-0.014	-0.136	0.108		0.011	-0.144	0.167																	
cis-Nonachlor	0.004	-0.104	0.112		-0.041	-0.177	0.096																	
trans-Nonachlor	0.012	-0.095	0.119		0.003	-0.133	0.139																	
p,p'-DDD	0.058	-0.008	0.123		0.070	-0.013	0.153	†																
o,p'-DDE	-0.016	-0.087	0.056		-0.082	-0.172	0.009	†																
p,p'-DDE	0.001	-0.092	0.095		-0.060	-0.178	0.058	*																
o,p'-DDT	-0.006	-0.087	0.075		-0.090	-0.191	0.011	†																
p,p'-DDT	0.007	-0.095	0.109		-0.063	-0.191	0.065																	
Dieldrin	-0.020	-0.140	0.101		-0.201	-0.354	-0.048	*																
cis-HCE	0.040	-0.074	0.153		-0.070	-0.215	0.075																	
HCB	-0.029	-0.200	0.141		-0.057	-0.273	0.159																	
β -HCH	-0.002	-0.107	0.103		-0.078	-0.212	0.056																	
Mirex	-0.041	-0.160	0.078		-0.029	-0.180	0.121																	
Parlar-26	0.008	-0.075	0.090		-0.066	-0.171	0.040																	
Parlar-50	-0.011	-0.099	0.077		-0.069	-0.181	0.043																	

β s were changes of hormone levels (log 10 transformed) of 10 fold increase of each OCP, calculated by linear regression model.

^{a)} P for OCP adjusted for Maternal age, Parity, Gestational age, and (sex) x (each OCP); ^{b)} P for interaction of sex and OCP.

*P < 0.05, **P < 0.01, †P < 0.1