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# **Research on Microplastics in Marine Life in the Coastal Areas of Japan**

(日本近海におけるマリンライフ中の  
マイクロプラスチックの研究)

Hokkaido University Graduate School of Global Food Resources Division  
of Global Food Resources Doctoral Course

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# Chapter 1: Introduction

## 1.1 Marine pollution

Oceans cover over 70% of the earth's surface. Millions of creatures exist, from microscopic plankton to giant whales [1]. Oceans also significantly impact the global climate system by regulating temperature, weather patterns, and the carbon cycle [2].

Marine pollution is a major global environmental concern because of the significant risks. It can harm the health and well-being of humans and marine ecosystems [3]. This affects all levels, from low-level microbes to apex predators, through various pathways that harm the entire marine ecosystem. Impacts on sediments, eutrophication, toxic substances, and marine debris are reported as marine pollution. Various pollution causes include intensive agriculture, urbanization, industrialization, deforestation, and coastal developments [4]. It occurs when harmful substances or materials are intentionally or unintentionally introduced into the ocean [5]. Chemical pollutants such as oil, plastic, and heavy metals, and biological contaminants such as pathogens and invasive species can be listed among these harmful substances [6]. Marine pollution harms marine ecosystems in several ways, including habitat destruction, species extinction, and ecosystem degradation [7]. These pollutants include chemical compounds, petroleum-derived substances, marine debris, sewage-associated pathogens, and excess nutrients with high environmental impacts [8].

Plastics, including microplastics and larger plastic debris (macro-plastics), are a big issue. Plastic is a durable and versatile material used in various products, such as packaging, toys, and construction materials [9], with much of it ending in the ocean [10]. Macro-plastics can lead to

entanglement and suffocation in birds, fish, and sea turtles [11]. Moore [12] found that approximately 60% of sea turtles have consumed plastic.

Moreover, microplastics can be hazardous to marine animals and human health and transport harmful chemicals and pathogens [13] into the food chain and end up in the seafood we eat [14].

Sustainable Development Goal 14 aims to “Conserve and sustainably use the oceans, seas and marine resources for sustainable development.” In order to achieve this goal, efforts from all sectors of society are required, including increased research on marine life [15]. Marine mammals are at the top of the marine food chain and can accumulate some of the highest levels of environmental pollutants of all wildlife [16]. Given the increasing prevalence and severity of marine wildlife diseases, it is important to understand the types and mechanisms of the toxicity of pollutants. For example, persistent pollutants have been associated with various toxic effects in marine mammals, including immunosuppression and the development of infectious diseases [17]. Considering the background knowledge relevant to this study, this chapter focuses on the impacts of microplastics in the marine ecosystem and summarises the research on microplastics and their importance for the main objective of this thesis.

## 1.2 Plastic pollution

Plastic pollution is a major environmental concern that has garnered significant attention in recent years. Plastic pollution is defined as the presence of plastic materials in the environment with adverse effects on wildlife, habitats, and human populations [18]. Marine litter (marine waste, garbage, or debris) is every type of man-made solid waste that deliberately or accidentally pollutes rivers, lakes, seas and coastlines [19]. Marine debris can be categorized into several diverse classes of material, including plastics, metal, glass, paper and so on [20]. Furthermore, plastic debris is a common component of garbage, and when not properly disposed of, it can end up in the environment, thus contributing to overall ocean pollution [21].

Degradation of plastics is difficult due to their xenobiotic origin and recalcitrant nature [22]. Global plastic production increased substantially to 311 million tonnes in 2014 and rose to 368 million tonnes in 2019 [23]. Plastics are generally inexpensive to manufacture and chemically resistant, which makes them difficult to degrade. They are used in various industries and products, including packaging, the food industry, medicine, electronics, textiles, and construction. Between 1950 and 2015, 6.3 billion tons of primary and secondary (recycled) plastic waste were generated. Approximately 9% were recycled, 12% incinerated, and the remaining 79% was either stored in landfills or released directly into the environment [24]. Approximately 40% of plastic packaging ends up in landfills, while 32% leaks from sewage systems [25].

A summary of the main categories of negative effects of plastics in the oceans allows for a basic understanding of the present problems facing marine life [26]:

- 1) Physical impacts: entanglement, suffocation, drowning, different injuries, and death through ingestion.

2) Chemical impacts: Plastic additives may release toxic products into the environment that can accumulate in the food chain.

3) Ecological impacts: ecosystem and habitat structure can be affected, and wildlife reproductive behaviour can be disrupted.

4) Economic impacts: industries, such as fisheries and tourism, can be heavily impacted through habitat destruction and loss of biodiversity.

Increased plastic pollution affects all levels of marine diversity [27]. They are widely distributed in northeast Pacific waters, including the “Great Pacific Garbage Patch,” an area of debris captured by the North Pacific central gyre and well known for its high concentration of plastic debris [28]. Lost or abandoned fishing gear is also major plastic pollution in the oceans [29]. It accumulates in the environment and is broken down into smaller pieces through weathering. Nevertheless, efforts are being made to reduce the amount of plastic entering the marine environment [30] and thus reduce the danger they pose to marine life.



## 1.3 Microplastic pollution

Microplastics are small plastic particles less than 5 millimeters in size [31] that can be found in the form of fibres, beads, or shards, and are often used in products such as cosmetics, clothing, and packaging materials [32]. They can negatively impact a wide range of organisms, including fish, birds, and other wildlife, potentially entering the human food chain through the consumption of seafood and other sources, thus negatively that impacting human health [33]. The occurrence of microplastics is derived from the degradation of large plastic debris [34]. As of late, microplastics are a problem in oceans world-wide [35].

The presence and accumulation of microplastics in the ocean are of great concern partly because marine biota consumes them [36]. The sharp rise in the number of publications on microplastics in aquatic ecosystems observed lately justifies the great concern. Most research focused on quantifying microplastic abundance to understand these impacts better. Additionally, some researchers count the combinations of microplastic fragments that vary in size, shape, colour, specific gravity, chemical composition, and other characteristics.

### 1.3.1 The history of microplastic research

Microplastic research can be traced back to the 1960s when scientists first began to study the presence of small plastic particles in the ocean [37,38]. Research has revealed that microplastics have been found in various environments, including land, oceans, rivers, and even in the air. One of the major challenges in studying microplastics is the difficulty in detecting and measuring them [39-41]. Various methods have been developed to identify and quantify microplastics, including microscopy, spectroscopy, and chemical analysis [42-44].

Microplastic may threaten marine life, such as shellfish, fish, and cetaceans. They can be ingested directly from the surrounding water or by consuming prey that has ingested microplastics, thus considerably impacting the entire food chain [45,46]. Only a few studies have looked for microplastics directly in their digestive tracts [47]. The Indo-Pacific humpback dolphin (*Sousa Chinensis*) was the first cetacean species found with microplastics, showing that this pollution may threaten coastal ecosystems [48].

Many shellfish, such as Bivalvia, are very susceptible to microplastic contamination [49]. Research on blue mussels showed that they could ingest microplastics transported from the stomach to the digestive glands [50]. It is also reported that microplastics in the gills of blue mussels [50]. Table 1.1 summarises research confirming the presence of microplastics in shellfish. Most studies consider ingestion as the main route of microplastics into aquatic organisms showing that the digestive system has the major accumulation of these particles. Therefore, studying the microplastic distribution in the digestive system of shellfish is crucial because it can reflect its abundance in the environment [51].

**Table 1.1** Studies dealing with microplastic contamination in shellfish worldwide [52].

<b>Species</b>	<b>Countries</b>	<b>The lower limit of the filter pore size (<math>\mu\text{m}</math>)</b>
<i>Mytilus edulis</i>	France	1.6
<i>M. edulis</i> and <i>Perna viridis</i>	China	1.6
<i>M. edulis</i>	Korea	20
<i>P. viridis</i>	India	11
<i>Siliqua patula</i>	United States	0.2
<i>Theodoxus fluviatilis</i>	Germany	0.2

One of the key areas of microplastic research has been the impact of these particles on human health. Some research reported that microplastics could enter the food chain from the ingestion by marine life, potentially negatively affecting other organisms' health [53,54]. The long-term effects of microplastic exposure in humans are not yet fully understood, but research is increasing in this area due to the increased public concern about the prevalence of microplastics in the environment. Governments and organizations around the world are starting to take action to address this issue, including banning or regulating their use in consumer products [55,56]. Overall, the research on microplastics has highlighted the need for a better understanding of their sources, impacts and effective reduction strategies.

### 1.3.2 Microplastics in oysters

Oysters are widely distributed in estuaries due to their high tolerance to salinity variation and pollution in general [57]. They are filter-feeding mollusks that settle gregariously and are directly exposed to contaminants in the water [58]. One oyster can filter 5-25 L of seawater per hour [59], and during this process, they are exposed to a high possibility of ingesting microplastics [60]. In brackish estuaries, most oysters are attached to hard surfaces such as rocks, mangrove roots, and other oyster shells [61]. They can also be found along the intertidal zone in many relatively shallow water, calm environments [62].

Oysters are famous worldwide as food. They also concentrate filtered particulate matter in excrement as bio deposits [63]. Oysters could capture and transport all sizes of microspheres and microfibrils [64].

It is important to understand the presence and accumulation of microplastics in oysters and their possible implications for human health because they are consumed globally. Unfortunately, we still lack a strong enough data set related to the presence and consequences of microplastics in these organisms. The Pacific oysters (*Magallana gigas*) from the Southwestern Atlantic were reported to contain microplastics [65]. Moreover, oysters are suitable for studying microplastics and bioaccumulation because oysters are easy to culture in the laboratory [66]. Their widespread distribution, filter-feeding behaviour, and commercial importance, especially for aquaculture, make oysters ideal for studying microplastics in marine environments. For this thesis, quantities, and types of microplastics in oysters were recorded. They were assumed to indicate the level of potential microplastic distribution in the marine environment.

### 1.3.3 Microplastics in cetaceans

Cetaceans occupy the top trophic level in the marine ecosystem [67]. Cetaceans are catalyzed by Mysticeti and Odontoceti [68]. Mysticeti feeds by filtering large volumes of water through their baleen [69]. Odontoceti includes various species, such as all species of dolphins, all species of porpoises, and some species of whales like sperm whale *Physeter macrocephalus* and beaked whales [70]. One of the defining characteristics of odontocetes is their teeth, which they use to catch and consume their prey. Odontocetes are active hunters that feed on fish, squid, and other marine animals [71]. As apex predators, cetaceans are critical in regulating the food web and maintaining ecosystem balance.

Mysticeti has a filtering system in their intestines. This system is particularly important for species that feed on small prey, such as krill and plankton, which they consume in large quantities [72]. Intestinal villi trap and filter out unwanted particles or debris from their food [73]. Chemical functional groups such as carboxyl, amine, and hydroxyl in the intestine walls interact with and bind to various substances, including microplastics, thus removing them from the body [74]. Consequently, because the intestines are a key site for bioaccumulation, it is essential for microplastic research in cetaceans.

Bioaccumulation is the process by which a substance, such as a toxin or a pollutant, builds up in the tissues of an organism over time [75]. This occurs when an organism ingests a substance at a rate faster than it can metabolize or eliminate it, leading to the accumulation of the substance in the body [76]. One example of bioaccumulation is the case of polychlorinated biphenyls (PCBs), a group of synthetic chemicals commonly used in industrial applications [77]. The harmful effects of PCB bioaccumulation are the case of killer whales. Studies have shown that these whales have some of the highest levels of PCBs among marine mammals, and the chemicals have been linked

to reproductive and immune system problems in the animals [78]. In addition, harmful substances, including heavy metals, can accumulate in animal tissues through bioaccumulation [79]. With the more harmful substances in the environment, including microplastics, animals such as cetaceans may accumulate more in their bodies. Cetaceans are more sensitive to microplastics.

At present, studies are not enough on microplastics in cetaceans. A study revealed the presence of an average of 5.5 particles per animal among the observed 50 stranded mammals [80]. A review [81] found that microplastics can interfere with the feeding and reproductive success of various marine organisms, including fish and invertebrates, which are food for cetaceans. Microplastics were reported in the gastrointestinal tract of bottlenose dolphins (*Tursiops truncatus*) stranded in South Carolina in a study covering materials from 2017 to 2018 [82] and were also found in the coastal species Indo-Pacific humpback dolphins (*Sousa chinensis*) [48]. As cetaceans rely on diverse food sources, microplastics in their prey may have cascading effects on cetacean populations. Therefore, further research on the prevalence and effects of microplastics in these animals is needed to help conservation efforts and shed light on the larger issue of plastic pollution in the oceans.

## 1.4 Impact of microplastics on seafood safety

A safe food supply is essential for a healthy society. The existing food systems are fraught with different types of risks. Food safety has been thought of narrowly, such as foodborne illnesses and risks directly related to food intake. A more comprehensive definition of food safety has been provided recently. Food safety refers to the measures taken to ensure that food is safe and free from harmful contaminants or substances that could cause illness or injury to consumers [83,84]. Zoonotic bacteria, viruses, parasites, and various physical hazards in food may pose safety risks to human health [85]. Among widely used food sources, seafood products are recognized globally for their high nutritional value and increasing popularity with consumers. Consumer preferences range from raw or minimally processed fresh products to prepared (salted, smoked, cured, canned) and ready-to-eat products. In addition, seafood is a major food category in international trade and is a major economic asset [86]. The importance of seafood safety will be discussed in the following section.

Seafood can contain many biological, chemical, and physical hazards. The most prevalent things are biogenic amines, biotoxins, and disease-causing bacteria and viruses [87]. More than 80% of seafood-borne outbreaks are associated with biotoxins (ciguatoxins), mackerel toxins, or the consumption of raw mollusks [88]. Some of the largest food poisoning outbreaks have been associated with seafood. Data on foodborne disease outbreaks collected over 10 years (1983–1993) in the United States demonstrated that fish were the third most reported category according to the vehicle of transmission [89].

On the other hand, there is concern about the possible impacts of microplastics on human health. Microplastic particles are ubiquitous in the environment, from the air we breathe to the food consumed by humans. Growing evidence suggests that humans are exposed to microplastics

through food and drink consumption [90]. Chemical toxicity may be caused by organic and inorganic components of microplastics or by the exogenous harmful chemicals, proteins, and toxins they may carry [91-93]. A study mentioned in the question measured the exposure of European shellfish consumers to microplastics through dietary intake [94]. Therefore, studies of the potential threat of microplastics represented through the food web are needed to evaluate the possible food safety issues [95].

In conclusion, seafood is an important source of nutrition worldwide. It is essential to understand the potential risks associated with consuming seafood containing microplastics to protect public health. It is also important for consumers to be aware of these potential risks. Research on seafood safety and microplastics can help inform policies in order to develop regulations and strategies to reduce the number of microplastics entering the oceans, thus protecting seafood sources. Furthermore, cetaceans consume lower trophic-level organisms that have ingested microplastics. Therefore, studying the levels of microplastics in whales can provide an indication of the level of microplastic contamination in the larger marine food web, including seafood that humans consume. Consequently, research on seafood safety and microplastics are critical for protecting public health, the health of the oceans, and the sustainability of seafood as a resource.



## **1.5 Thesis objectives and outline**

### **1.5.1 Main Objectives**

The main purpose of this thesis was to study the presence and accumulation of microplastics in oysters and cetaceans, for better understanding their potential impacts on these species and the marine environment. In the case of oysters, the possible relationship between microplastics and food safety was discussed. The research involved analyzing the presence of fibres and their quantification in oysters and cetaceans from different locations and comparing the results to determine trends or patterns. The results of this research will contribute to the growing body of knowledge on the impacts of microplastics on marine life and the ecosystem. Obtained results can be used to inform efforts directed at the mitigation and management of this emerging environmental issue. Finally, the potential effects of microplastics on human health are summarised based on literature data analyzing the relations between plastic and toxicity. The conclusion discusses possible ways to reduce the impacts of microplastic pollution from the point of view of political measures and social awareness.

## 1.5.2 Thesis Outline

This thesis comprises 5 chapters.

Chapter 1 describes the current situation of marine pollution, especially microplastic.

Chapter 2 introduces a newly developed portable observation system used in this research.

Chapter 3 presents results from the analyses to detect and quantify microplastics in cultured oysters from the coastal areas of Japan.

Chapter 4 discusses the presence of microplastics in the intestines of stranded cetaceans.

Chapter 5 summarises the toxicity of microplastic from the viewpoint of quantitative analysis using published data.

## 1.6 References

- [1] Penesyan, A., Kjelleberg, S., & Egan, S. (2010). Development of novel drugs from marine surface associated microorganisms. *Marine Drugs*, 8(3), 438-459.
- [2] Smith Jr, K. L., Ruhl, H. A., Bett, B. J., Billett, D. S. M., Lampitt, R. S., & Kaufmann, R. S. (2009). Climate, carbon cycling, and deep-ocean ecosystems. *Proceedings of the National Academy of Sciences*, 106(46), 19211-19218.
- [3] Cole, M., Lindeque, P., Halsband, C., & Galloway, T. S. (2011). Microplastics as contaminants in the marine environment: a review. *Marine pollution bulletin*, 62(12), 2588-2597.
- [4] Todd, P. A., Ong, X., & Chou, L. M. (2010). Impacts of pollution on marine life in Southeast Asia. *Biodiversity and Conservation*, 19(4), 1063-1082.
- [5] Tornero, V., & Hanke, G. (2016). Chemical contaminants entering the marine environment from sea-based sources: a review with a focus on European seas. *Marine Pollution Bulletin*, 112(1-2), 17-38.
- [6] Sohrabi, H., Hemmati, A., Majidi, M. R., Eyvazi, S., Jahanban-Esfahlan, A., Baradaran, B., Adlpour-Azar R., Mokhtarzadeh A., & de la Guardia, M. (2021). Recent advances on portable sensing and biosensing assays applied for detection of main chemical and biological pollutant agents in water samples: A critical review. *TrAC Trends in Analytical Chemistry*, 143, 116344.
- [7] Shahidul Islam, M. S., & Tanaka, M. (2004). Impacts of pollution on coastal and marine ecosystems including coastal and marine fisheries and approach for management: a review and synthesis. *Marine Pollution Bulletin*, 48(7-8), 624-649.
- [8] Desforges, J. P. W., Sonne, C., Levin, M., Siebert, U., De Guise, S., & Dietz, R. (2016). Immunotoxic effects of environmental pollutants in marine mammals. *Environment International*, 86, 126-139.

- [9] Heller, M. C., Mazor, M. H., & Keoleian, G. A. (2020). Plastics in the US: toward a material flow characterization of production, markets and end of life. *Environmental Research Letters*, *15*(9), 094034.
- [10] Gewert, B., Plassmann, M. M., & MacLeod, M. (2015). Pathways for degradation of plastic polymers floating in the marine environment. *Environmental Science. Processes and Impacts*, *17*(9), 1513-1521.
- [11] Laist, D. W. (1997). Impacts of marine debris: entanglement of marine life in marine debris including a comprehensive list of species with entanglement and ingestion records. *In Marine debris* (pp. 99-139). Springer, New York, NY.
- [12] Moore, C. J. (2008). Synthetic polymers in the marine environment: a rapidly increasing, long-term threat. *Environmental Research*, *108*(2), 131-139.
- [13] Campanale, C., Massarelli, C., Savino, I., Locaputo, V., & Uricchio, V. F. (2020). A detailed review study on potential effects of microplastics and additives of concern on human health. *International journal of environmental research and public health*, *17*(4), 1212.
- [14] Seltenrich, N. (2015). New link in the food chain? Marine plastic pollution and seafood safety. *Environmental Health Perspectives*, *123*(2), A34-A41.
- [15] Duarte, C. M., Agusti, S., Barbier, E., Britten, G. L., Castilla, J. C., Gattuso, J. P., Fulweiler, R. W., Hughes, T. P., Knowlton, N., Lovelock, C. E., Lotze, H. K., Predragovic, M., Poloczanska, E., Roberts, C., & Worm, B. (2020). Rebuilding marine life. *Nature*, *580*(7801), 39-51.
- [16] Aguilar, A., Borrell, A., & Reijnders, P. J. H. (2002). Geographical and temporal variation in levels of organochlorine contaminants in marine mammals. *Marine Environmental Research*, *53*(5), 425-452.

- [17] Whittaker, R. J., Araújo, M. B., Jepson, P., Ladle, R. J., Watson, J. E. M., & Willis, K. J. (2005). Conservation biogeography: assessment and prospect. *Diversity and Distributions*, 11(1), 3-23.
- [18] Sheavly, S. B., & Register, K. M. (2007). Marine debris & plastics: environmental concerns, sources, impacts and solutions. *Journal of Polymers and the Environment*, 15(4), 301-305.
- [19] Valavanidis, A., & Vlachogianni, T. (2012). Marine litter: man-made solid waste pollution in the Mediterranean Sea and coastline. Abundance, composition and sources identification. *Science Advances on Environmental Chemistry, Toxicology and Ecotoxicology*, 1, 1-18.
- [20] Attamimi, A., Purba, N. P., Anggraini, S. R., Harahap, S. A., & Husrin, S. (2015). Investigation of marine debris in Kuta Beach, Bali. *Proceedings of Environmental Engineering and Water Technology, Integrated Water System and Governance (Malang, East Java, Indonesia)*, C1-7.
- [21] Jambeck, J. R., Geyer, R., Wilcox, C., Siegler, T. R., Perryman, M., Andrady, A., Narayan, R., & Law, K. L. (2015). Plastic waste inputs from land into the ocean. *Science*, 347(6223), 768-771.
- [22] Padmanabhan, L., Varghese, S., Patil, R. K., Rajath, H. M., Krishnasree, R. K., & Shareef, M. I. (2020). Ecofriendly degradation of polyethylene plastics using oil degrading microbes. *Recent Innovations in Chemical Engineering (Formerly Recent Patents on Chemical Engineering)*, 13(1), 29-40.
- [23] Abhijith, R., Jayasankar, R., Shilta, M. T., Asokan, P. K., & Vinod, K. (2022). Save the Sea say 'NO' to plastics.
- [24] Geyer, R., Jambeck, J. R., & Law, K. L. (2017). Production, use, and fate of all plastics ever made. *Science Advances*, 3(7), e1700782.

- [25] Rhodes, C. J. (2018). Plastic pollution and potential solutions. *Science progress*, 101(3), 207-260.
- [26] LI, W. C., Tse, H. F., & Fok, L. (2016). Plastic waste in the marine environment: A review of sources, occurrence and effects. *Science of the total environment*, 566, 333-349.
- [27] Fowler, C. W. (1987). Marine debris and northern fur seals: a case study. *Marine Pollution Bulletin*, 18(6), 326-335.
- [28] Moore, C. J., Moore, S. L., Leecaster, M. K., & Weisberg, S. B. (2001). A comparison of plastic and plankton in the North Pacific central gyre. *Marine Pollution Bulletin*, 42(12), 1297-1300.
- [29] Stelfox, M., Hudgins, J., & Sweet, M. (2016). A review of ghost gear entanglement amongst marine mammals, reptiles and elasmobranchs. *Marine Pollution Bulletin*, 111(1-2), 6-17.
- [30] Villarrubia-Gómez, P., Cornell, S. E., & Fabres, J. (2018). Marine plastic pollution as a planetary boundary threat—The drifting piece in the sustainability puzzle. *Marine policy*, 96, 213-220.
- [31] Jung, S., Cho, S. H., Kim, K. H., & Kwon, E. E. (2021). Progress in quantitative analysis of microplastics in the environment: a review. *Chemical Engineering Journal*, 422, 130154.
- [32] Hammer, J., Kraak, M. H., & Parsons, J. R. (2012). Plastics in the marine environment: the dark side of a modern gift. *Reviews of environmental contamination and toxicology*, 1-44.
- [33] Waring, R. H., Harris, R. M., & Mitchell, S. C. (2018). Plastic contamination of the food chain: A threat to human health? *Maturitas*, 115, 64-68.
- [34] Egessa, R., Nankabirwa, A., Ocaya, H., & Pabire, W. G. (2020). Microplastic pollution in surface water of Lake Victoria. *Science of the Total Environment*, 741, 140201.

- [35] Acosta-Coley, I., & Olivero-Verbel, J. (2015). Microplastic resin pellets on an urban tropical beach in Colombia. *Environmental Monitoring and Assessment*, 187, 1-14.
- [36] Moore, C. J. (2008). Synthetic polymers in the marine environment: a rapidly increasing, long-term threat. *Environmental Research*, 108(2), 131-139.
- [37] Shabbir, S., Faheem, M., Ali, N., Kerr, P. G., Wang, L. F., Kuppusamy, S., & Li, Y. (2020). Periphytic biofilm: An innovative approach for biodegradation of microplastics. *Science of the Total Environment*, 717, 137064.
- [38] Wallen, I. E. (1968). Materials problems in the utilization of marine biology resources. *Ocean Engineering*, 1(2), 149-157.
- [39] Zhang, Y., Kang, S., Allen, S., Allen, D., Gao, T., & Sillanpää, M. (2020). Atmospheric microplastics: a review on current status and perspectives. *Earth-Science Reviews*, 203, 103118.
- [40] Prata, J. C., da Costa, J. P., Duarte, A. C., & Rocha-Santos, T. (2019). Methods for sampling and detection of microplastics in water and sediment: a critical review. *TrAC Trends in Analytical Chemistry*, 110, 150-159.
- [41] Du, H., & Wang, J. (2021). Characterization and environmental impacts of microplastics. *Gondwana Research*, 98, 63-75.
- [42] Chen, G., Fu, Z., Yang, H., & Wang, J. (2020). An overview of analytical methods for detecting microplastics in the atmosphere. *TrAC Trends in Analytical Chemistry*, 130, 115981.
- [43] Lv, L., Yan, X., Feng, L., Jiang, S., Lu, Z., Xie, H., Sun, S., Chen, J., & Li, C. (2021). Challenge for the detection of microplastics in the environment. *Water Environment Research*, 93(1), 5-15.
- [44] Löder, M. G., & Gerdts, G. (2015). Methodology used for the detection and identification of microplastics—a critical appraisal. *Marine Anthropogenic Litter*, 201-227.

- [45] Santos, A. G., da Rocha, G. O., & de Andrade, J. B. (2019). Occurrence of the potent mutagens 2-nitrobenzanthrone and 3-nitrobenzanthrone in fine airborne particles. *Scientific Reports*, 9(1), 1.
- [46] Nelms, S. E., Barnett, J., Brownlow, A., Davison, N. J., Deaville, R., Galloway, T. S., Lindeque, P. K., Santillo, D., & Godley, B. J. (2019). Microplastics in marine mammals stranded around the British coast: ubiquitous but transitory? *Scientific Reports*, 9(1), 1075.
- [47] Novillo, O., Raga, J. A., & Tomás, J. (2020). Evaluating the presence of microplastics in striped dolphins (*Stenella coeruleoalba*) stranded in the western Mediterranean Sea. *Marine Pollution Bulletin*, 160, 111557.
- [48] Zhu, J., Yu, X., Zhang, Q., Li, Y., Tan, S., Li, D., Yang, Z., & Wang, J. (2019). Cetaceans and microplastics: first report of microplastic ingestion by a coastal delphinid, *Sousa chinensis*. *Science of the Total Environment*, 659, 649-654.
- [49] Li, Q., Sun, C., Wang, Y., Cai, H., Li, L., Li, J., Shi, H. (2019). Fusion of microplastics into the mussel byssus. *Environmental Pollution*, 252(A), 420-426.
- [50] Von Moos, N., Burkhardt-Holm, P., & Köhler, A. (2012). Uptake and effects of microplastics on cells and tissue of the blue mussel *Mytilus edulis* L. after an experimental exposure. *Environmental Science and Technology*, 46(20), 11327-11335.
- [51] Ding, J., Li, J., Sun, C., Jiang, F., Ju, P., Qu, L., Zheng, Y., & He, C. (2019). Detection of microplastics in local marine organisms using a multi-technology system. *Analytical Methods*, 11(1), 78-87.
- [52] Li, Q., Ma, C., Zhang, Q., & Shi, H. (2021). Microplastics in shellfish and implications for food safety. *Current Opinion in Food Science*, 40, 192-197.



- [53] Campanale, C., Massarelli, C., Savino, I., Locaputo, V., & Uricchio, V. F. (2020). A detailed review study on potential effects of microplastics and additives of concern on human health. *International Journal of Environmental Research and Public Health*, 17(4), 1212.
- [54] Athey, S. N., Albotra, S. D., Gordon, C. A., Monteleone, B., Seaton, P., Andrady, A. L., Taylor, A. R., & Brander, S. M. (2020). Trophic transfer of microplastics in an estuarine food chain and the effects of a sorbed legacy pollutant. *Limnology and Oceanography Letters*, 5(1), 154-162.
- [55] Mitrano, D. M., & Wohlleben, W. (2020). Microplastic regulation should be more precise to incentivize both innovation and environmental safety. *Nature Communications*, 11(1), 1-12.
- [56] Anagnosti, L., Varvaresou, A., Pavlou, P., Protopapa, E., & Carayanni, V. (2021). Worldwide actions against plastic pollution from microbeads and microplastics in cosmetics focusing on European policies. Has the issue been handled effectively? *Marine Pollution Bulletin*, 162, 111883.
- [57] Li, H. X., Ma, L. S., Lin, L., Ni, Z. X., Xu, X. R., Shi, H. H., & Rittschof, D. (2018). Microplastics in oysters *Saccostrea cucullata* along the Pearl River estuary, China. *Environmental Pollution*, 236, 619-625.
- [58] Wang, W. X., Pan, K., Tan, Q., Guo, L., & Simpson, S. L. (2014). Estuarine pollution of metals in China: science and mitigation.
- [59] Korringa, P. (1952). Recent advances in oyster biology. *The quarterly review of biology*, 27(3), 266-308.
- [60] Setälä, O., Norkko, J., & Lehtiniemi, M. (2016). Feeding type affects microplastic ingestion in a coastal invertebrate community. *Marine pollution bulletin*, 102(1), 95-101.
- [61] Arapov, J., Ezgeta–Balić, D., Peharda, M., & Ninčević Gladan, Ž. (2010). Bivalve feeding—how and what they eat?. *Croatian Journal of Fisheries: Ribarstvo*, 68(3), 105-116.

- [62] Zuschin, M., & Baal, C. (2007). Large gryphaeid oysters as habitats for numerous sclerobionts: a case study from the northern Red Sea. *Facies*, 53(3), 319-327.
- [63] Cole, M., Lindeque, P. K., Fileman, E., Clark, J., Lewis, C., Halsband, C., & Galloway, T. S. (2016). Microplastics alter the properties and sinking rates of zooplankton faecal pellets. *Environmental science & technology*, 50(6), 3239-3246.
- [64] Ward, J. E., Zhao, S., Holohan, B. A., Mladinich, K. M., Griffin, T. W., Wozniak, J., & Shumway, S. E. (2019). Selective ingestion and egestion of plastic particles by the blue mussel (*Mytilus edulis*) and eastern oyster (*Crassostrea virginica*): implications for using bivalves as bioindicators of microplastic pollution. *Environmental science & technology*, 53(15), 8776-8784.
- [65] Severini, M. D. F., Villagran, D. M., Buzzi, N. S., & Sartor, G. C. (2019). Microplastics in oysters (*Crassostrea gigas*) and water at the Bahía Blanca Estuary (Southwestern Atlantic): An emerging issue of global concern. *Regional Studies in Marine Science*, 32, 100829.
- [66] Thomas, M., Jon, B., Craig, S., Edward, R., Ruth, H., John, B., & Matthew, S. (2020). The world is your oyster: low-dose, long-term microplastic exposure of juvenile oysters. *Heliyon*, 6(1), e03103.
- [67] Trites, A. W. (2019). Marine mammal trophic levels and trophic interactions. *Encyclopedia of Ocean Sciences*. Elsevier, 589-594.
- [68] Geisler, J. H., McGowen, M. R., Yang, G., & Gatesy, J. (2011). A supermatrix analysis of genomic, morphological, and paleontological data from crown Cetacea. *BMC evolutionary biology*, 11, 1-33.
- [69] Bannister, J. L. (2009). Baleen whales (mysticetes). In *Encyclopedia of marine mammals* (pp. 80-89). Academic Press.

- [70] Hooker, S. K. (2018). Toothed whales (Odontoceti). In *Encyclopedia of marine mammals* (pp. 1004-1010). Academic Press.
- [71] Heithaus, M. R. (2001). Predator–prey and competitive interactions between sharks (order Selachii) and dolphins (suborder Odontoceti): a review. *Journal of Zoology*, 253(1), 53-68.
- [72] Dalpadado, P., & Mowbray, F. (2013). Comparative analysis of feeding ecology of capelin from two shelf ecosystems, off Newfoundland and in the Barents Sea. *Progress in Oceanography*, 114, 97-105.
- [73] Baumgart, D. C., & Dignass, A. U. (2002). Intestinal barrier function. *Current Opinion in Clinical Nutrition & Metabolic Care*, 5(6), 685-694.
- [74] Kanwar, R., & Kaur, I. (2012). Role of functional groups of adsorbents in the adsorption process. *Reviews in Chemical Engineering*, 28(1-2), 47-87.
- [75] Daley, J. M., Paterson, G., & Drouillard, K. G. (2014). Bioamplification as a bioaccumulation mechanism for persistent organic pollutants (POPs) in wildlife. *Reviews of Environmental Contamination and Toxicology, Volume 227*, 107-155.
- [76] Gobas, F. A. P. C., & Morrison, H. A. (2000). Bioconcentration and biomagnification in the aquatic environment. *Handbook of property estimation methods for chemicals*, 189-231.
- [77] Reddy, A. V. B., Moniruzzaman, M., & Aminabhavi, T. M. (2019). Polychlorinated biphenyls (PCBs) in the environment: Recent updates on sampling, pretreatment, cleanup technologies and their analysis. *Chemical Engineering Journal*, 358, 1186-1207.
- [78] Desforges, J. P., Hall, A., McConnell, B., Rosing-Asvid, A., Barber, J. L., Brownlow, A., Guise, S. D., Eulaers, I., Jepson, P. D., Letcher, R. J., Levin, M., Ross, P. S., Vikingson, G., Sonne, C., & Dietz, R. (2018). Predicting global killer whale population collapse from PCB pollution. *Science*, 361(6409), 1373-1376.

- [79] Ali, H., Khan, E., & Ilahi, I. (2019). Environmental chemistry and ecotoxicology of hazardous heavy metals: environmental persistence, toxicity, and bioaccumulation. *Journal of Chemistry*, 2019.
- [80] Ajith, N., Arumugam, S., Parthasarathy, S., Manupoori, S., & Janakiraman, S. (2020). Global distribution of microplastics and its impact on marine environment—a review. *Environmental Science and Pollution Research*, 27, 25970-25986.
- [81] Rochman, C. M., Tahir, A., Williams, S. L., Baxa, D. V., Lam, R., Miller, J. T., & Teh, S. J. (2015). Anthropogenic debris in seafood: Plastic debris and fibers from textiles in fish and bivalves sold for human consumption. *Scientific reports*, 5(1), 1-10.
- [82] Battaglia, F. M., Beckingham, B. A., & McFee, W. E. (2020). First report from North America of microplastics in the gastrointestinal tract of stranded bottlenose dolphins (*Tursiops truncatus*). *Marine Pollution Bulletin*, 160, 111677.
- [83] Jia, C., & Jukes, D. (2013). The national food safety control system of China—a systematic review. *Food control*, 32(1), 236-245.
- [84] Escanciano, C., & Santos-Vijande, M. L. (2014). Reasons and constraints to implementing an ISO 22000 food safety management system: Evidence from Spain. *Food Control*, 40, 50-57.
- [85] Das, A. K., Nanda, P. K., Das, A., & Biswas, S. (2019). Hazards and safety issues of meat and meat products. In *Food safety and human health* (pp. 145-168). Academic Press.
- [86] Amagliani, G., Brandi, G., & Schiavano, G. F. (2012). Incidence and role of Salmonella in seafood safety. *Food Research International*, 45(2), 780-788.
- [87] El Sheikha, A. F., & Xu, J. (2017). Traceability as a key of seafood safety: reassessment and possible applications. *Reviews in fisheries science & aquaculture*, 25(2), 158-170.

- [88] Huss, H. H., Reilly, A., & Karim Ben Embarek, P. K. B. (2000). Prevention and control of hazards in seafood. *Food Control*, 11(2), 149-156.
- [89] Lipp, E. K., & Rose, J. B. (1997). The role of seafood in foodborne diseases in the United States of America. *Revue Scientifique et Technique*, 16(2), 620-640.
- [90] Blackburn, K., & Green, D. (2022). The potential effects of microplastics on human health: what is known and what is unknown. *Ambio*, 51(3), 518-530.
- [91] Vethaak, A. D., & Legler, J. (2021). Microplastics and human health. *Science*, 371(6530), 672-674.
- [92] Koelmans, A. A. (2016). *Environmental Toxicology and Chemistry*, 50, 3315.
- [93] Ribeiro, F., O'Brien, J. W., Galloway, T., & Thomas, K. V. (2019). Accumulation and fate of nano- and micro-plastics and associated contaminants in organisms. *TrAC Trends in Analytical Chemistry*, 111, 139-147.
- [94] Van Cauwenberghe, L., & Janssen, C. R. (2014). Microplastics in bivalves cultured for human consumption. *Environmental pollution*, 193, 65-70.
- [95] Ahmad, M., Li, J. L., Wang, P. D., Hozzein, W. N., & Li, W. J. (2020). Environmental perspectives of microplastic pollution in the aquatic environment: a review. *Marine Life Science & Technology*, 2(4), 414-430.

## **Chapter 2: Innovative Methodology**

### **2.1 Observation methods for the study of microplastics**

It is important to learn how to accurately identify and quantify the presence of microplastics in various samples. Here we describe an improved technique using the optical microscope and fluorescence microscope commonly used for microplastic observation techniques [1].

Optical and fluorescence microscopy enables researchers to identify and characterize the physical characteristics of microplastics, such as their size, shape, and composition. However, these techniques can be time-consuming and labor-intensive and may not be appropriate for large-scale studies [2]. Chemical analysis techniques, such as Fourier transform infrared spectroscopy (FTIR) and gas chromatography-mass spectrometry (GC-MS), are another method to detect the presence of microplastics [3] allowing the determination of the chemical compositions of microplastics [4]. The advantages and disadvantages of these methods are summarised in Table 2.1. However, not all observation techniques are appropriate for on-site observation. In other words, they are too heavy to carry to the field. Therefore, an innovative portable observation system was devised for this study.

**Table 2.1** Comparison of analytical devices [5-7].

<b>Device name</b>	<b>Advantages</b>	<b>Disadvantages</b>
Optical microscopy	Simple, fast, easy	No chemical confirmation – High possibility of false positive data – High possibility of missing small and transparent particles – No polymer composition data
FTIR spectroscopy	–Higher reliability concerning possible false positive data by chemical confirmation of all plastic-like particles – Reduction of false negative data –Nondestructive – Detection of less than 10 mm fibres –Automatic mapping (FPA reflectance)	– Expensive – Laborious work and time-consuming for whole particle identification –Contact analysis (ATR)
Raman spectroscopy	–Higher reliability concerning possible false positive data by chemical confirmation of all plastic-like particles – Reduction of false negative data – Detection of less than 1 mm of plastics – Nondestructive – Noncontact	–Expensive

## 2.2 Portable observation system

The new portable observation system used in this research consists of a sample stage, light sources, and digital microscope assembled with a Wi-Fi system. The specification of the digital microscope is 5.8 x 1.6 inches. The resolution is 1,920 x 1,080P (2 million pixels). There are eight adjustable LED lights on the end of the camera with excellent colour temperature in dark environments. Other features include portability, observation using various light sources, and simple handling. It is also very useful for fieldwork research. Wi-Fi enables the use of a smart phone or tablet instead of a computer.

Additionally, the sample stage has a drainage system for easy dissection of experimental samples. Another advantage is that it can be powered through USB mobile batteries. Consequently, it can function without an AC power supply. Figure 2.1 depicts the procedure of dissecting a sample (oysters are used as an example). The oyster shells were removed prior to the digestion experiment and the surface of the oysters was slightly washed. The tissue was then settled on the stage. After dissection, the gills and the digestive system were observed. Microplastics are easily visible under UV light. The surface of the intestine was also examined for microplastic detection using the same method. Following the digestion experiment, a digital microscope was used again to observe the microplastics on the surface of the filter paper.



1. Experimental material: Oysters, Potassium Hydroxide (KOH) (Wako), Filter paper(20  $\mu\text{m}$ ~25  $\mu\text{m}$ ), Scaler, Optical microscope Nikon ECLIPSE 50i.



2. Removed the shell of oyster and washed the surface of oyster slightly.



3. Poured 10% KOH solution and used the water bath to accelerate the digestion.



4. Used the filtration system, and the solution was sucked through the filter paper by a vacuum pump.



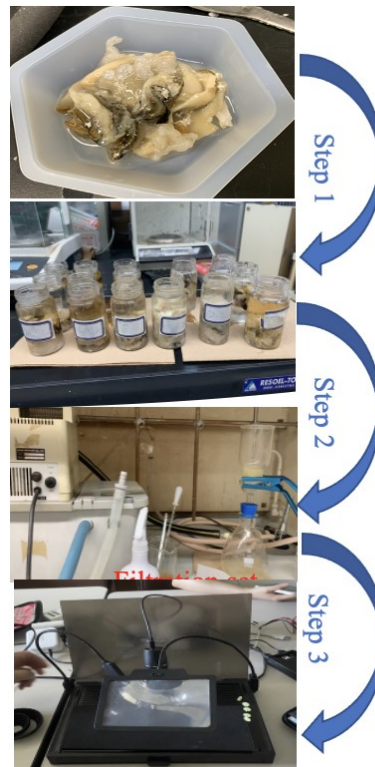
5. Fix filter paper onto flat and labelled plastic disk which is then placed into a labelled Petri dish. Seal Petri dish with sticky tape.



6. Observed by digital microscope.

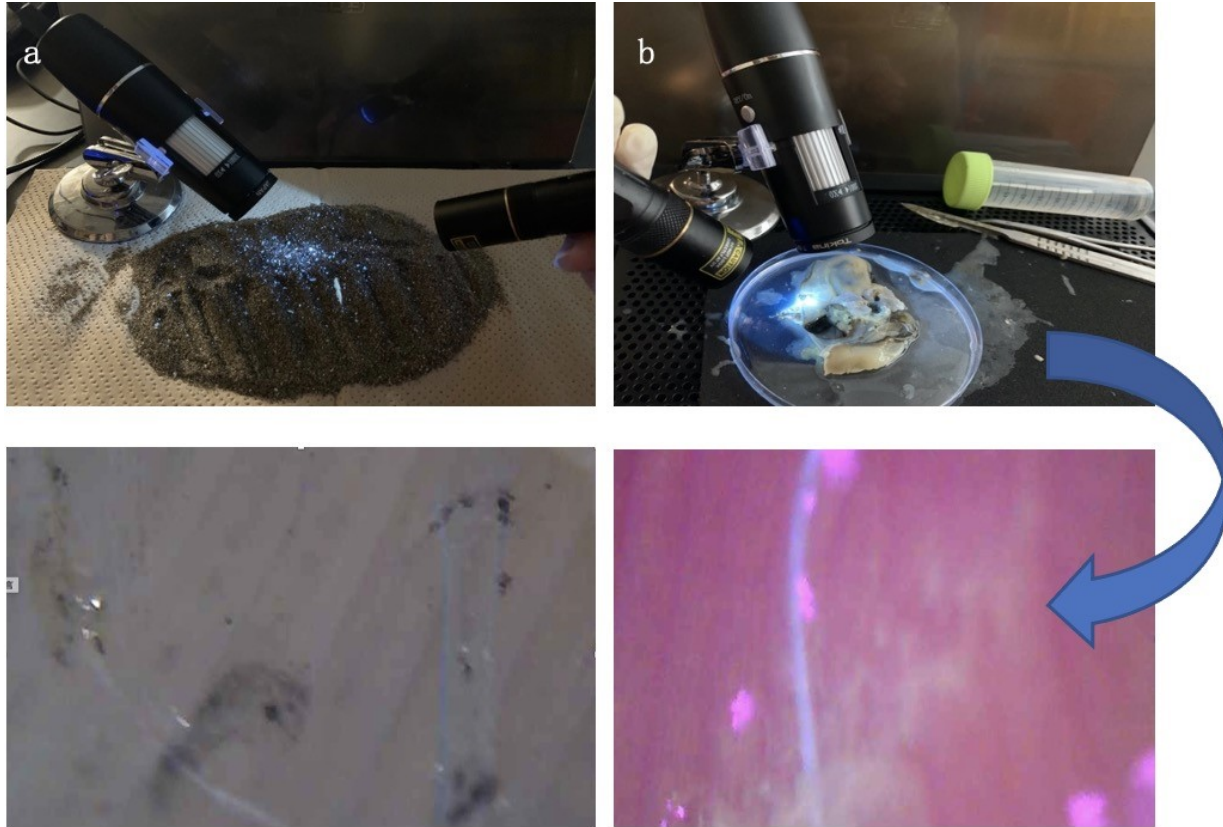


**Experiment Measurement by IR or Raman**



**Figure 2.1** Flow chart showing the use of the new portable system machines and chemical materials for the oyster study.

Fluorescence lights were used to observe the microplastics in the samples. In order to excite the plastics, a light source with a wavelength of 355 nm was used, and a 420 nm filter was mounted onto the lens of the digital microscope. During the observation process, it is necessary to switch filters with different wavelengths. As shown in Figure 2.2, several bright spots were observed, indicating the presence of microplastics in the sample. The optical magnification zoom can reach up to 1,000 times. The system can be used to observe microscopic objects invisible to the naked eye, monitor pores and skin, check cavities of animals and plants, close-up of electronic circuits, and machinery, and check substrates. The bottom right image shows the gills of oysters. Water and light reflections were found when the object was observed at different angles. Flexible arms are well suited to find good angles of observation. This procedure was applied in this research to observe oysters and cetacean intestine samples. This portable observation system allows the observation of both wet and dry samples which is very convenient. Furthermore, its numerous advantages allow for its use in the field.



**Figure 2.2** Typical microplastics found in oysters and sand: a) a digital microscope is used to observe the reflective parts in the sand under UV light. b) White and translucent blue fibres found in the gills of oysters.

## 2.3 Reference

- [1] Mendoza, L. M. R., & Balcer, M. (2019). Microplastics in freshwater environments: a review of quantification assessment. *TrAC Trends in Analytical Chemistry*, *113*, 402-408.
- [2] Auta, H. S., Emenike, C. U., & Fauziah, S. H. (2017). Distribution and importance of microplastics in the marine environment: a review of the sources, fate, effects, and potential solutions. *Environment International*, *102*, 165-176.
- [3] Primpke, S., Fischer, M., Lorenz, C., Gerdt, G., & Scholz-Böttcher, B. M. (2020). Comparison of pyrolysis gas chromatography/mass spectrometry and hyperspectral FTIR imaging spectroscopy for the analysis of microplastics. *Analytical and Bioanalytical Chemistry*, *412*(30), 8283-8298.
- [4] Campanale, C., Massarelli, C., Savino, I., Locaputo, V., & Uricchio, V. F. (2020). A detailed review study on potential effects of microplastics and additives of concern on human health. *International Journal of Environmental Research and Public Health*, *17*(4), 1212.
- [5] Wang, Z. M., Wagner, J., Ghosal, S., Bedi, G., & Wall, S. (2017). SEM/EDS and optical microscopy analyses of microplastics in ocean trawl and fish guts. *Science of the Total Environment*, *603*, 616-626.
- [6] Veerasingam, S., Ranjani, M., Venkatachalapathy, R., Bagaev, A., Mukhanov, V., Litvinyuk, D., Mugilarasan, M., Gurumoorrthi, K., Gunganathan, L., Aboobacker, V. M., & Vethamony, P. (2021). Contributions of Fourier transform infrared spectroscopy in microplastic pollution research: a review. *Critical Reviews in Environmental Science and Technology*, *51*(22), 2681-2743.
- [7] Lenz, R., Enders, K., Stedmon, C. A., Mackenzie, D. M., & Nielsen, T. G. (2015). A critical assessment of visual identification of marine microplastic using Raman spectroscopy for analysis improvement. *Marine Pollution Bulletin*, *100*(1), 82-91.

# **Chapter 3: Microplastics in cultured oysters from different coastal areas of Japan**

## **3.1 Introduction**

Studies dealing with the current extent of human impacts on marine ecosystems show that the oceans are becoming highly affected by human activities [1]. The accumulation of mismanaged plastic waste is now a global concern [2].

Since the 1970s, research on the presence of plastic debris in the ocean has shown an increase in the average concentration of plastic particles. For example, a study from the western Sargasso Sea indicated an average of 3,500 pieces and 290 grams per square kilometer [3]. Not being biodegradable in most cases, plastics widespread use constitutes a persistent environmental burden [4,5].

Plastic is a cheap, lightweight material that can easily be molded into various shapes. This has increased its annual world production from 1.7 million tons in the 1950s to 288 million tons in 2012 [6]. Our lives now seem to be inseparable from plastic. However, its negative effects, including many adverse effects on life and the environment, are becoming of the utmost concern [6].

Data show that most plastic waste in the oceans comes from terrestrial sources and is carried mostly by winds, wastewater treatment systems and road runoff [7,8], with the remainder coming from artifacts such as fish nets and ropes, aquaculture cages, and so on. [9]. Among microplastic garbage, plastic bags harm marine life the most [10]. Modern plastics are usually complex mixtures of polymers with chemical additives [11]. Plastic waste is usually colonized by bacteria [12] and often leaches chemical pollutants [13] into the waters. As mentioned, plastics are

widely distributed on the surface and coastal waters of the global oceans, and their impact on marine life is enormous. For example, 44% of all seabird species are known to ingest plastic, as do sea turtles, cetaceans, and fish [14-16].

In recent years, more attention has been paid to microplastics [17], particles less than 5 mm that result from the breakdown of larger items [3]. Among the possible sources of these particles are waters from washing synthetic textiles discharged into the ocean. From the perspective of its composition, polypropylene (PP), polyethylene (PE), polystyrene (PS), polyvinyl chloride (PVC) and polyethylene terephthalate (PET) are the main components with higher probabilities of ending up in the ocean environment [18]. Indeed, several studies have found that both food and drinking water [19,20]. may contain different types of microplastics. They have also been found in human feces [21], increasing awareness of their possible effects on human health.

Compared with large plastics, microplastics are much more abundant and easily absorbed by organisms, thus increasing their toxicity potential. A Beijiang River surface sediment study found that microplastics were loaded with metals such as Ni, Cd, Pb, Cu, Zn, and Ti, mostly derived from inherent load [22]. Strong sorption of PCBs (polychlorinated biphenyls) to nano plastics and microplastics was also observed [23], as well as the presence of contaminants, such as polyfluoroalkyl substances (PFAS) [24].

Microplastics can also transport pollutants into marine food webs [25]. Moreover, in China, researchers established a model of microplastic accumulation in Baiyangdian Lake. The results showed that they spread and accumulate very quickly throughout the food web, eventually reaching high trophic-level aquatic organisms [21]. Studies from six locations along the French–Belgian–Dutch coastline looked at the absorption of microplastics under field conditions detecting them in mussels (*Mytilus edulis*) and lugworms (*Arenicola marina*). In France, microplastics were

detected throughout the Bay of Brest, with polyethylene (PE, 53-67%), polypropylene (PP, 16-30%), and polystyrene (PS, 16-17%) fragments dominating both the surface water and the sediments [26]. Microplastic contamination of bivalves from the French Atlantic coast showed a high ratio of grey colour particles with sizes ranging from 50 to 100  $\mu\text{m}$  in blue mussels (*Mytilus edulis*) and Pacific oysters (*Magallana gigas*) [27].

In Japan, microplastics were found in 31 out of 36 sites along 29 rivers where samples were collected in surface waters. Results show that the concentration of microplastics highly correlates with urbanization and population density and that the number of particles in rivers seems to depend on the human activities along the river basins [28]. Furthermore, microplastics were found in the digestive tract of 64 anchovies (*Engraulis japonicus*) in Tokyo Bay, representing 77% of the samples studied [17]. The correlation of microplastics' amount and human activities was also found in a study of coastal surface water done in the subtropical island of Okinawa, Japan [29].

Studies from the eastern coast of Australia seaports, report large amounts of microplastics in both sediments and oysters. Note that the density of particles in oysters was significantly higher than in sediments and black fibres between 0.1 mm–0.5 mm in size were the most abundant [30]. Experiments with oysters (*Saccostrea glomerata*) show that it can ingest and accumulate microplastics of 2  $\mu\text{m}$  and 0.5  $\mu\text{m}$ , that are then transferred through the membranes of the digestive glands to the haemolymph [31]. These studies emphasize the necessity to monitor microplastics in aquaculture and in the ocean, in order to enforce seafood safety.

Shellfish is an important component of marine ecosystems. Oysters are filter feeding organisms and as such they are especially easy to come into contact with microplastics [32]. Accumulating floating microparticles during the filtration process. Seafood products are important sources of proteins, polyunsaturated lipids, and phospholipids. Economic losses (Microplastic

toxicity in marine organism, the cost of saving a huge amount of marine litter, etc.) due to microplastic pollution and marine debris in the marine environment are expected [33-35]. In Japan, oysters are very popular as food. With the development of the aquaculture industry, oysters (Mostly *Crassostrea spp.*) have become even more looked after. So far, Japanese researchers have predominantly studied viruses and their threat to oysters [36-38]. As the threat of microplastics to marine life becomes ever more serious, oyster aquaculture is expected to be highly affected.

Japan is an important oyster producer. Its oysters are exported worldwide and considered of prime quality. In this context, material from various aquaculture farms in Japan are studied in order to analyze the presence, quantity, size, and type of microplastics present to assess its the level in farmed oysters. Normally, farmed oysters are generally pre-washed, and UV irradiated before entering the market, so our research also aims to test the efficiency of these pretreatments on microplastics content. The possible sources of microplastics and its potential health risks are also discussed.



## 3.2 Materials and methods

### 3.2.1 Sampling sites

Oysters (*Magallana gigas*) were obtained at 12 locations along the Northern Island, Pacific Ocean, Sea of Japan, Inland Island, and Southern Islands. The objectives were to: (1) collect oysters from various agricultural farms and search for wild oysters near the coast; (2) take near-shore photos (environmental survey); (3) survey the surrounding living facilities (factories, living areas, etc.); and (4) compare the distribution of ocean currents. For this survey, the Fishery Adjustment Division of the Fishery Association Management Section of the Fisheries and Forest Department of the Hokkaido Government issued permission to take natural oysters. Fishery Cooperative Associations also helped to sample natural oysters in each area. The primary oyster breeding places in Japan were considered when choosing the sample sites for investigation.

The sampling also included cultured oysters bought from twelve sites (Table 3.1). This material was approximately two years old (personal communication from the seller). At least four oysters were sampled from each location. In order to avoid microplastic contamination, oysters were kept under freezing conditions, wrapped in aluminium foil for further experiments. A total of 106 collected samples were studied (Table 3.1).

During the collection of wild material, plastic waste was observed in a considerable amount of fishing nets and lines, as well as other plastic garbage abandoned on beaches near the areas where some of the studied material came. It is assumed that these discarded materials could be sources for the microplastics found in the studied materials.

**Table 3.1** Location and average weights of oysters (without the shell) from different coastal areas of Japan.

	<b>Location</b>	<b>Number of specimens</b>	<b>Average weight per oyster without shell (g)</b>
S1	Northern island	10	30.50
S2	Pacific Ocean	20	35.19
S3	Northern island	5	41.50
S4	Pacific Ocean	5	24.39
S5	Pacific Ocean	5	19.95
S6	Pacific Ocean	4	49.68
S7	Sea of Japan	12	31.33
S8	Pacific Ocean	10	33.61
S9	Pacific Ocean	10	21.44
S10	Inland island	10	18.96
S11	Pacific Ocean	5	23.87
S12	Southern island	10	25.02

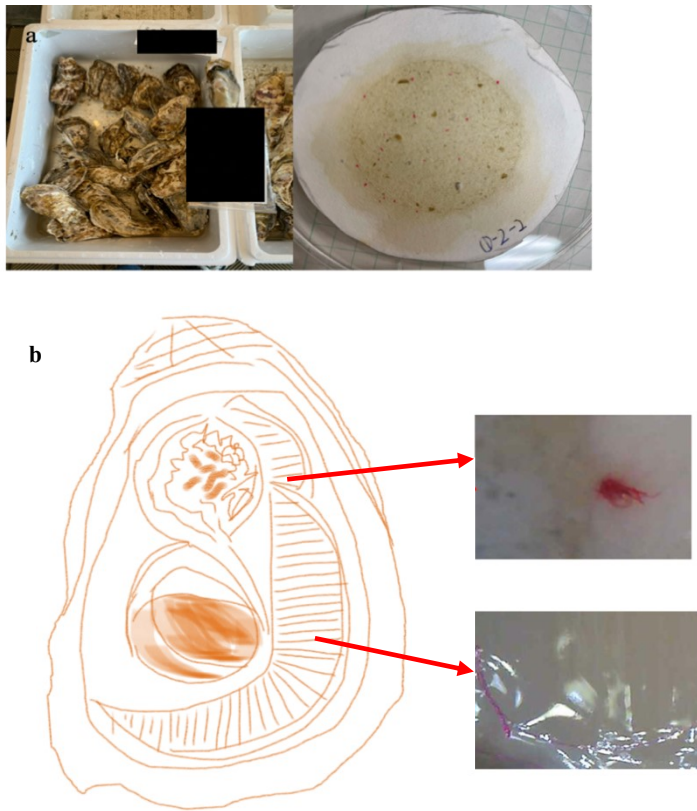
### 3.2.2 Methods

Microplastics were extracted from oysters following Chen's procedure [33]. The shells were removed, and the body's surface was slightly washed. The body was then placed into a 10% KOH solution, and a water bath was used to accelerate the digestion process. With the help of a vacuum pump, the solution was filtered through filter paper (20  $\mu\text{m}$ -25  $\mu\text{m}$ ), and this was fixed onto a flat, labelled plastic disk and then placed into a labelled Petri dish. These were then sealed with sticky tape. A similar filtration procedure was applied to the dissected digestive tract of the oysters.

The dissected digestive and gill systems were observed through a digital microscope to assess the presence of microplastics (Figure 3.1) preliminarily. This simple procedure can be used to obtain images and evaluate the shapes and sizes of the microplastics. The colour can also be used to determine the presence of microplastics. In addition, tweezers were used to assess if the fibres would break. Raman spectra (RENISHAW inVia Raman Microscope) and optical microscopy (Nikon ECLIPSE 50i) were used to identify whether or not a particle was plastic. Raman methodology combines Raman spectroscopy and optical microscopy. It is one of the most efficient and effective ways to identify plastics by focusing a laser beam onto a small spot to obtain a spectrum. Raman spectroscopy can specify different polymers by probing different vibrational modes in the molecule.

Once a spectrum is acquired, the next step is identifying which polymers make up the unknown sample. A complete, high-quality spectral database is critical for accurate spectral matching. In our research, the software Open Specy was used. In this typical database, spectra of reference materials are recorded and stored with their chemical and physical properties in the sample record metadata. The spectra were matched with the spectra from the database. The

software displays the highest matching polymer spectra. The spectra obtained through Raman were loaded into the gallery software, Open Specy, for analysis and identification of the plastic polymer type.



**Figure 3.1** a) Oysters purchased at seafood markets and local breeding bases. After the tissue digestion procedure, microplastics are shown in colour on the filter paper. b) Fibres can be observed in samples using digital microscopy before the tissue digestion procedure.

### 3.3 Results

Microplastics were found in almost all oysters studied. Our results indicate that microplastics were widely detected in cultured oysters in Japan. Microplastics (fibres and fragments mainly) were found in almost all oysters, and showed differences in length, size, and colour. A total of 333 microplastic pieces were collected from 106 oysters. The average number of microplastics per individual amounted to  $3.14 \pm 4.00$  particles.

The average number of microplastics per oyster S4 material contains the highest value (up to  $11.8 \pm 8.5$  items), followed by S5 (up to  $7.8 \pm 10.2$  microplastics/oyster). The least microplastics per individual oyster come from S2, S3 and S10 ( $1.5 \pm 1.5$ ,  $1.4 \pm 1.3$ , and  $1.5 \pm 1.0$ , respectively) (Table 3.2).

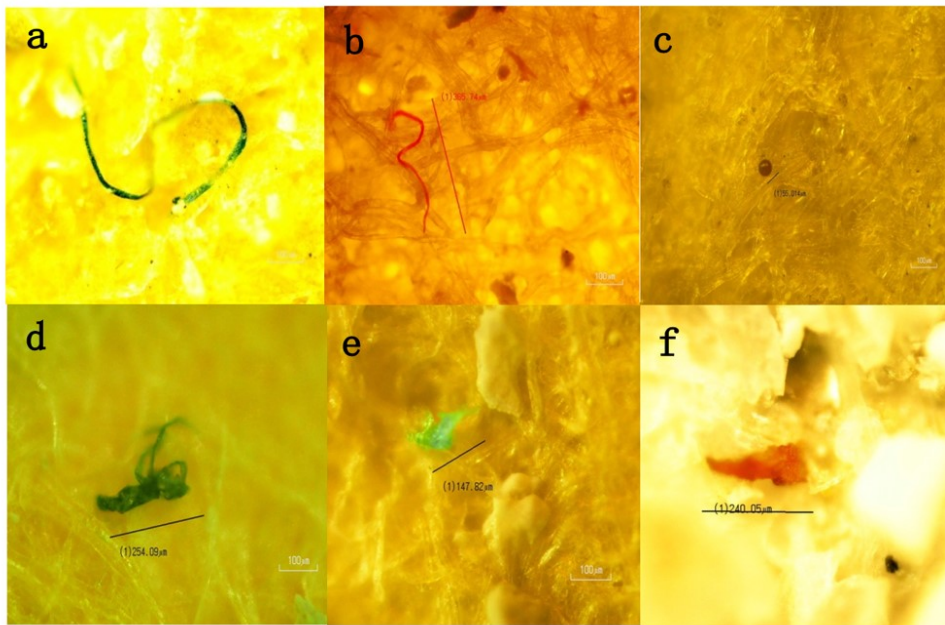
The most common colours of typical fibres were red and green, whereas plastic fragments were mainly red and blue. Indeed, green particles constituted 57% of the total microplastics found, with red accounting for 17%. Fewer microplastics were blue (3%), grey or brown (About 5% in both cases). From the microscopic images, some fibre-formed clusters were observed (Figure 3.2, d). Sometimes, the edges of the fragment are damaged because of the degradation phenomenon (Figure 3.2, e).

Fibres were the most common type forming 88% of the total microplastics. The rest were mostly fragments. The most abundant size (32%) found was 100 to 500, followed by particles over 1000 (30%). Microplastics smaller than 100 were only 9% (Figure 3.3). The observation of dissected oyster gills and digestive systems confirmed the presence of microplastics in the gills of some oysters. The filtered material after the digestion of these organs also showed the presence of microplastic fibres (Figure 3.2). The analyses of the Raman spectra showed that the microplastics found in oyster samples are nylon 6, high-density polyethylene and polypropylene (Figure 3.4).

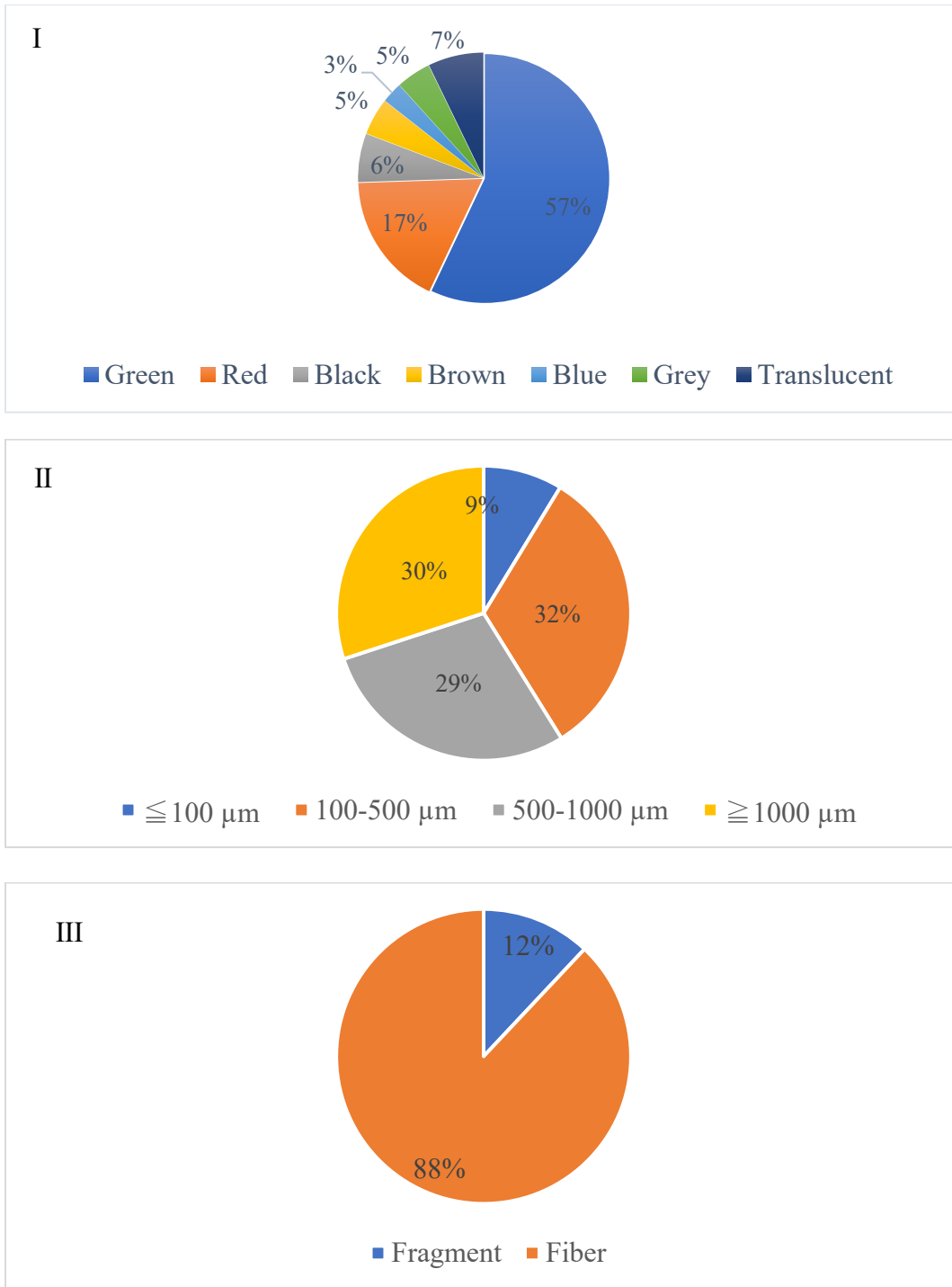
Pictures taken during the field work on several beaches show the presence of domestic garbage, as well as discarded fishing nets and lines (Figure 3.5).

**Table 3.2** Microplastics in oysters from various locations in Japan.

<b>Sample location</b>	<b>Number of specimens</b>	<b>Total number of microplastics per sample</b>	<b>Average number of microplastics per oyster</b>
S1	10	24	$2.4 \pm 1.7$
S2	20	30	$1.5 \pm 1.5$
S3	5	7	$1.4 \pm 1.3$
S4	5	59	$11.8 \pm 8.5$
S5	5	39	$7.8 \pm 10.2$
S6	4	30	$7.5 \pm 5.9$
S7	12	33	$2.8 \pm 1.3$
S8	10	29	$2.9 \pm 1.7$
S9	10	19	$1.9 \pm 1.2$
S10	10	15	$1.5 \pm 1.0$
S11	5	12	$2.4 \pm 1.1$
S12	10	36	$3.6 \pm 2.2$
<b>Total</b>	<b>106</b>	<b>333</b>	<b><math>3.14 \pm 4.0</math></b>

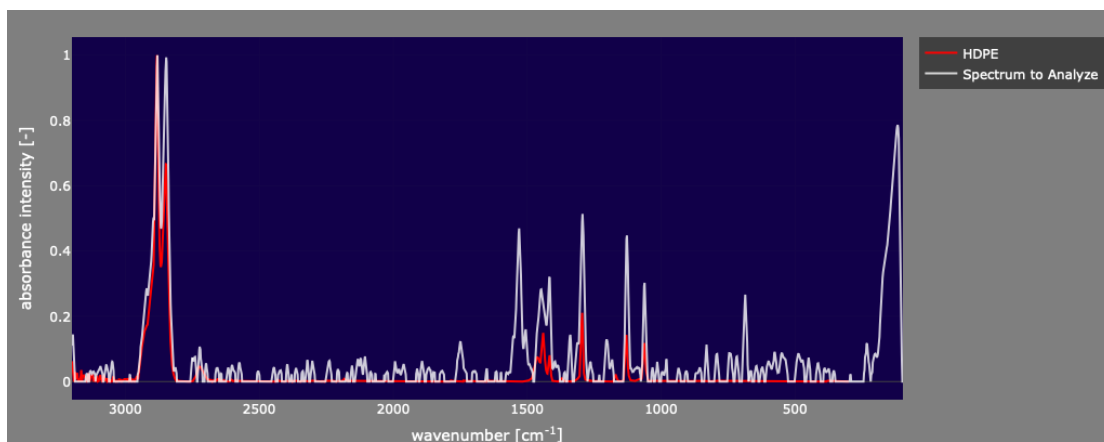
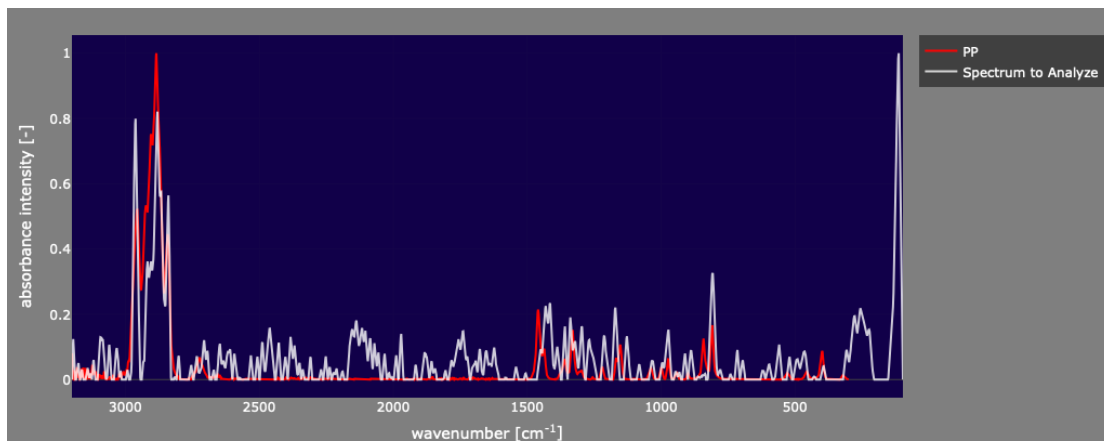
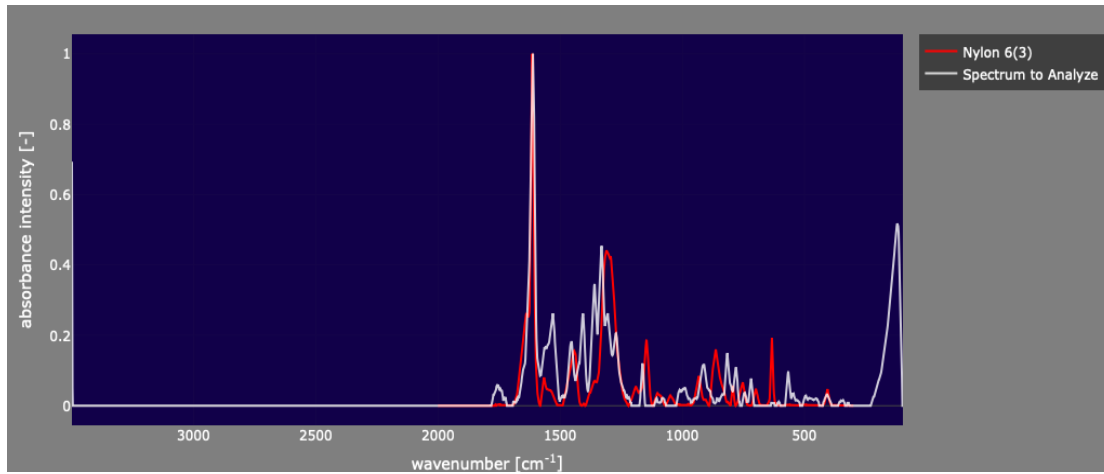


**Figure 3.2** Typical microplastics found in oysters: a) green fibres; b) red fibres; c) brown bead; d) green fibres; e) green fragment; f) red fragment.



**Figure 3.3** Distribution of microplastics found in oysters in Japan according to: (I) colour; (II) size; (III): shape.





**Figure 3.4** Raman spectra of selected microplastics: nylon 6, polypropylene (PP), high density polyethylene (HDPE).



**Figure 3.5** Fishing gills, nets, and domestic garbage stranded on a beach. Hokkaido, Japan, August 2021.

### 3.4 Discussion and conclusions

The most common type of microplastics found in this study were fibres (88%) which agrees with data from 17 coastal sites in China where fibres constitute 60.67% of all microplastics found [22]. It should be noted that our material showed less diversity in the shape of plastics found. Microplastics found in this study can be divided into fragments and fibres (12% and 88%, respectively). Similarly, a study using oysters along the Pearl River Estuary in China indicated the dominance of fibres with a ratio of 69.4% and microplastic size in rock oysters (*Saccostrea cucullata*) shows 83.9% of which were less than 100  $\mu\text{m}$ . On the other hand, our research results show only 9% smaller than 100  $\mu\text{m}$ , whereas the fraction from 100 to 500  $\mu\text{m}$  has the highest proportion, up to 32%. In another study from the French Atlantic coast, the largest fraction of the particles found in the blue mussel (*Mytilus edulis*) and the Pacific oyster (*Magallana gigas*) ranged from 50 to 100  $\mu\text{m}$  [39]. In Korea, four popular bivalve species, oyster (*Magallana gigas*), mussel (*Mytilus edulis*), Manila clam (*Tapes philippinarum*) and scallop (*Patinopecten yessoensis*) were evaluated, finding mainly fragments and particles smaller than 300  $\mu\text{m}$  (76% and 65% respectively) of the total microplastics found [40]. In our study, fibres were the main shape of found.

In this study, the average of microplastics per oyster was 3.14 particles. This is less than what was found in *Perna viridis* (3.28 items/individual) in India (See Table 3.3 for a reference), but the most abundant type found was fibres, the same as in China. In both European and Asian literature, the size of microplastics found was smaller than that of Japanese oysters. On the other hand, the number of microplastics in Japan oysters was higher, and their size was larger than reported in other studies (Table 3.3). Regarding the type of microplastics identified using optical microscopy, most researchers use FT-IR and Raman spectroscopy, which gives confidence to our findings. Therefore, our data can be compared to that of other studies with high confidence.

From our observation, farmed oysters were cleaned and irradiated with ultraviolet rays before being sold. This procedure aims to reduce or eliminate possible microbial contamination of food. Although they were also cleaned using Milli-Q water to rinse both the shell and oysters' bodies before the digestion procedure, microplastics were found. All the cleaning methods done before the oysters are put on the market cannot remove the microplastics completely.

One of the problems concerns the source of the found microplastics. During the collection trips, a large amount of domestic garbage was found stranded on beaches (Figure 3.5). Note that the dates in some plastic packages were relatively recent. In addition, many fishing gills, nets, and cages used in aquaculture lay on the beaches. These are all made up of nylon. Table 3.2 shows that the average number of microplastics in S4, S5, and S6 is higher than in other sites. These sites are located on the eastern Pacific coast, so the pollutants' origin may be domestic garbage and that from neighbouring countries where ocean currents carry it.

Regarding S2, it is located in a relatively closed bay. This is also an aquaculture area, and local fishermen are more likely to pay attention to keeping the water-body healthy.

One research characterized microplastics in the heavily urbanized, brackish water of Vembanad Lake (India), focusing on some commercially important bottom-feeding fishes and shellfish (*Arius maculatus*, *Etroplus suratensis*, *E. maculatus* and *Villorita sp.*). Its results also showed the presence of polyethylene in the samples [41]. Polypropylene and nylon 6 were also reported in south India [42]. Similarly, the main polymers in shellfish from Dongshan Bay, southeastern China, were polyester and Polyethylene terephthalate [43]. The abundance and types of microplastics found exhibited great variation among species. Various microplastics are present in marine organisms due to their durability and wide range of applications in fishing activities and packing and textile industries.

Shellfish consumed as food may carry microplastics that will affect human health. As they are often consumed whole, the accumulation process in the consumers is a real possibility. Furthermore, plastics may have heavy metals and other toxic coatings that will carry on to the final consumer. These concerns lead us to recommend more studies of the possible pathologic effects of microplastics ingested with shellfish to understand their impacts in public health better.

**Table 3.3** Studies showing microplastic contamination of marine life: location, type of digestion, identification method of microplastics, quantity, and abundant type with the corresponding references [22,44,45,46].

Species	Location	Digestion	Identification	Quantity	Most common types
<i>Mytilus edulis</i>	Germany	HNO <sub>3</sub> 69%	Raman spectroscopy	0.36 ± 0.07 items/g (without depuration)	fibre and particle
<i>Magallana gigas</i>	France	HNO <sub>3</sub> 69%	Raman spectroscopy	0.47±0.16 items/g (without depuration)	fibre and particle
<i>Perna viridis</i>	India	10% KOH	Fluorescence microscope	3.28 ± 0.87 items/individual	fragment
<i>Meretrix meretrix</i>	India	10% KOH	Raman spectroscopy	0.5 ± 0.11 items/individual	fragment
<i>Mytilus galloprovincialis</i>	Portugal	10% KOH	FT-IR spectroscopy	0.45 ± 0.67 items/individual	fibre
<i>Scrobicularia plana</i>	Portugal	10% KOH	FT-IR spectroscopy	0.30 ± 0.63 items/individual	fibre
<i>Marphysa sanguinea</i>	Portugal	10% KOH	FT-IR spectroscopy	0.40 ± 0.88 items/individual	fragment
<i>Trachurus trachurus</i>	Portugal	10% KOH	FT-IR spectroscopy	2.24 ± 2.05 items/individual	fibre
<i>Scomber colias</i>	Portugal	10% KOH	FT-IR spectroscopy	2.46 ± 4.12 items/individual	fibre
<i>Magallana gigas</i> <i>Crassostrea angulate</i> <i>Crassostrea hongkongensis</i> <i>Crassostrea sikamea</i>	China	10% (m/v) KOH + 30% H <sub>2</sub> O <sub>2</sub>	FT-IR spectroscopy	2.93 items/individual	fibre

### 3.5 Reference

- [1] Halpern, B. S., Walbridge, S., Selkoe, K. A., Kappel, C. V., Micheli, F., d'Agrosa, C., & Watson, R. (2008). A global map of human impact on marine ecosystems. *science*, 319(5865), 948-952.
- [2] Lebreton, L., & Andrady, A. (2019). Future scenarios of global plastic waste generation and disposal. *Palgrave Communications*, 5(1), 1-11.
- [3] Carpenter, E. J., & Smith Jr, K. L. (1972). Plastics on the Sargasso Sea surface. *Science*, 175(4027), 1240-1241.
- [4] Brems, A., Baeyens, J., & Dewil, R. (2012). Recycling and recovery of post-consumer plastic solid waste in a European context. *Thermal Science*, 16(3), 669-685.
- [5] Haribowo, R., Yoshimura, M., Sekine, M., Imai, T., Yamamoto, K., Higuchi, T., & Kanno, A. (2017). Behavior of toxicity in river basins dominated by residential areas. *Contemporary Engineering Sciences*, 10(7), 305-315.
- [6] Geyer, R., Jambeck, J. R., & Law, K. L. (2017). Production, use, and fate of all plastics ever made. *Science advances*, 3(7), e1700782.
- [7] Paul, R., Chatterjee, D. P., & Dutta, K. (2016). Concerns regarding “plastic” pollution: reasons, effects and needs to generate public awareness. *Int. J. Humanities Social Sci. Studies*, 3, 123-148.
- [8] Boucher, J., & Friot, D. (2017). *Primary microplastics in the oceans: a global evaluation of sources* (Vol. 10). Gland, Switzerland: Iucn.
- [9] Sebille D. E. V., Spathi, D. C., & Gilbert A (2016). The ocean plastic pollution challenge: towards solutions in the UK. *Grant Brief Pap 19*, 1-16.

- [10] Hardesty, B. D., Good, T. P., & Wilcox, C. (2015). Novel methods, new results and science-based solutions to tackle marine debris impacts on wildlife. *Ocean & Coastal Management*, *115*, 4-9.
- [11] Hahladakis, J. N., Velis, C. A., Weber, R., Iacovidou, E., & Purnell, P. (2018). An overview of chemical additives presents in plastics: Migration, release, fate and environmental impact during their use, disposal and recycling. *Journal of hazardous materials*, *344*, 179-199.
- [12] Zettler, E. R., Mincer, T. J., & Amaral-Zettler, L. A. (2013). Life in the “plastisphere”: microbial communities on plastic marine debris. *Environmental science & technology*, *47*(13), 7137-7146.
- [13] Koelmans, A. A., Bakir, A., Burton, G. A., & Janssen, C. R. (2016). Microplastic as a vector for chemicals in the aquatic environment: critical review and model-supported reinterpretation of empirical studies. *Environmental science & technology*, *50*(7), 3315-3326.
- [14] Moore, C. J. (2008). Synthetic polymers in the marine environment: a rapidly increasing, long-term threat. *Environmental research*, *108*(2), 131-139.
- [15] Derraik, J. G. (2002). The pollution of the marine environment by plastic debris: a review. *Marine pollution bulletin*, *44*(9), 842-852.
- [16] Sivan, A. (2011). New perspectives in plastic biodegradation. *Current opinion in biotechnology*, *22*(3), 422-426.
- [17] Tanaka, K., & Takada, H. (2016). Microplastic fragments and microbeads in digestive tracts of planktivorous fish from urban coastal waters. *Scientific reports*, *6*(1), 1-8.
- [18] Andrady, A. L. (2011). Microplastics in the marine environment. *Marine pollution bulletin*, *62*(8), 1596-1605.



- [19] Barboza, L. G. A., Vethaak, A. D., Lavorante, B. R., Lundebye, A. K., & Guilhermino, L. (2018). Marine microplastic debris: An emerging issue for food security, food safety and human health. *Marine pollution bulletin*, *133*, 336-348.
- [20] Danopoulos, E., Twiddy, M., & Rotchell, J. M. (2020). Microplastic contamination of drinking water: A systematic review. *PloS one*, *15*(7), e0236838.
- [21] Zhang, N., Li, Y. B., He, H. R., Zhang, J. F., & Ma, G. S. (2021). You are what you eat: Microplastics in the feces of young men living in Beijing. *Science of the total environment*, *767*, 144345.
- [22] Teng, J., Wang, Q., Ran, W., Wu, D., Liu, Y., Sun, S., Liu, H., Cao, R., & Zhao, J. (2019). Microplastic in cultured oysters from different coastal areas of China. *Science of the total environment*, *653*, 1282-1292.
- [23] Velzeboer, I., Kwadijk, C. J. A. F., & Koelmans, A. A. (2014). Strong sorption of PCBs to nanoplastics, microplastics, carbon nanotubes, and fullerenes. *Environmental science & technology*, *48*(9), 4869-4876.
- [24] Scott, J. W., Gunderson, K. G., Green, L. A., Rediske, R. R., & Steinman, A. D. (2021). Perfluoroalkylated substances (Pfas) associated with microplastics in a lake environment. *Toxics*, *9*(5), 106.
- [25] Carbery, M., O'Connor, W., & Palanisami, T. (2018). Trophic transfer of microplastics and mixed contaminants in the marine food web and implications for human health. *Environment international*, *115*, 400-409.
- [26] Frere, L., Paul-Pont, I., Rinnert, E., Petton, S., Jaffré, J., Bihannic, I., Soudant, P., Lambert, C., & Huvet, A. (2017). Influence of environmental and anthropogenic factors on the composition,

concentration and spatial distribution of microplastics: a case study of the Bay of Brest (Brittany, France). *Environmental Pollution*, 225, 211-222.

[27] Koelmans, A. A., Bakir, A., Burton, G. A., & Janssen, C. R. (2016). Microplastic as a vector for chemicals in the aquatic environment: critical review and model-supported reinterpretation of empirical studies. *Environmental science & technology*, 50(7), 3315-3326.

[28] Kataoka, T., Nihei, Y., Kudou, K., & Hinata, H. (2019). Assessment of the sources and inflow processes of microplastics in the river environments of Japan. *Environmental pollution*, 244, 958-965.

[29] Ripken, C., Kotsifaki, D. G., & Chormaic, S. N. (2021). Analysis of small microplastics in coastal surface water samples of the subtropical island of Okinawa, Japan. *Science of the total Environment*, 760, 143927.

[30] Jahan, S., Strezov, V., Weldekidan, H., Kumar, R., Kan, T., Sarkodie, S. A., He, J., Dastjerdi, B., & Wilson, S. P. (2019). Interrelationship of microplastic pollution in sediments and oysters in a seaport environment of the eastern coast of Australia. *Science of the Total Environment*, 695, 133924.

[31] Scanes, E., Wood, H., & Ross, P. (2019). Microplastics detected in haemolymph of the Sydney rock oyster *Saccostrea glomerata*. *Marine Pollution Bulletin*, 149, 110537.

[32] Thomas, M., Jon, B., Craig, S., Edward, R., Ruth, H., John, B., Dick, V. A., Heather, L. A., & Matthew, S. (2020). The world is your oyster: low-dose, long-term microplastic exposure of juvenile oysters. *Heliyon*, 6(1), e03103.

[33] Chen, J. Y. S., Lee, Y. C., & Walther, B. A. (2020). Microplastic contamination of three commonly consumed seafood species from Taiwan: A pilot study. *Sustainability*, 12(22), 9543.

- [34] Cho, Y., Shim, W. J., Jang, M., Han, G. M., & Hong, S. H. (2019). Abundance and characteristics of microplastics in market bivalves from South Korea. *Environmental pollution*, 245, 1107-1116.
- [35] McIlgorm, A., Campbell, H. F., & Rule, M. J. (2011). The economic cost and control of marine debris damage in the Asia-Pacific region. *Ocean & Coastal Management*, 54(9), 643-651.
- [36] Iritani, N., Kaida, A., Abe, N., Kubo, H., Sekiguchi, J. I., Yamamoto, S. P., Goto, K., Tanaka, T., & Noda, M. (2014). Detection and genetic characterization of human enteric viruses in oyster-associated gastroenteritis outbreaks between 2001 and 2012 in Osaka City, Japan. *Journal of medical virology*, 86(12), 2019-2025.
- [37] Nishida, T., Kimura, H., Saitoh, M., Shinohara, M., Kato, M., Fukuda, S., Munemura, T., Mikami, T., Kawamoto, A., Akiyama, M., Kato, Y., Nishi, K., Kozawa, K., & Nishio, O. (2003). Detection, quantitation, and phylogenetic analysis of noroviruses in Japanese oysters. *Applied and environmental microbiology*, 69(10), 5782-5786.
- [38] Ueki, Y., Amarasiri, M., Kamio, S., Sakagami, A., Ito, H., Uprety, S., Umam, A. N., Miura, T., Nguyen, T. H., & Sano, D. (2021). Human norovirus disease burden of consuming *Crassostrea gigas* oysters: A case-study from Japan. *Food Control*, 121, 107556.
- [39] Phuong, N. N., Poirier, L., Pham, Q. T., Lagarde, F., & Zalouk-Vergnoux, A. (2018). Factors influencing the microplastic contamination of bivalves from the French Atlantic coast: location, season and/or mode of life?. *Marine Pollution Bulletin*, 129(2), 664-674.
- [40] Cho, Y., Shim, W. J., Jang, M., Han, G. M., & Hong, S. H. (2019). Abundance and characteristics of microplastics in market bivalves from South Korea. *Environmental pollution*, 245, 1107-1116.

- [41] Nikki, R., Jaleel, K. A., Ragesh, S., Shini, S., Saha, M., & Kumar, P. D. (2021). Abundance and characteristics of microplastics in commercially important bottom dwelling finfishes and shellfish of the Vembanad Lake, India. *Marine Pollution Bulletin*, 172, 112803.
- [42] Devi, S. S., Sreedevi, A. V., & Kumar, A. B. (2020). First report of microplastic ingestion by the alien fish Pirapitinga (*Piaractus brachypomus*) in the Ramsar site Vembanad Lake, south India. *Marine Pollution Bulletin*, 160, 111637.
- [43] Pan, Z., Liu, Q., Xu, J., Li, W., & Lin, H. (2022). Microplastic contamination in seafood from Dongshan Bay in southeastern China and its health risk implication for human consumption. *Environmental Pollution*, 303, 119163.
- [44] Van Cauwenberghe, L., & Janssen, C. R. (2014). Microplastics in bivalves cultured for human consumption. *Environmental pollution*, 193, 65-70.
- [45] Dowarah, K., Patchaiyappan, A., Thirunavukkarasu, C., Jayakumar, S., & Devipriya, S. P. (2020). Quantification of microplastics using Nile Red in two bivalve species *Perna viridis* and *Meretrix meretrix* from three estuaries in Pondicherry, India and microplastic uptake by local communities through bivalve diet. *Marine Pollution Bulletin*, 153, 110982.
- [46] Pequeno, J., Antunes, J., Dhimmer, V., Bessa, F., & Sobral, P. (2021). Microplastics in marine and estuarine species from the coast of Portugal. *Frontiers in Environmental Science*, 18.

# **Chapter 4: Microplastic in the intestines of stranded cetaceans**

## **4.1 Introduction**

Cetaceans can be a soundness indicator of marine ecosystems because they occupy the higher trophic level of the ecosystem [1,2]. The cetacean study can provide current or potential impacts in marine ecosystems, helping to characterise strategies for dealing with the impacts [2,3].

Most cetacean species have ingested a wide range of plastics [4]. It ranges from microplastics, such as particles and fibres, to macroplastics, such as plastic sheets, fishing nets, fishing lines, household items, etc. [see 5 for a review]. The International Whaling Commission considered plastic pollution an urgent issue since it may affect the mortality of cetaceans [6].

Two extant cetacean suborders have a large difference in the feeding ecology, affecting evaluating microplastics' effects in these organisms. Baleen whale (Mysticeti) species can be exposed to plastic pollution from direct uptake of the water and prey contamination as they filter a large amount of water while feeding [7-11]. On the other hand, toothed whales (Odontoceti), which hunt and swallow the prey whole or torn into pieces, should be more prone to be affected by plastics acquired through prey ingestion and trophic transfer [12-16].

Especially, ingestion of microplastics can occur directly from seawater filter feeding and indirectly from prey consumption [5,17]. It may lead bioaccumulation of microplastics [18-20]. A relatively high number of synthetic particles (mean value 0.057 particles/g) was found in krill ingested by fin whales on the western coast of Iceland [21]. Data from off the East coast of Asia found about 45 plastic particles in the body of a fin whale, including fishing lines, parts of fishing nets, plastic filaments, and Styrofoam pieces [22]. Several polymer types (polyethene and

propylene, nylon, and polyethene terephthalate, among others) were all found in eastern north Pacific humpback whales (*Megaptera novaeangliae*) [17,23,24]. Blue whales (*Balaenoptera musculus*) in the Gulf of California feeding on krill are much more exposed to microplastic contamination through prey transfer than humpbacks which feed predominantly on fish [25].

Toothed (Odontoceti) whales are also commonly affected by microplastic pollution from the consumed prey. Beluga whales (*Delphinapterus leucas*) from the Canadian Northwest Territories were observed with a high amount of microplastic particles in their gastrointestinal tracts [26,27]. Similarly, a study of True's beaked whales (*Mesoplodon mirus*) stranded on the coasts of Ireland reported a large number of microplastics mixed with the remains of mesopelagic fish and cephalopods among their stomach contents [28].

It is crucial to increase our understanding of the ingestion of microplastics in cetaceans for various reasons. Firstly, cetaceans are apex predators in the ocean, and they can accumulate high levels of microplastics that can potentially damage their health and survival [29]. Furthermore, these species are valuable indicators of the soundness of marine ecosystems, so microplastics in their bodies can provide valuable insights into the extent and impact of marine plastic pollution [30]. Therefore, the description of microplastics in cetaceans is essential in understanding the impacts of microplastics.

Hokkaido is located in the Northern part of Japan, surrounded by the Pacific Ocean, the Okhotsk sea, and the Sea of Japan. The Northwestern Pacific Ocean is known for its high levels of microplastic pollution [31]. On average, 70 cetacean strandings are observed annually, and the Stranding Network Hokkaido collects and distributes the specimen of stranded cetaceans for academic purposes.

The objective of the study in this chapter was to evaluate the presence of microplastics in cetaceans collected in Hokkaido, Japan, in order to increase the available information related to their potential impacts on marine ecosystems.

## 4.2 Materials and methods

### 4.2.1 Material

Specimens collected by the Stranding Network Hokkaido (SNH) were used for this research. SNH is a non-profit organisation aiming to collect information and specimens of cetacean strandings, including beaching, drifting, and bycatch, established in 2007. They called on the general public, administrative bodies, fishery officials, and others to report information on stranding to the dedicated receiving and reporting desk ‘IRUKA KUJIRA 110’ and collected an average of about 70 cases per year, with a cumulative total of over 900 cases of stranding information by 2021 [32]. Upon receiving information on stranding, SNH investigates the approachability and condition of the specimen and determines the feasibility of conducting a survey. If a survey is deemed feasible, permission is obtained from the relevant authorities, and the survey is carried out. The survey may involve a visit to the site for a full body recovery, or the survey may be carried out on-site where the specimen has drifted ashore or at the disposal site. In either case, specimens are collected by searching for external morphological information and external injuries. The SNH obtains requests in advance from the research institutions for the samples needed for the study. When the specimens are available, SNH sends specimens to the institutes free of charge and unconditionally [33]. The use of stranded specimens allows these studies to proceed without unnecessary collections of living specimens, thus contributing to conservation efforts.

Intestines of 11 cetacean specimens were obtained from different locations in Hokkaido. Table 4.1 shows the detailed information for each cetacean specimen used in this study. A total of 17 samples were prepared and analysed for microplastic detection. A five-centimetre sample was



taken for each received intestine, keeping in mind that the sample was clean from sand and sediment. In 3 cases, plural samples were obtained: SNH21007 (five samples), SNH20091 and SNH190032 (two samples each). In order to avoid contamination, samples were wrapped in aluminium foil and frozen before the analysis. The study of carcasses can be complicated due to uncertainties in determining the stage of decomposition. However, it is crucial to have a system in place to define the quality of the material being studied. The code system was established by the Smithsonian Institution's Scientific Event Alert Network. In Table 4.1, CODE 2 points to carcass in good condition. CODE 3 means the carcass was decomposed, but organs basically intact [52]. In the research, the experimental intestines were proved to be eligible.

**Table 4.1** Stranded locations and species of sampled cetaceans.

<b>Sample ID</b>	<b>Stranded Location</b>	<b>Species Name</b>	<b>Common Name</b>	<b>Suborder</b>	<b>Body length</b>	<b>Sex</b>	<b>The stage of decomposition</b>
<b>SNH20091</b>	41.986064N	<i>Balaenoptera</i>	Minke	Mysticeti	495.0 cm	Male	CODE 2
	140.906982E	<i>acutorostrata</i>	whale				
<b>SNH19001</b>	41.868298N	<i>Lagenorhynch</i>	Pacific	Odontoceti	219.9 cm	Male	CODE 3
	140.115493E	<i>us obliquidens</i>	white-sided dolphin				
<b>SNH20032</b>	41.940520N	<i>Stenella</i>	Striped	Odontoceti	176.0 cm	Male	CODE 2
	143.242260E	<i>coeruleoalba</i>	dolphin				
<b>SNH21060</b>	42.389176N	<i>Stenella</i>	Striped	Odontoceti	191.1 cm	Female	CODE 2
	141.084797E	<i>coeruleoalba</i>	dolphin				
<b>SNH21052</b>	42.600722N	<i>Stenella</i>	Striped	Odontoceti	196.9 cm	Female	CODE 2
	141.488556E	<i>coeruleoalba</i>	dolphin				
<b>SNH22001</b>	42.010660N	<i>Lagenorhynch</i>	Pacific	Odontoceti	202.4 cm	Male	CODE 3
	140.104330E	<i>us obliquidens</i>	white-sided dolphin				
<b>SNH22006</b>	41.468139N	<i>Phocoenoides</i>	Dall's	Odontoceti	196.5 cm	Male	CODE 3
	140.029861E	<i>dalli</i>	porpoise				
<b>SNH21066</b>	42.619847N	<i>Stenella</i>	Striped	Odontoceti	215.7 cm	Male	CODE 2
	141.564035E	<i>coeruleoalba</i>	dolphin				
<b>SNH21007</b>	42.697500N	<i>Phocoena</i>	Harbour	Odontoceti	178.6 cm	Female	CODE 3
	140.051944E	<i>phocoena</i>	porpoise				
<b>SNH22016</b>	42.331027N	<i>Phocoena</i>	Harbour	Odontoceti	136.1 cm	Female	CODE 2
	141.022174E	<i>phocoena</i>	porpoise				
<b>SNH21036</b>	42.393920N	<i>Stenella</i>	Striped	Odontoceti	186.8 cm	Female	CODE 2
	140.906860E	<i>coeruleoalba</i>	dolphin				

In order to ascertain if all plastics emit fluorescent light and can thus be observed under a fluorescence microscope, a test experiment was done. Small pieces (less than 5 mm) were cut from several plastic items commonly used in household and industrial items, marked and placed in one of the intestine samples. This sample was then examined under a fluorescence microscope using the same conditions as the intestine samples used in this study. The materials used for this test were: Polyurethane (PU), Polyethylene terephthalate (PET), Poly (methyl methacrylate), i.e. acrylic plastic and plexiglass (PMMA), and Polystyrene, i.e. foams (PS). Observed microplastic fibers were characterised through their size, colour, shapes and possible origins. Furthermore, the chemical composition (polymer types) of the found fibres was identified using Raman analysis.

## 4.2.2 Methods

Raman spectroscopy is a useful tool for the analysis of microplastics. It can provide specific information about the chemical composition and structure of a sample without the need for sample preparation or destruction. This makes it a non-destructive and efficient method for identifying microplastics [34]. Another merit of Raman spectroscopy is its ability to differentiate between different types of polymers based on their unique Raman spectral signatures [35]. Based on their characteristic peaks, the spectra of polyethylene, polypropylene, and polystyrene can be easily separated. Raman spectra (RENISHAW inVia Raman Microscope) were used to identify the microplastics found. The measured spectra were uploaded to Open Specy and compared with those in the database library. Open Specy automatically displays the highest matching polymer spectra, thus allowing the identification of the studied sample.

In order to find microplastics more efficiently in the studied samples, the observation methods were divided into three types.

### **a) Digestion**

Microplastics were extracted from the intestines following Chen's procedure [36]. The surface of the intestines was first lightly washed with Milli-Q water. They were then placed into a 10% KOH solution, and a water bath was used to accelerate the digestion process. With the help of a vacuum pump, the solution was filtered through filter paper (20–25  $\mu\text{m}$ ), which was then fixed onto a flat. Labelled plastic disks were placed into a labelled Petri dish and sealed with sticky tape. An optical microscope (Nikon ECLIPSE 50i) was used to preliminary identify the particles to assess if they were plastic. Observations were made under 100x magnification. Obtained images were used to evaluate the shapes and sizes of the observed microplastics.

### **b) Direct observation**

After the intestines were washed with Mili-Q ultrapure water, they were opened and flattened with the help of clips to expose the inner wall (Figure 4.1). A digital microscope was used to search for larger plastic fragments in the sample under a magnification of 200x.



**Figure 4.1** Dissection of the intestine to expose the inner wall.

### c) Fluorescence microscope

Fluorescence microscopy was used to reduce the likelihood of missing microplastics in the studied samples. A cryostat (Leica CM 3050S) was used to slice the samples for further observation under a fluorescence microscope. Figure 4.2 shows the stages of this process.



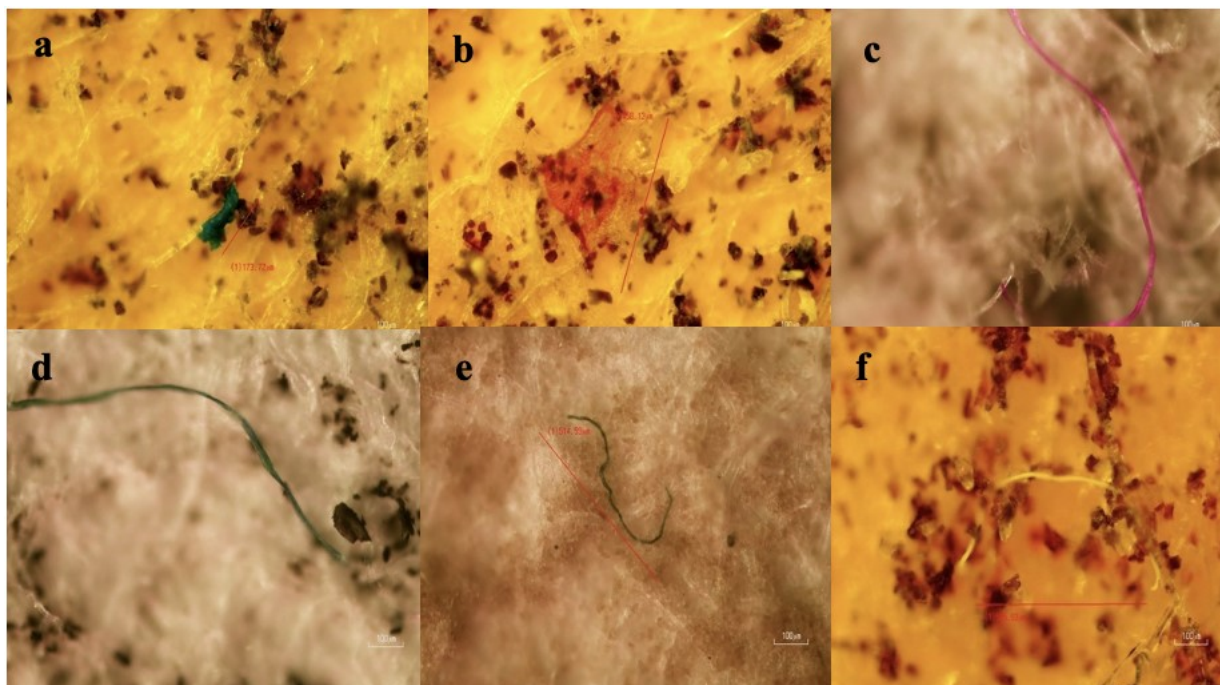
**Figure 4.2** Flow chart of the intestines' sample preparation for observation under the fluorescence microscope.

## 4.3 Results

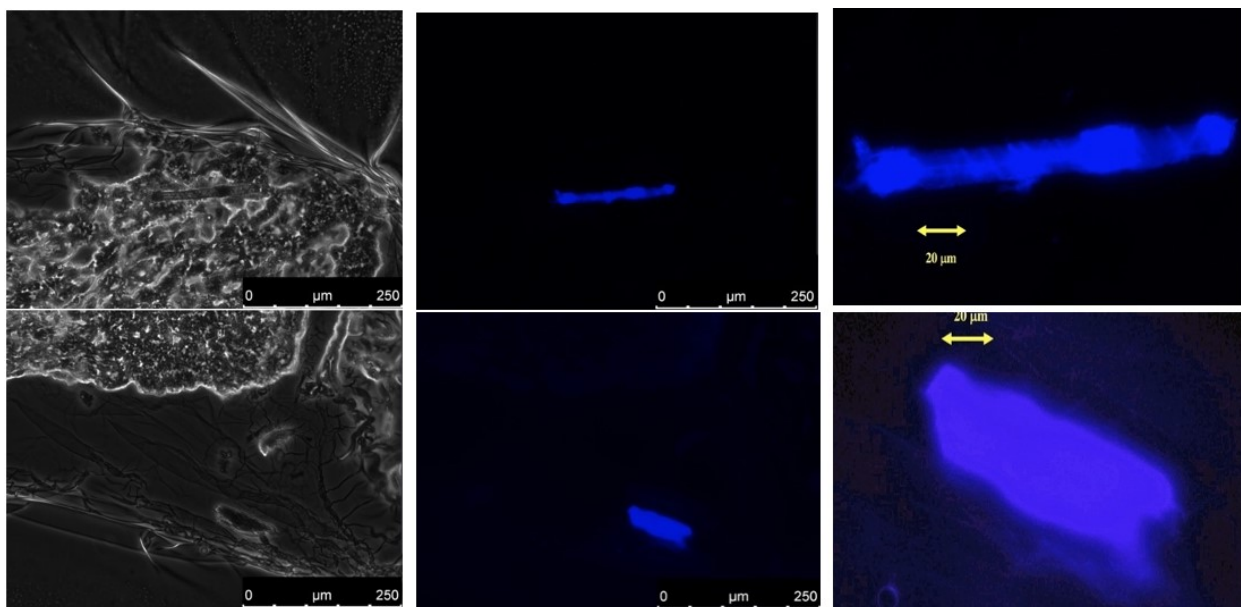
Some microplastics were found in 13 out of 17 samples of cetacean intestines (Table 4.2). The two samples without microplastics were from SNH20032 and SNH22006 (Table 4.2). Observed microplastics showed differences in length, size, and colour (Figure 4.3). Most microplastics observed were fibres (Figure 4.4). Only Nylon 6,6 was identified by Raman spectroscopy (Figure 4.5). Figure 4.6 shows the test experiment results using common household items made of different plastic polymers.

**Table 4.2** Microplastics presence in stranded cetaceans from various locations in Japan. For species names and detailed location information, refer to Table 4.1.

<b>Sample ID</b>	<b>Number of samples taken</b>	<b>Observation of microplastics in the samples</b>
<b>SNH20091</b>	2	Presence
<b>SNH19001</b>	1	Presence
<b>SNH20032</b>	2	Absence
<b>SNH21060</b>	1	Presence
<b>SNH21052</b>	1	Presence
<b>SNH22001</b>	1	Presence
<b>SNH22006</b>	1	Absence
<b>SNH21066</b>	1	Presence
<b>SNH21007</b>	5	Presence
<b>SNH22016</b>	1	Presence
<b>SNH21036</b>	1	Absence

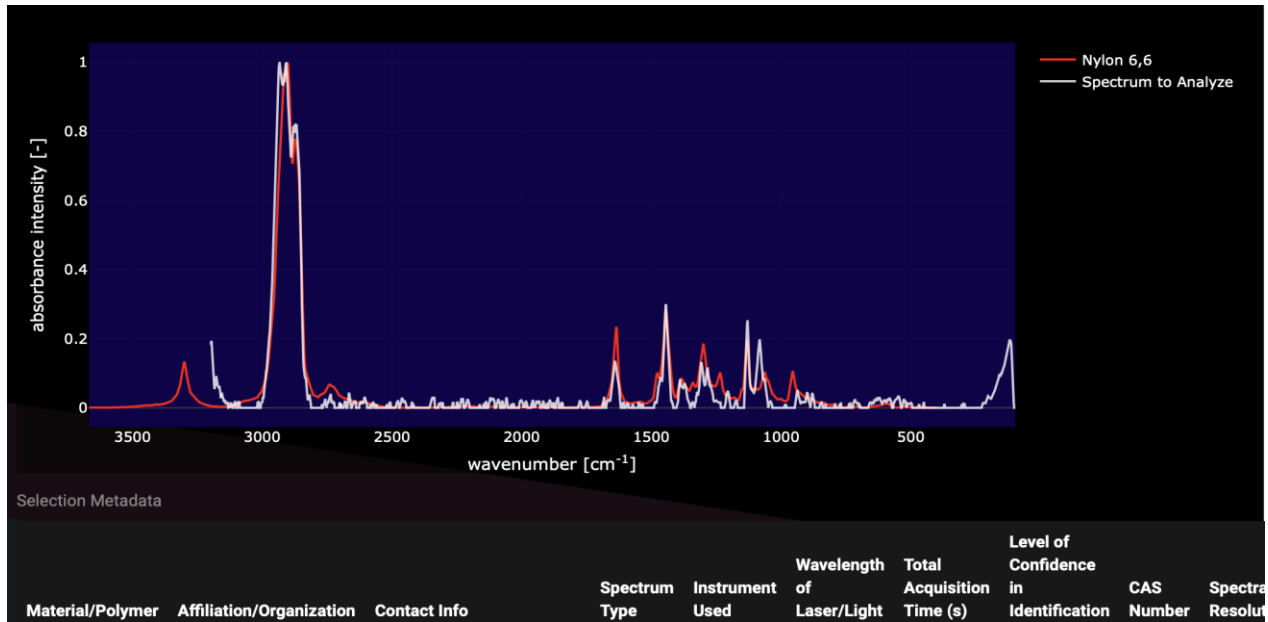


**Figure 4.3** Typical microplastics found in cetacean intestines: a) green fragment; b) red fragment; c) pink fibre; d) green fibre; e) green fibre; f) yellow fibre.

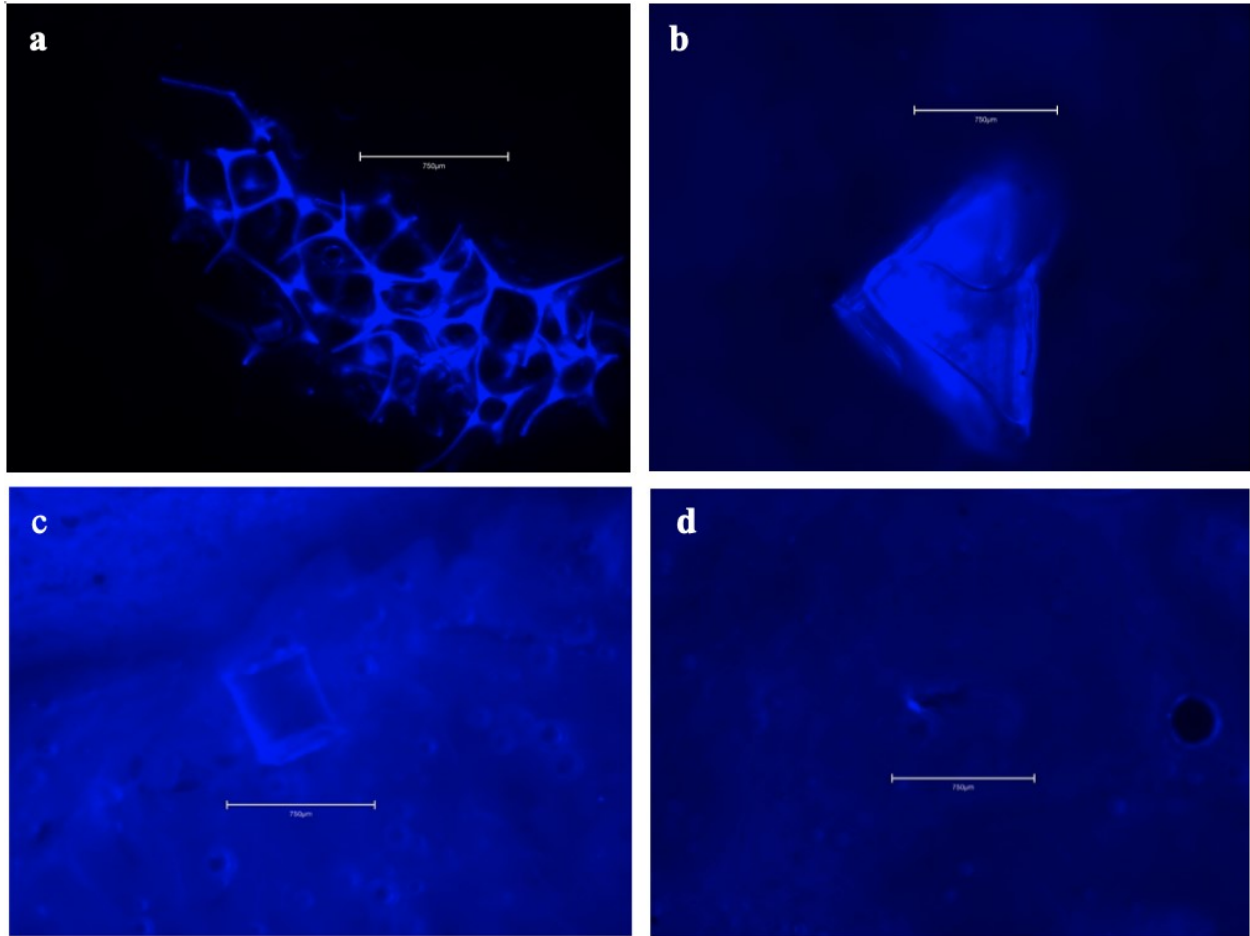


**Figure 4.4** Detail images of fragments and fibres from cetacean intestines taken with a fluorescence microscope.





**Figure 4.5** Raman spectra of Nylon 6,6.



**Figure 4.6** Images of the test done with different materials. a) PU (Polyurethane) and b) PET (Polyethylene terephthalate) are very obvious under the fluorescence microscope. c) PS (Polystyrene) and d) PMMA (Poly (methyl methacrylate)) cannot be distinguished from the surrounding tissue under the fluorescence microscope.

## 4.4 Discussion

Our study aimed to ascertain the presence of microplastics in the intestines of stranded cetaceans in several areas of Japan and their characterisation. Plastic production and its use are continuously increasing [37] despite the alarming data from many studies [38]. Indeed, plastic pollution and its consequences for marine diversity, in general, have been widely documented [6,17, 39]. Although observations of microplastic ingestion in cetaceans have increased [see 24 for a review], they are still lagging behind those of other marine groups. Cetaceans are top predators and can be well positioned as sentinel species to evaluate the health of the world's oceans [1-3]. Thus, it is crucial to increase our knowledge of the consequences of plastic pollution to develop better recommendations and guidelines for this important group of marine organisms.

As mentioned before, cetaceans' feeding habits range from top carnivore species (belonging to the toothed or Odontoceti suborder) through filter-feeding baleen species (Mysticeti suborder) [6,17]. Microplastic particles were found in most analysed samples through fluorescence and optical microscopy. The small amount found may be related to the fact that the studied material consisted mostly of toothed cetacean species. Indeed, only one baleen species was available for this study. Toothed whales are more exposed to macroplastics because of their feeding ecology [26,40]. Indeed, predictions based on observations and modelling consider that fish and cephalopod-feeding toothed cetaceans would be less impacted by microplastics when compared with krill-feeding baleen species [25]. The highest number of observed microplastics were fibres which agrees with several other studies in cetaceans and other marine vertebrates [41,28,42].

In order to compare data, the same methodology should be followed if at all possible. Unfortunately, we could not access the instruments usually used for this type of study. Consequently, it was decided to use fluorescence microscopy. The use of this method to observe

microplastics in cetacean intestine tissue with cetaceans has not been reported in previous studies, so this also had value as a methodological first trial. Even though fluorescence microscopy could clearly capture microplastic images, sample processing was long and inefficient. Therefore, only one sample was used for this trial, confirming that fluorescence microscopy can be used to obtain clear images. We would recommend this method for studies that work with smaller samples.

The small number of microplastics found could be due to several reasons. First, fluorescence microscopy may not be the most efficient method to observe microplastics in the intestines of marine mammals. Fluorescence microscopy is normally used to detect microplastics in water [43-45]. However, many microplastics have low intrinsic fluorescence signals or lack them together [46-48], making their observation quite challenging. Figure 4.6 shows that while Polyurethane and Polyethylene terephthalate has strong fluorescence emission signals, the signals are very weak and cannot be observed with fluorescence microscopes. A way to improve the fluorescence emission signals is to stain the samples before observation using fluorescent dyes such as Nile Red [49]. We wanted to check the reliability of this method in tissues. Unfortunately, our samples were very brittle after being processed for the microtome, so they could not be used further to separate the microplastics from the medium. Therefore, we relied on the digestion method to assess the presence of microplastics in the samples.

It is generally difficult to compare studies of microplastic contents in cetaceans due to methodological issues. As explained above, the use of fluorescence microscopy was chosen due to the lack of other instruments usually referenced in other reports. Unfortunately, our experiments with this method did not yield reliable results, and less-used methods were used, such as the digestion of tissues. Furthermore, some studies examine both stomach and intestines [28,42], while others use subsamples to extrapolate the total potential amount [27]. This study only used intestines

which maybe was not enough to detect microplastics. Several studies report blue and black as the most frequent microplastic colours [41]. On the contrary, this study found mostly green fibres and fragments (Figure 4.3). We also found red and yellow fragments.

This study was the first observation of Nylon 6,6 in cetaceans [50]. This robust plastic, often used for pelagic fishing nets, is considered a common source of ocean microplastics [41]. It should be observed that Nylon is a very common polymer found in ocean sediments as well as macroalgae beds [51]. More studies are necessary to assess this polymer's possible transference to the animal food chain in the oceans.

As plastic pollution increases in the oceans, we need more data from ecologically critical prey species to accurately characterise microplastics' danger. New data using different species at all trophic levels would allow a complete understanding of plastic presence in marine food webs and its consequences for apex predators, including predatory fish, seabirds, and marine mammals.

## 4.5 References

- [1] Rogers, S. I., & Greenaway, B. (2005). A UK perspective on the development of marine ecosystem indicators. *Marine pollution bulletin*, 50(1), 9-19.
- [2] Bossart, G. D. (2006). Marine mammals as sentinel species for oceans and human health. *Oceanography*, 19(2):134–137.
- [3] Guzzetti, E., Sureda, A., Tejada, S., & Faggio, C. (2018). Microplastic in marine organisms: Environmental and toxicological effects. *Environmental toxicology and pharmacology*, 64, 164-171.
- [4] Ryan, P. G. (2019). Ingestion of plastics by marine organisms. *Hazardous chemicals associated with plastics in the marine environment*, 235-266.
- [5] Fossi, M. C., Baini, M., & Simmonds, M. P. (2020). Cetaceans as ocean health indicators of marine litter impact at global scale. *Frontier Environment Science*.
- [6] International Whaling Commission. (2020). *Report of the IWC Workshop on Marine Debris: The Way Forward, 3-5 December 2019, La Garriga, Catalonia, Spain* (Vol. 68). Paper SC.
- [7] Bannister, J. L. (2009). Baleen whales (*Mysticetes*). In *Encyclopedia of marine mammals* (pp. 80-89). Academic Press.
- [8] Germanov, E. S., Marshall, A. D., Bejder, L., Fossi, M. C., & Loneragan, N. R. (2018). Microplastics: no small problem for filter-feeding megafauna. *Trends in ecology & evolution*, 33(4), 227-232.
- [9] Guerrini, F., Mari, L., & Casagrandi, R. (2019). Modelling plastics exposure for the marine biota: risk maps for fin whales in the Pelagos Sanctuary (Northwestern Mediterranean). *Frontiers in Marine Science*, 6, 299.

- [10] Burkhardt-Holm, P., & N’Guyen, A. (2019). Ingestion of microplastics by fish and other prey organisms of cetaceans, exemplified for two large baleen whale species. *Marine pollution bulletin*, 144, 224-234.
- [11] Desforges, J. P. W., Galbraith, M., & Ross, P. S. (2015). Ingestion of microplastics by zooplankton in the Northeast Pacific Ocean. *Archives of environmental contamination and toxicology*, 69(3), 320-330.
- [12] Wursig, B., & Perrin, W. F. (Eds.). (2009). *Encyclopedia of marine mammals*. Academic Press.
- [13] Au, S. Y., Lee, C. M., Weinstein, J. E., van den Hurk, P., & Klaine, S. J. (2017). Trophic transfer of microplastics in aquatic ecosystems: identifying critical research needs. *Integrated environmental assessment and management*, 13(3), 505-509.
- [14] do Sul, J. A. I., & Costa, M. F. (2014). The present and future of microplastic pollution in the marine environment. *Environmental pollution*, 185, 352-364.
- [15] Nelms, S. E., Galloway, T. S., Godley, B. J., Jarvis, D. S., & Lindeque, P. K. (2018). Investigating microplastic trophic transfer in marine top predators. *Environmental pollution*, 238, 999-1007.
- [16] Perez-Venegas, D. J., Seguel, M., Pavés, H., Pulgar, J., Urbina, M., Ahrendt, C., & Galbán-Malagón, C. (2018). First detection of plastic microfibers in a wild population of South American fur seals (*Arctocephalus australis*) in the Chilean Northern Patagonia. *Marine Pollution Bulletin*, 136, 50-54.
- [17] Zantis, L. J., Carroll, E. L., Nelms, S. E., & Bosker, T. (2021). Marine mammals and microplastics: A systematic review and call for standardisation. *Environmental Pollution*, 269, 116142.

- [18] Koelmans, A. A. (2015). Modeling the role of microplastics in bioaccumulation of organic chemicals to marine aquatic organisms. A critical review. *Marine anthropogenic litter*, 309-324.
- [19] Miller, M. E., Hamann, M., & Kroon, F. J. (2020). Bioaccumulation and biomagnification of microplastics in marine organisms: A review and meta-analysis of current data. *PLoS One*, 15(10), e0240792.
- [20] Xu, S., Ma, J., Ji, R., Pan, K., & Miao, A. J. (2020). Microplastics in aquatic environments: occurrence, accumulation, and biological effects. *Science of the Total Environment*, 703, 134699.
- [21] Garcia-Garin, O., A. Aguilar, M. Vighi, G. A. Vikingsson, V. Chosson, A. Borrell. (2021). Ingestion of synthetic particles by fin whales feeding off Western Iceland in summer. *Chemosphere* 279, 130564.
- [22] Im, J., Joo, S., Lee, Y., Kim, B. Y., & Kim, T. (2020). First record of plastic debris ingestion by a fin whale (*Balaenoptera physalus*) in the sea off East Asia. *Marine Pollution Bulletin*, 159, 111514.
- [23] Alava, J. J. (2020). Modelling the bioaccumulation and biomagnification potential of microplastics in a cetacean foodweb of the northeastern pacific: a prospective tool to assess the risk exposure to plastic particles. *Frontiers in Marine Science*, 7, 566101.
- [24] Besseling, E., Foekema, E. M., Van Franeker, J. A., Leopold, M. F., Kühn, S., Rebolledo, E. B., Heße, E., Mielke, L., IJzer, J., Kamminga, P., & Koelmans, A. A. (2015). Microplastic in a macro filter feeder: humpback whale *Megaptera novaeangliae*. *Marine pollution bulletin*, 95(1), 248-252.
- [25] Kahane-Rapport, S. R., Czapanskiy, M. F., Fahlbusch, J. A., Friedlaender, A. S., Calambokidis, J., Hazen, E. L., Goldbogen, J. A., & Savoca, M. S. (2022). Field measurements



reveal exposure risk to microplastic ingestion by filter-feeding megafauna. *Nature Communications*, 13(1), 1-11.

[26] Jacobsen, J. K., Massey, L., & Gulland, F. (2010). Fatal ingestion of floating net debris by two sperm whales (*Physeter macrocephalus*). *Marine Pollution Bulletin*, 60(5), 765-767.

[27] Moore, R. C., Loseto, L., Noel, M., Etemadifar, A., Brewster, J. D., MacPhee, S., Bendell, L., & Ross, P. S. (2020). Microplastics in beluga whales (*Delphinapterus leucas*) from the Eastern Beaufort Sea. *Marine Pollution Bulletin*, 150, 110723.

[28] Lusher, A. L., Hernandez-Milian, G., O'Brien, J., Berrow, S., O'Connor, I., & Officer, R. (2015). Microplastic and macroplastic ingestion by a deep diving, oceanic cetacean: the True's beaked whale *Mesoplodon mirus*. *Environmental pollution*, 199, 185-191.

[29] Harlacher, J. (2020). Whale, What Do We Have Here? Evidence of Microplastics in Top Predators: Analysis of Two Populations of Resident Killer Whale Fecal Samples. University of Washington.

[30] Wilcox, C., Mallos, N. J., Leonard, G. H., Rodriguez, A., & Hardesty, B. D. (2016). Using expert elicitation to estimate the impacts of plastic pollution on marine wildlife. *Marine Policy*, 65, 107-114.

[31] Pan, Z., Liu, Q., Sun, Y., Sun, X., & Lin, H. (2019). Environmental implications of microplastic pollution in the Northwestern Pacific Ocean. *Marine pollution bulletin*, 146, 215-224.

[32] Stranding Network Hokkaido (2022). Cetacean Stranding Data in Hokkaido. Version 1.9. ROIS. Sampling event dataset Access on 23-12-2022 <https://doi.org/10.15468/f9y3xd>

[33] Stranding Network Hokkaido Access on 23-12-2022 <https://kujira110.com/>

[34] Shim, W. J., Hong, S. H., & Eo, S. E. (2017). Identification methods in microplastic analysis: a review. *Analytical methods*, 9(9), 1384-1391.

- [35] Vankeirsbilck, T., Vercauteren, A., Baeyens, W., Van der Weken, G., Verpoort, F., Vergote, G., & Remon, J. P. (2002). Applications of Raman spectroscopy in pharmaceutical analysis. *TrAC trends in analytical chemistry*, 21(12), 869-877.
- [36] Chen, J. Y. S., Lee, Y. C., & Walther, B. A. (2020). Microplastic contamination of three commonly consumed seafood species from Taiwan: A pilot study. *Sustainability*, 12(22), 9543.
- [37] Rhodes, C. J. (2018). Plastic pollution and potential solutions. *Science Progress*, 101(3), 207-260.
- [38] Thushari, G. G. N., & Senevirathna, J. D. M. (2020). Plastic pollution in the marine environment. *Heliyon*, 6(8), e04709.
- [39] Dias, B. F. D. S., & Lovejoy, T. (2012). Impacts of marine debris on biodiversity: Current status and potential solutions. Secretariat of the Convention on Biological Diversity and the Scientific and Technical Advisory Panel—GEF. *Montreal, Technical Series*, (67), 61.
- [40] Unger, B., Rebolledo, E. L. B., Deaville, R., Gröne, A., IJsseldijk, L. L., Leopold, M. F., Siebert, U., Spitz, J., Wohlsein, P., & Herr, H. (2016). Large amounts of marine debris found in sperm whales stranded along the North Sea coast in early 2016. *Marine pollution bulletin*, 112(1-2), 134-141.
- [41] Nelms, S.E., J. Barnett, A. Brownlow, N. J. Davison, R. Deaville, T. S. Galloway, p. K. Lindeque, D. Santillo & B. J. Godley. (2019). Microplastics in marine mammals stranded around the British coast: ubiquitous but transitory? *Scientific Reports*, 9(1), 1-8.
- [42] Lusher, A. L., Hernandez-Milian, G., Berrow, S., Rogan, E., & 'O'Connor, I. (2018). Incidence of marine debris in cetaceans stranded and bycaught in Ireland: Recent findings and a review of historical knowledge. *Environmental Pollution*, 232, 467-476.

- [43] Dehghani, S., Moore, F., & Akhbarizadeh, R. (2017). Microplastic pollution in deposited urban dust, Tehran metropolis, Iran. *Environmental Science and Pollution Research*, 24(25), 20360-20371.
- [44] Scircle, A., & Cizdziel, J. V. (2019). Detecting and quantifying microplastics in bottled water using fluorescence microscopy: A new experiment for instrumental analysis and environmental chemistry courses. *Journal of Chemical Education*, 97(1), 234-238.
- [45] Qiu, Q., Peng, J., Yu, X., Chen, F., Wang, J., & Dong, F. (2015). Occurrence of microplastics in the coastal marine environment: first observation on sediment of China. *Marine Pollution Bulletin*, 98(1-2), 274-280.
- [46] Karakolis, E. G., Nguyen, B., You, J. B., Rochman, C. M., & Sinton, D. (2019). Fluorescent dyes for visualising microplastic particles and fibbers in laboratory-based studies. *Environmental Science & Technology Letters*, 6(6), 334-340.
- [47] Yan, J. J., Wang, Z. K., Lin, X. S., Hong, C. Y., Liang, H. J., Pan, C. Y., & You, Y. Z. (2012). Polymerising nonfluorescent monomers without incorporating any fluorescent agent produces strong fluorescent polymers. *Advanced Materials*, 24(41), 5617-5624.
- [48] Spizzichino, V., Caneve, L., Colao, F., & Ruggiero, L. (2016). Characterisation and discrimination of plastic materials using laser-induced fluorescence. *Applied spectroscopy*, 70(6), 1001-1008.
- [49] Maes, T., Jessop, R., Wellner, N., Haupt, K., & Mayes, A. G. (2017). A rapid-screening approach to detect and quantify microplastics based on fluorescent tagging with Nile Red. *Scientific reports*, 7(1), 1-10.

[50] Liu, L. C., Wu, C. A., & Chen, C. M. (2022). Modified recycled polyamide 6, 6 (r-PA 6, 6) based on pelagic fishing nets and their physical properties. *Journal of the Indian Chemical Society*, 99(6), 100450.

[51] Sukanya Hongthong, Hannah S. Leese, Michael J. Allen, Christopher J. Chuck. (2021). Assessing the Conversion of Various Nylon Polymers in the Hydrothermal Liquefaction of Macroalgae. *Environments 2021*, 8(4), 34.

[52] Geraci, J. R., & Lounsbury, V. J. (2005). *Marine mammals ashore: a field guide for strandings*. National Aquarium in Baltimore.

## Chapter 5: General conclusion

The study detected microplastics in cultured oysters from Japan, mainly in the form of fibres and fragments with red and green being the most common colours. Fibres were the most common type, with an average of  $3.14 \pm 4.00$  particles per individual. Microplastics were found in oyster gills and digestive systems, and the Raman spectra analysis showed they were nylon 6, high-density polyethylene and polypropylene. Furthermore, Microplastics were detected in 13 out of 17 samples of cetacean intestines. The observed microplastics showed variations in length, size, and colour, most of which were fibres. Raman spectroscopy analysis identified Nylon 6,6 as the sole polymer present in the microplastics.

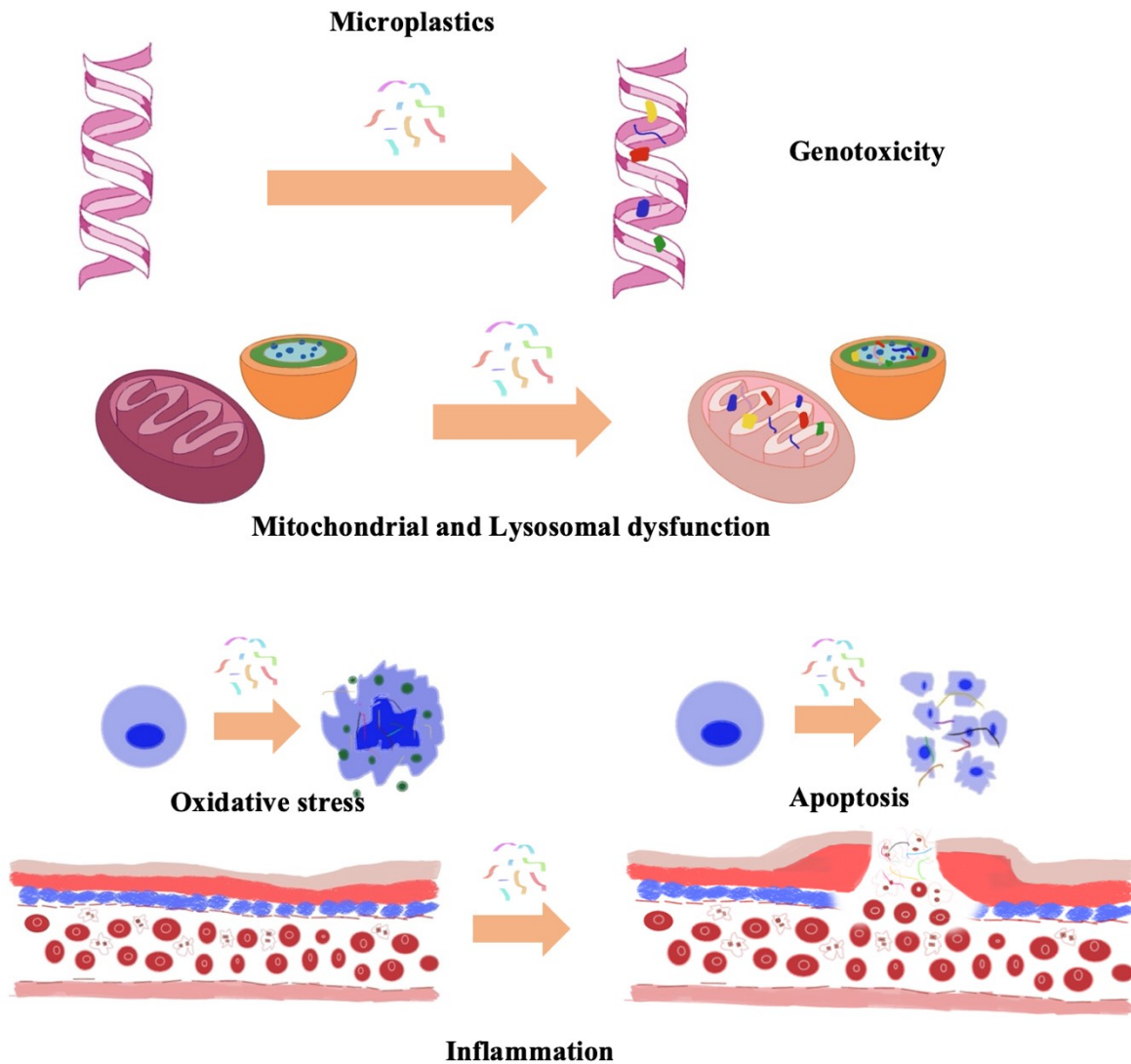
The thesis investigated the presence and accumulation of microplastics in oysters and cetaceans. The potential impacts on these species and the marine environment were discussed. The research also examined the potential relationship between microplastics and seafood safety in oysters. Additionally, the thesis analyzed the potential effects of microplastics on human health based on existing literature and discussed possible measures to reduce the impact of microplastic pollution from a political and societal perspective. The results of the study will contribute to the current knowledge of the impact of microplastics on marine life and ecosystems and can aid in developing strategies to mitigate and manage this environmental issue.

## 5.1 Toxicity of microplastic

When researchers discuss the social impact of microplastics, different aspects of their situation in life need to be considered. Studies of microplastics in various organisms were summarised. Micro pieces have been proven to be ubiquitous in the environment [1,2]. In recent years, microplastics have accumulated in oceans and sediments worldwide, reaching a maximum capacity of 100,000 particles per m<sup>3</sup> [3]. Because of their small size, microplastics can even be ingested by microorganisms [4]. However, the food chain is a more important means of transporting and storing nutrients. Plastic fragments, spheres, and fibres are already consumed by commercially available species such as marine molluscs, crustaceans, and fish. A growing body of research indicates that microplastic ingestion directly or indirectly affects marine organisms [5,6]. Researchers have found microplastics in the intestines of four kinds of seafood: cod, dab, flounder, and mackerel. These four kinds of fish are seafood with a high sales rate in the market [7]. The social impact of microplastics is also very crucial.

Regarding pathology, the toxicity of microplastics will cause oxidative stress, including disturbance of redox balance and damage to cellular components (Apoptosis) [8]. Because of microplastic, the proliferation of small vessels and infiltration of chronic inflammatory cells happened (Inflammation) [9]. Microplastics directly interfere with the normal synthesis of ATP. This phenomenon directly induces mitochondrial depolarization. Microplastics can interfere with the normal metabolism of ATP (Adenosine triphosphate) [10]. Nanoparticles (microplastics) of different sizes entered into cells. Confocal microscopy revealed that the entry route of these particles changed from direct penetration to endocytosis with increasing particle size, and then

subcellular localization changed as well. The researchers hypothesized that DNA damage and repair balance were broken [11] (Figure 5.1).



**Figure 5.1** Polystyrene microparticles show oxidative stress, apoptosis, inflammation, mitochondrial and lysosomal dysfunction, and genotoxicity in cell cultures [12].

Studies have shown that microplastics may cause inflammation, mitochondrial dysfunction, lysosomal dysfunction, and apoptosis [12]. Toxic effects depend on the dose used in the experiment, the dose rate, and the duration of exposure [13-15]. For humans, microplastics are everywhere. Studies have shown that there are also many microplastics in the air [16-18]. A study summarises data on microplastics found indoors. They also calculated the estimated daily intake of MPs from inhaling indoor dust. The adult consumes 0.23 MPs/kg BW/day [19]. The reference shows that microplastics are estimated to account for 33% of the fibres in indoor environments [20]. In the outdoors, one study found six different microplastics in Australian road dust: polypropylene, polystyrene, polyethylene terephthalate, polyvinyl chloride, poly (methyl methacrylate), and PE. Their densities range from 0.5 to 6 mg/g [21]. More microplastics are in the air than humans ingest accidentally by eating an oyster (3.14 particles per oyster). It can be inferred that microplastics in oysters do not pose a significant threat to human health.

It is difficult to determine a specific threshold at which microplastics begin to pose a threat to oyster mortality. According to the data on oysters in the current study, the average microplastic of an individual oyster is 3.14. The microplastics observed are nylon 6, nylon 6,6, polypropylene (PP), and high-density polyethylene (HDPE). The density of nylon 6,6 is 1.14 g/cm<sup>3</sup> [22], nylon 6 is 1.08 g/cm<sup>3</sup> [23], and polypropylene is 0.946 g/cm<sup>3</sup> [24]. HDPE is well known for its high strength-to-density ratio. The density of HDPE ranges from 930 to 970 kg/m<sup>3</sup> [25]. The average length of all microplastics was 593.23 µm. The average diameter of the microplastics is 10 µm. According to the data, 88% of microplastics are fibres. The average volume of microplastic can be obtained through cylindrical calculation. The average volume of the microplastic is 46,568.56 µm<sup>3</sup> (4.66 × 10<sup>-8</sup> cm<sup>3</sup>). In an old study, researchers injected nylon 6 into mice. When the injection dose was 500 mg/kg, convulsions were observed in mice. Mice died at injection doses of 800 mg/kg



(per body weight of mice) [26]. Another experiment revealed that when rats were given oral administration of over 5,000 mg/kg of HDPE, toxicity occurred, and rats died [27]. Nylon-66 has been approved as indirect food additives as polymers used for food contact surfaces [26]. Fiume et al. previously reported that adipic acid (5%) was not carcinogenic in rats fed diets [28,29]. Polypropylene is a colourless, odourless granular solid. The lethal dose for rats was determined to be 5,000 mg/kg [30]. The average weight of oysters is 29.62 g. An oyster will die if it ingests at least 0.02 g of nylon 6. The oyster would die if it consumed over 0.15 g of PP or HDPE. In table 5.1, the pollution level of microplastics observed in the oyster was much lower than the lethal level.

**Table 5.1** Comparison of the lethal dose of plastics and the weight of microplastics in oysters.

<b>Name of the microplastic</b>	<b>Density</b>	<b>Weight of microplastic</b>	<b>Lethal dose</b>
Nylon 6	1.08 g/cm <sup>3</sup>	$1.59 \times 10^{-7}$ g	0.80 g/kg
Nylon 6,6	1.14 g/cm <sup>3</sup>	$1.67 \times 10^{-7}$ g	N. A
Polyethylene	0.95 g/cm <sup>3</sup>	$4.37 \times 10^{-8}$ g	5.00 g/kg
Polypropylene	0.95 g/cm <sup>3</sup>	$4.41 \times 10^{-8}$ g	5.00 g/kg

Because the density of plastic in oysters is far below lethal levels, it is strongly suggested that the microplastic content in oysters has no impact on human health. If a 50 kg adult ingests 40 g of nylon 6 or 250 g of PE or PP, the microplastics will not reach the lethal level. In other words, if adults consume at least  $2.52 \times 10^8$  oysters amounts to LD50. In terms of quantity, humans cannot consume that many oysters.

## 5.2 Standardization of the analytic methods for microplastics

Correctly analyzing microplastics in the environment is an important aspect of understanding their impacts. A major challenge in this field is the need to standardize the analytic methods for detecting and measuring microplastics. The Japanese government (Ministry of the Environment) promulgated the River Microplastic Survey Guidelines in June 2021 [31] to provide microplastic collection, preservation, and analysis guidelines. According to the guideline, the study solutions should be filtered and then treated with 30 % H<sub>2</sub>O<sub>2</sub> solution for digestion at 50°C for 24 h. The dried microplastics should be identified using Fourier Transform Infrared Spectroscopy. At the same time, the Ministry of the Environment is encouraging a worldwide horizontal distribution mapping of microplastic densities on the ocean surface [32]. Several studies are conducting similar efforts. A meta-analysis of 68 studies found that selective, volume-reduced, and bulk sampling were the main sampling strategies, whereas density separation, filtration, sieving, and visual sorting of microplastics were commonly used to process the samples, followed by the identification and quantification of the fibres found [33]. Therefore, this research proposes the analysis method for microplastics in marine life to evaluate their influence on marine life.

While the Japanese government and previous studies may have proposed a standard procedure for digesting and observing microplastics, there are several reasons why developing a new procedure may still be necessary. Firstly, we compared the different methods for studying microplastics in organisms and water. We found that KOH digestion was the most convenient. Secondly, our experimental method is not limited to filtering microplastics from water but also allows for the extraction of microplastics from organisms. In comparison to large conventional equipment such as microscopes and Raman spectroscopy, our portable developed portable digital microscope is more convenient to carry and can be easily disseminated to the general public.

## 5.3 Current policy of plastic production

The governments of various countries have formulated related policies because of the social impact of microplastics. Some policies are extremely effective. Some policies are difficult to implement.

Public health is relatively backward in African countries. Policies to control microplastics are also being implemented [34]. First, policies to combat microplastic pollution are difficult to implement in a single sector. Policies require multiparty cooperation among all participants in the plastic value chain and scientific researchers [35]. Thirty four African countries have implemented full and partial bans on single-use plastics [36]. All Southern African Development Community members, including 16 countries, have released plastic bag ban policies [37]—the Southern African Development Community taxes plastic bags [37]. In terms of technology, it is also one of the most efficient ways for relevant departments to improve plastic waste and further recycling [38]. Despite strong government consultations, plastic producers cite potential profit losses as the reason. Businesspeople have refused to comply with the government’s plastic restrictions [39].

The Southeast Asian country (Malaysia), like some African countries, also imposes a tax on plastic bags. This policy is only enforced on Saturdays [40]. Inventing alternatives is much more welcome than charging for plastic bags [41]. In 2002 and 2005, the Indian government banned the use of plastic bags. In 2009, Hong Kong also introduced a tax on plastic bags [42].

However, the EU’s political and economic decision-making appears more effective. The EU has developed a strategy from product design to cover the entire post-consumption and end-of-life process by stimulating regional recycling capacity and ensuring demand for recycled plastics [43]. Plastic waste is more widely used as a solid recovery fuel in cement kilns than in municipal solid waste incineration [44].

In China, in response to the problem of plastic pollution, the Chinese government issued “Opinions on Further Strengthening Plastic Pollution Control” in January 2020. The government formulates laws to actively promote plastic substitutes, improve the standards of plastic substitutes, and cultivate consumers’ green consumption concepts through daily education [45]. In addition, the Chinese government recently implemented a new policy prohibiting most plastic waste imports [46]. Policies on plastics are widely applied in both developing and developed countries.

Japan has no microplastic pollution when compared to other countries. It may be because of the Japanese government’s strong environmental protection awareness. Despite implementing the plastic bag payment policy in Japan, Japan’s research on biodegradable polymers and plastics are constantly innovating [47]. Japanese researchers investigated polylactic acid and polyhydroxyalkanoates (as the most potential substitutes) [48]. Using biobased plastics produced from renewable resources and biodegradable plastics can help us solve global environmental and waste management issues [49]. Microplastics are not a serious problem in Japan because of technological advances and increased environmental awareness.

## **5.4 Recommendation**

Although countries worldwide have effectively prevented the accumulation of microplastics by enacting laws, the best way to approach microplastic pollution should be to start from the source of plastics. The government should foster a favourable scientific research environment for scientists and invest funds in research of degradable packaging materials. At the same time, several methods, such as increasing taxes and charging plastic bags, remain necessary. Consumers should be encouraged to use reusable shopping bags, while public units should strive to promote the attitude of environmental protection and raise awareness of decreasing plastic waste in school education and daily community activities. It is crucial to be aware of the potential environmental problems caused by microplastics despite the data in this study showing that the amount of microplastics is far below the lethal amount.

## 5.5 References

- [1] Wright, S. L., Thompson, R. C., & Galloway, T. S. (2013). The physical impacts of microplastics on marine organisms: A review. *Environmental Pollution*, *178*, 483-492.
- [2] Wang, J., Qin, X., Guo, J., Jia, W., Wang, Q., Zhang, M., & Huang, Y. (2020). Evidence of selective enrichment of bacterial assemblages and antibiotic resistant genes by microplastics in urban rivers. *Water Research*, *183*, 116113.
- [3] Connors, K. A., Dyer, S. D., & Belanger, S. E. (2017). Advancing the quality of environmental microplastic research. *Environmental Toxicology and Chemistry*, *36*(7), 1697-1703.
- [4] Sievert, S. M., Brinkhoff, T., Muyzer, G., Ziebis, W., & Kuever, J. (1999). Spatial heterogeneity of bacterial populations along an environmental gradient at a shallow submarine hydrothermal vent near Milos Island (Greece). *Applied and Environmental Microbiology*, *65*(9), 3834-3842.
- [5] Danopoulos, E., Jenner, L. C., Twiddy, M., & Rotchell, J. M. (2020). Microplastic contamination of seafood intended for human consumption: A systematic review and meta-analysis. *Environmental Health Perspectives*, *128*(12), 126002.
- [6] Santillo, D., Miller, K., & Johnston, P. (2017). Microplastics as contaminants in commercially important seafood species. *Integrated Environmental Assessment and Management*, *13*(3), 516-521.
- [7] Rummel, C. D., Löder, M. G., Fricke, N. F., Lang, T., Griebeler, E. M., Janke, M., & Gerdtts, G. (2016). Plastic ingestion by pelagic and demersal fish from the North Sea and Baltic Sea. *Marine Pollution Bulletin*, *102*(1), 134-141.
- [8] Solomando, A., Capó, X., Alomar, C., Álvarez, E., Compa, M., Valencia, J. M., Pinya, S., Deudero, S., & Sureda, A. (2020). Long-term exposure to microplastics induces oxidative stress



and a pro-inflammatory response in the gut of *Sparus aurata* Linnaeus, 1758. *Environmental Pollution*, 266, 115295.

[9] Li, B., Ding, Y., Cheng, X., Sheng, D., Xu, Z., Rong, Q., Wu, Y., Zhao, H., Ji, X., & Zhang, Y. (2020). Polyethylene microplastics affect the distribution of gut microbiota and inflammation development in mice. *Chemosphere*, 244, 125492.

[10] Ježek, P., & Plecítá-Hlavatá, L. (2009). Mitochondrial reticulum network dynamics in relation to oxidative stress, redox regulation, and hypoxia. *The international journal of biochemistry & cell biology*, 41(10), 1790-1804.

[11] Zhang, M., Li, J., Xing, G., He, R., Li, W., Song, Y., & Guo, H. (2011). Variation in the internalization of differently sized nanoparticles induces different DNA-damaging effects on a macrophage cell line. *Archives of toxicology*, 85(12), 1575-1588

[12] Pironti, C., Ricciardi, M., Motta, O., Miele, Y., Proto, A., & Montano, L. (2021). Microplastics in the environment: Intake through the food web, Human Exposure and Toxicological Effects. *Toxics*, 9(9), 224.

[13] Schirinzi, G. F., Pérez-Pomeda, I., Sanchís, J., Rossini, C., Farré, M., & Barceló, D. (2017). Cytotoxic effects of commonly used nanomaterials and microplastics on cerebral and epithelial human cells. *Environmental Research*, 159, 579-587.

[14] Fröhlich, E., Meindl, C., Wagner, K., Leitinger, G., & Roblegg, E. (2014). Use of whole genome expression analysis in the toxicity screening of nanoparticles. *Toxicology and Applied Pharmacology*, 280(2), 272-284.

[15] Wu, B., Wu, X., Liu, S., Wang, Z., & Chen, L. (2019). Size-dependent effects of polystyrene microplastics on cytotoxicity and efflux pump inhibition in human Caco-2 cells. *Chemosphere*, 221, 333-341.

- [16] Gasperi, J., Wright, S. L., Dris, R., Collard, F., Mandin, C., Guerrouache, M., Langlois, V., Kelly, F. J., & Tassin, B. (2018). Microplastics in air: Are we breathing it in? *Current Opinion in Environmental Science and Health*, 1, 1-5.
- [17] Zhang, Q., Xu, E. G., Li, J., Chen, Q., Ma, L., Zeng, E. Y., & Shi, H. (2020). A review of microplastics in table salt, drinking water, and air: Direct human exposure. *Environmental Science and Technology*, 54(7), 3740-3751.
- [18] Wesch, C., Elert, A. M., Wörner, M., Braun, U., Klein, R., & Paulus, M. (2017). Assuring quality in microplastic monitoring: About the value of clean-air devices as essentials for verified data. *Scientific Reports*, 7(1), 5424.
- [19] Zhu, J., Zhang, X., Liao, K., Wu, P., & Jin, H. (2022). Microplastics in dust from different indoor environments. *Science of the Total Environment*, 833, 155256.
- [20] Dris, R., Gasperi, J., Mirande, C., Mandin, C., Guerrouache, M., Langlois, V., & Tassin, B. (2017). A first overview of textile fibers, including microplastics, in indoor and outdoor environments. *Environmental Pollution*, 221, 453-458.
- [21] O'Brien, S., Okoffo, E. D., Rauert, C., O'Brien, J. W., Ribeiro, F., Burrows, S. D., Toapanta, T., Wang, X., & Thomas, K. V. (2021). Quantification of selected microplastics in Australian urban road dust. *Journal of Hazardous Materials*, 416, 125811.
- [22] Chang, H. H., Chen, S. C., Lin, D. J., & Cheng, L. P. (2013). Preparation of bi-continuous Nylon-66 porous membranes by coagulation of incipient dopes in soft non-solvent baths. *Desalination*, 313, 77-86.
- [23] Curcó, D., & Alemán, C. (2004). Performance of SuSi: A method for generating atomistic models of amorphous polymers based on a random search of energy minima. *Journal of Computational Chemistry*, 25(6), 790-798.

- [24] Shin, J. S., Park, J. M., Lee, Y. H., & Kim, H. D. (2014). Preparation and properties of eco-friendly waterborne polyurethane-urea primer for thermoplastic polypropylene applied to automobile interiors. *Clean Technology*, 20(3), 232-240.
- [25] Rodrigue, D., Kavianiboroujeni, A., & Cloutier, A. (2017, December). Determination of the optimum coupling agent content for composites based on hemp and high-density polyethylene. In *AIP Conference Proceedings* (Vol. 1914, No. 1, p. 030003). AIP Publishing LLC.
- [26] Burnett, C., Heldreth, B., Bergfeld, W. F., Belsito, D. V., Hill, R. A., Klaassen, C. D., Liebler, D. C., Marks, J. G., Shank, R. C., Slaga, T. J., Snyder, P. W., & Andersen, F. A. (2014). Safety assessment of nylon as used in cosmetics. *International Journal of Toxicology*, 33(4), 47S-60S.
- [27] MATERIAL SAFETY DATA SHEET high density polyethylene (HDPE): [https://www.opalindia.in/PDF/HDPE/Material%20Safety%20Data%20Sheet\\_HDPE.pdf](https://www.opalindia.in/PDF/HDPE/Material%20Safety%20Data%20Sheet_HDPE.pdf).
- [28] Jacobson, D. W. (2017). An industrial process for the production of nylon 6, 6 through the step-growth reaction of adipic acid and hexamethylenediamine.
- [29] Fiume, M. M., Eldreth, H., Bergfeld, W. F., Belsito, D. V., Hill, R. A., Klaassen, C. D., Liebler, D., Marks, J. G., Shank, R. C., Slaga, T. J., Snyder, P. W., & Andersen, F. A. (2012). Final report of the cosmetic ingredient review expert panel on the safety assessment of dicarboxylic acids, salts, and esters. *International Journal of Toxicology*, 31(4\_suppl), 5S-76S.
- [30] Product risk assessment: [https://www.ril.com/DownloadFiles/Polymers/assessment/prarepol\\_pp\\_homopolymer.pdf](https://www.ril.com/DownloadFiles/Polymers/assessment/prarepol_pp_homopolymer.pdf).
- [31] River Microplastic Survey Guidelines (2021). available at: <https://www.env.go.jp/content/900543325.pdf>
- [32] Michida, Y., Chavanich, S., Chiba, S., Cordova, M. R., Cozsar Cabanas, A., Glagani, F., Haggmann, P., Hinata, H., Isobe, A., Kershaw, P., Kozlovskii, N., Li, D., Lushe, A. L., Marti, E.,

Mason, S. A., Mu, J., Saito, H., Shim, W. J., Syakti, A. D., Takada, H., Thompson, R., Tokai, T., Uchida, K., Vasilenko, K., & Wang, J. (2019). Guidelines for Harmonizing Ocean Surface Microplastic Monitoring Methods. Version 1.1.

[33] Hidalgo-Ruz, V., Gutow, L., Thompson, R. C., & Thiel, M. (2012). Microplastics in the marine environment: a review of the methods used for identification and quantification. *Environmental science & technology*, 46(6), 3060-3075.

[34] Deme, G. G., Ewusi-Mensah, D., Olagbaju, O. A., Okeke, E. S., Okoye, C. O., Odii, E. C., Ejeromedoghene, O., Igun, E., Onyekwere, J. O., Oderinde, O. K., & Sanganyado, E. (2022). Macro problems from microplastics: Toward a sustainable policy framework for managing microplastic waste in Africa. *Science of the Total Environment*, 804, 150170.

[35] Worm, B., Lotze, H. K., Jubinville, I., Wilcox, C., & Jambeck, J. (2017). Plastic as a persistent marine pollutant. *Annual Review of Environment and Resources*, 42(1), 1-26.

[36] Babayemi, J. O., Nnorom, I. C., Osibanjo, O., & Weber, R. (2019). Ensuring sustainability in plastics use in Africa: Consumption, waste generation, and projections. *Environmental Sciences Europe*, 31(1), 1-20.

[37] Carlos Bezerra, J. C., Walker, T. R., Clayton, C. A., & Adam, I. (2021). Single-use plastic bag policies in the Southern African Development Community. *Environmental Challenges*, 3, 100029.

[38] Bayo, J., Rojo, D., & Olmos, S. (2019). Abundance, morphology and chemical composition of microplastics in sand and sediments from a protected coastal area: The Mar Menor lagoon (SE Spain). *Environmental Pollution*, 252(B), 1357-1366.

- [39] Behuria, P. (2021). Ban the (plastic) bag? Explaining variation in the implementation of plastic bag bans in Rwanda, Kenya and Uganda. *Environment and Planning C: Politics and Space*, 39(8), 1791-1808.
- [40] Asmuni, S., Hussin, N. B., Khalili, J. M., & Zain, Z. M. (2015). Public participation and effectiveness of the no plastic bag day program in Malaysia. *Procedia – Social and Behavioral Sciences*, 168, 328-340.
- [41] Jalil, M. A., Mian, M. N., & Rahman, M. K. (2013). Using plastic bags and its damaging impact on environment and agriculture: An alternative proposal. *International Journal of Learning and Development*, 3(4), 1-14.
- [42] Xanthos, D., & Walker, T. R. (2017). International policies to reduce plastic marine pollution from single-use plastics (plastic bags and microbeads): A review. *Marine Pollution Bulletin*, 118(1-2), 17-26.
- [43] Leal Filho, W., Saari, U., Fedoruk, M., Iital, A., Moora, H., Klöga, M., & Voronova, V. (2019). An overview of the problems posed by plastic products and the role of extended producer responsibility in Europe. *Journal of Cleaner Production*, 214, 550-558.
- [44] Lazarevic, D., Aoustin, E., Buclet, N., & Brandt, N. (2010). Plastic waste management in the context of a European recycling society: Comparing results and uncertainties in a life cycle perspective. *Resources, Conservation and Recycling*, 55(2), 246-259.
- [45] Kang, A., Ren, L., Hua, C., Dong, M., Fang, Z., & Zhu, M. (2021). Stakeholders' views towards plastic restriction policy in China: Based on text mining of media text. *Waste Management*, 136, 36-46.
- [46] Brooks, A. L., Wang, S., & Jambeck, J. R. (2018). The Chinese import ban and its impact on global plastic waste trade. *Science Advances*, 4(6), eaat0131.

[47] Lenz, R. W. (1995). Biodegradable polymers and plastics in Japan: Research, development, and applications. Japanese Technology Evaluation Center, JTEC/WTEC Program Loyola College in Maryland, Baltimore, Maryland.

[48] Sudesh, K., & Iwata, T. (2008). Sustainability of biobased and biodegradable plastics. *CLEAN – Soil, Air, Water*, 36(5-6), 433-442.

[49] Iwata, T. (2015). Biodegradable and bio-based polymers: Future prospects of eco-friendly plastics. *Angewandte Chemie*, 54(11), 3210-3215.

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