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Dicarboxylic acids and water-soluble organic carbon in aerosols in New Delhi, India in winter: Characteristics and formation processes

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Short title: DIACIDS AND WSOC IN NEW DELHI

Abstract.

Day- and night-time aerosol samples were collected at an urban site in New Delhi, India, in winter 2006-2007. They were studied for low molecular weight dicarboxylic acids and related compounds as well as total water-soluble organic carbon (TWSOC). High concentrations of diacids (up to 6.03 μg m^{-3}), TWSOC, and OC were obtained, which are substantially higher than those previously observed at other urban sites in Asia. Daytime TWSOC/OC ratio (37%) was on average higher than that in nighttime (25%). In particular, more water-soluble OC (M-WSOC) to TWSOC ratio in daytime (50%) was twice higher than in nighttime (27%), suggesting that aerosols in New Delhi are photochemically more processed in daytime to result in more water-soluble organic compounds. Oxalic acid (C_2) was found as the most abundant dicarboxylic acid, followed by succinic (C_4) and malonic (C_3) acids. Contributions of C_2 to M-WSOC were greater (av. 8%) in nighttime than daytime (av. 3%). Positive correlations of C_2 with malic acid (hC_4), glyoxylic acid (ωC_2), and relative humidity suggest that secondary production of C_2 probably in aqueous phase is important in nighttime via the oxidation of both longer-chain diacids and ωC_2. C_2 also showed a positive correlation with potassium (K^+) in nighttime, suggesting that the enhanced C_2 concentrations are associated with biomass/biofuel burning. More tight, positive correlation between less water-soluble OC (L-WSOC) and K^+ was found in both day- and night-time, suggesting that L-WSOC, characterized by longer chain and/or higher molecular weight compounds, is significantly influenced by primary emissions from biomass/biofuel burning.

1. Introduction

Water-soluble organic carbon (WSOC) can significantly alter the hygroscopicity of aerosols and is important in determining the cloud condensation nuclei (CCN) activity of particles [*Novakov and Penner*, 1993; *Saxena et al*., 1995; *Facchini et al*., 1999]. WSOC is also considered to be associated with a major fraction of secondary organic aerosols (SOA), which is formed by oxidation of volatile organic compounds (VOCs) followed by condensation on existing particles and/or nucleation. Oxalic acid (C_2), generally the most abundant dicarboxylic acid in the atmosphere [*Kawamura and Sakaguchi*, 1999], is formed from oxidation of VOCs and aqueous phase chemistry in cloud/fog droplets, as well as from primary emissions from fossil fuel combustion, and biomass burning [*Norton et al*., 1983; *Kawamura and Kaplan*, 1987; *Kawamura et al*., 1996; *Narukawa et al*., 1999; *Warneck*, 2003; *Kawamura and Yasui*, 2005; *Sorooshian et al*., 2006].

Warneck [2003] suggested that oxalic acid is produced from glyoxal in cloud droplets in the marine atmosphere, and glyoxal is formed by oxidation of acetylene and glycolaldehyde formed by oxidation of ethene. It is noted that in these processes glyoxylic acid is a key intermediate, whereas diacids, except for oxalic acid, are not produced. *Ervens et al*. [2004] also consider the production of glyoxal from toluene and of glycolaldehyde from isoprene as well as aqueous phase reactions of adipic and glutaric acids produced by the oxidation of cyclohexene. Recent studies further suggest the formation of oxalic acid by the oxidation of methylglyoxal, an oxidation product of toluene and isoprene, via intermediates such as pyruvic and acetic acids [*Lim et al*., 2005; *Carlton et al*., 2006]. *Altieri et al*. [2006] suggested that this reaction pathway also forms oligomers. The knowledge of the sources of diacids and WSOC is important to understand the formation processes of SOA that act as CCN. This information should provide important implications for regional air quality and global climate.

High loadings of carbonaceous aerosols in northern India have been reported in recent years [*Prasad et al*., 2006; *Rengarajan et al*., 2007]. In fact, the largest OC emission per unit surface area in the world occurs over India [*Ramanathan et al*., 2007]. The northern part of India has a population of more than 210 millions and thus has large anthropogenic emission sources. *Venkataraman et al*. [2005] calculated that residential biofuel combustion is the most important source of carbonaceous aerosols in India, pointing out its potential impact on climate change in the South Asian region. Carbonaceous aerosols emitted from the region can be transported downwind to Indian Ocean [e.g., *Neusüß et al*., 2002; *Ramanathan et al*., 2002; *Stone et al*., 2007] and Himalayan regions [*Prasad et al*., 2006; *Srivastava et al*., 2006]. Field observations of the Indian Ocean Experiment (INDOEX) have indicated that visible amounts of carbonaceous aerosols are present annually over the northern Indian Ocean, beginning in November and lasting until April [e.g., *Neusüß et al*., 2002; *Ramanathan et al*., 2002]. The meteorology during the period is characterized by low-level northeasterly winds with little precipitation, thus atmospheric particles are scarcely scavenged in this region [*Jayaraman et al*., 1998].

A few previous studies on aerosol organic compounds in India have been made primarily at several urban sites [e.g., *Sharma et al.*, 2003; *Chowdhury et al*., 2007]. These studies identified solvent-extractable organic compounds, including *n*-alkanes and polycyclic aromatic hydrocarbons (PAH), mainly from the viewpoint of source identification. City of New Delhi, the capital of India, is located within the Delhi metropolitan region. In spite of their importance in regional air quality and climate, organic aerosols including dicarboxylic acids and WSOC have not been characterized in New Delhi.

In this study, we conducted the first measurements of dicarboxylic acids and WSOC in the aerosols collected at an urban site of New Delhi, India, from September 2006 to April 2007.

Here, we present the temporal variability of the concentrations and the mass fractions of dicarboxylic acids in WSOC, and discuss the factors that control the molecular distributions of dicarboxylic acids and levels of WSOC in the Indian urban atmosphere.

2. Experiments

2.1. Aerosol Sampling

The sampling location in New Delhi (28.37ºN, 77.13ºE), India, is shown in Figure 1. The figure also shows the annual emission rates of anthropogenic OC and VOC over India estimated for the year 2006 [*Zhang et al*., 2009], which are given at http://www.cgrer.uiowa.edu/EMISSION_DATA_new/data/intex-b_emissions/. They include emissions from power plants, industry, residential biofuel/fossil fuel, and transportation. Anthropogenic OC and VOC are strongly emitted from the northwest to southeast of India (22°–34°N, 70°–90°E).

Total suspended particulate (TSP) samples were collected on a rooftop (15 m above the ground) of a building on a campus of National Physical Laboratory (NPL) in New Delhi from September 2006 to April 2007. The samplings were conducted using pre-combusted (at 450ºC for at least 6 h) quartz fiber filters (25 × 20 cm) and a high-volume air sampler at a flow rate of 1200 L min^{-1}. The total volume of the samples ranged between 800 and 990 m^3. The average face velocity of the TSP samplings was ~42 cm s^{-1}, which is close to the dry-deposition velocity of particles with a diameter of ~100 μm roughly calculated for the sampling conditions [*Jacobson*, 2005]. This suggests that the effect of dry deposition on the measured aerosol concentrations is insignificant for particles with diameters smaller than ~100 μm.

The sampling site is located at a residential area of New Delhi, which has less undue influence of local emissions from nearby traffic or industries [*Chowdhury et al*., 2007]. Although most of the samples were collected in wintertime (November 2006-February 2007),

three samples obtained in late summer (September) and one sample in late spring (April) were also studied here. The sampling time for each sample was approximately 12 hrs, starting at 06:00 LT and 18:00 LT, which are referred to as daytime and nighttime samples, respectively. During the course of the sampling, the local wind speed typically ranged between 2-8 m s^{-1}. The wind direction generally remained northwesterly to southwesterly with no systematic differences in daytime and nighttime. Analytical results of sixteen samples are shown in this paper.

2.2. Dicarboxylic Acids and Related Water-Soluble Compounds

The filter samples were analyzed for dicarboxylic acids (DCAs) by the method of *Kawamura and Ikushima* [1993] and *Kawamura* [1993]. A part of the quartz-fiber filter was cut in pieces and soaked in 5 ml milli-Q water, and then water-soluble species were extracted with pure water (5 ml × 3 times) under ultrasonication. The extracts were combined in a 50 ml flask after filtration with quartz wool, and concentrated to almost dryness, to which 14% BF_3/*n*-butanol was added. The extracts and reagents were then heated for 1 h to convert the carboxyl groups to butyl esters and aldehyde groups to dibutoxy acetals. The derived butyl esters and acetals were extracted with *n*-hexane. They were concentrated and dissolved in 50 μl of *n*-hexane. The butyl esters and acetals were determined by a capillary gas chromatograph (Hewlett-Packard GC6890N) with a flame ionization detector. Each compound was identified based on retention times of GC peaks with those of authentic standards and mass spectra obtained by GC/mass spectrometry.

Recoveries of authentic standards spiked on a precombusted quartz-fiber filter were 73% and 81% for oxalic and malonic acids, respectively, and greater than 95% for succinic and adipic acids. This recovery was taken into account to calculate the concentrations of the DCAs

97 in the filter samples. The variations due to analytical errors were tested by analyzing different
98 sections of the same filter sample repeatedly, and the coefficient of deviation was found to be
99 less than 8% for major diacids. The levels of field blanks were generally less than 17% of mass
100 concentrations for the major species. Concentrations of diacids and related compounds
101 presented here were corrected for the field blanks.
102
103 **2.3. Total and Fractionated WSOC and Inorganic Species**
104 In a broad definition, WSOC can be divided into two fractions, more and less soluble in
105 water. To isolate the WSOC components into more and less water-soluble fractions, we used a
106 macroporous nonionic resin (DAX-8) with TOC detection [*Duarte and Duarte*, 2005; *Sullivan*
107 *and Weber*, 2006; *Miyazaki et al*., 2009]. A number of previous studies have used the XAD-8
108 resin to isolate less water-soluble organic fractions of organic compounds, mainly humic
109 substances, from natural water samples. However, because the XAD-8 resin is no longer
110 available commercially, the comparability of a substitute resin, DAX-8, was utilized in the
111 present study. The technical specifications (e.g., pore size and surface area) and adsorption
112 characteristics of these two resins are almost the same [e.g., *Peuravuori et al*., 2001]. The
113 mechanism for the adsorption of organic solutes onto the resin is related to the molecular size
114 of the solute in a water sample and to the certain interactions between more or less
115 water-soluble organic compounds and nonionic adsorbing resin under pre-adjusted pH
116 conditions.
117 In the present WSOC analysis, the volume of DAX-8 resin packed in the column was 5 ml.
118 A filter cut of 1.54 cm^2 was extracted with ultra pure Milli-Q water using an ultrasonic bath (10
119 min × 3 times). The total extracts (100 ml) were then filtrated with a disc filter (Millex-GV,
120 0.22 μm, Millipore). Total WSOC (TWSOC) in the extracts was then determined by a total

121 organic carbon (TOC) analyzer (Sievers, 810) [*Miyazaki et al*., 2006].

122 Another aliquot of water extracts was adjusted to pH = 2 using hydrochloric acid (HCl)
123 before being pumped onto the DAX-8 resin. The DAX resin then retains hydrophobic
124 compounds in the sample solution. In this study, we define more water-soluble OC (M-WSOC)
125 as such that pass through the DAX-8 column, whereas less water-soluble OC (L-WSOC) as
126 those retained on the DAX-8 resin. L-WSOC is calculated as TWSOC – M-WSOC. The
127 variations due to analytical errors were tested by analyzing different sections of the same filter
128 sample three times. The resulting coefficients of deviation were less than 4% and 7% for
129 TWSOC and M-WSOC, respectively. The levels of field blanks were less than 5% (TWSOC)
130 and 13% (M-WSOC) of the ambient aerosol concentrations.

131 The fractions of M-WSOC and L-WSOC are operational, but they are related to the
132 carbon chain length and number of functional groups per molecule. Briefly, M-WSOC is
133 composed of highly oxygenated and/or lower molecular weight organic compounds. These
134 compounds are more hydrophilic. L-WSOC is comprised of organic compounds that contain
135 longer carbon chain and higher molecular weight structures. To characterize the DAX-8 resin
136 under the current experimental conditions, authentic organic compounds relevant to
137 atmospheric aerosol components were tested in the laboratory. This laboratory experiments
138 were made in order to determine whether each standard pass through the column. Briefly,
139 authentic water-soluble organic species were dissolved in purified water and adjusted to pH = 2
140 using HCl. They were then pumped onto the column, followed by detection by the TOC
141 analyzer.

142 Passing tests of selected organic compounds are summarized in Table 1. The results
143 showed that aliphatic dicarboxylic acids and carbonyls (<4 carbons), saccharides, and amines
144 were classified as M-WSOC, whereas aliphatic dicarboxylic acids and carbonyls (>3-4

145 carbons), aromatic acids, phenols, organic nitrates, cyclic acids, and fulvic acids were
146 classified as L-WSOC. We confirmed that M-WSOC is associated with highly water soluble
147 and/or low molecular weight organics, while L-WSOC is with less water-soluble and/or higher
148 molecular weight organics. Our results are basically same as those by XAD-8 resin [*Sullivan*
149 *and Weber*, 2006]. For C_4 dicarboxylic acids, maleic and fumaric acids are cis-trans isomers,
150 having identical molecular weights. However, the water solubility of maleic acid is about two
151 orders of magnitude higher than that of fumaric acid (Table 1). This difference with respect to
152 chemical properties can explain the result that fumaric acid was retained on the DAX resin
153 whereas maleic acid passed through the resin. On the basis of the water solubility of the
154 compounds listed in Table 1, separation between M-WSOC and L-WSOC occurs in the water
155 solubility between 3-8 g per 100 g for the compounds studied here. It should be noted that
156 aromatic compounds are included in L-WSOC, even if their molecular weights are low (e.g.,
157 catechol) and water solubility is relatively high.

158 Another aliquot of filtrated water extracts from the samples was used to determine major
159 anions and cations with Metrohm ion chromatographs (Model 761 compact IC, Metrohm,
160 Herisau, Switzerland). Anions were measured using a SI-90 4E (SHODEX) column equipped
161 with a suppressor with eluent of 1.8 mM Na_2CO_3 + 1.7 mM $NaHCO_3$. For cation analysis, we
162 used a YK-421 (SHODEX) column with 4mM H_3PO_4 as eluent.

163

164 **2.4. OC and EC**

165 Mass concentrations of OC and EC were measured using a Sunset lab EC/OC analyzer
166 (Sunset Laboratory, Inc., Tigard, OR, USA). In the present study, we used a temperature
167 protocol based on that proposed by the National Institute for Occupational Safety and Health
168 (NIOSH) [*Birch and Cary*, 1996; *Miyazaki et al*., 2007].

169 Possible interferences were assessed by the measurement of field blanks. The OC and EC
170 values for a filter punch of field blanks were 1.6 and 0.0 μgC m^{-3}, respectively, in
171 correspondence with a sample volume of 936 m^3 of air. The equivalent concentration of OC
172 accounted for 2% of the average OC concentrations of the real samples. Data of OC presented
173 here are all corrected against the field blanks. The variations due to analytical errors were
174 tested by analyzing different sections of the same filter sample three times, and the coefficient
175 of deviation was less than 6% for OC and 3% for EC. Using the measured mass concentrations
176 of OC and TWSOC, water-insoluble OC (WIOC) is defined as WIOC = OC − TWSOC.
177

178 ## 3. Results and Discussion

179 ### 3.1. Mass Concentrations and Molecular Distributions of Dicarboxylic Acids

180 Figure 2 presents temporal variations in the mass concentrations of water-soluble organic
181 components as well as some meteorological parameters from September 2006 to April 2007. In
182 general, temporal trends of TWSOC and total dicarboxylic acids (TDCA) were similar (Figure
183 2a-b), with $r^2 = 0.51$. Average concentrations of OC, WIOC, M-WSOC, and L-WSOC were
184 68.4±37.9, 47.9±31.8, 7.8±3.7, and 12.3±4.6 μgC m^{-3}, respectively. TWSOC accounted for
185 31±11% of OC, where M-WSOC accounted for 36±12% of TWSOC during the study period.
186 The TWSOC/OC ratios are within a range (20-40%) reported for other urban sites in Asia in
187 winter [*Yang et al*., 2005; *Miyazaki et al*., 2006; *Ho et al*., 2007]. Concentrations of TDCA
188 ranged between 0.84-6.03 μg m^{-3}, with an average of 2.33±1.31 μg m^{-3}. This value is 2-5 times
189 larger than those reported in Chinese cities [*Wang et al*., 2006; *Ho et al*., 2007] and urban
190 Tokyo [*Kawamura and Ikushima*, 1993]. Because these samples were all obtained at urban
191 sites in the same season (winter), the difference in the concentrations can be attributed to the
192 largest OC emission per unit surface area being found in India [*Ramanathan et al*., 2007].

193 Average molecular distributions of dicarboxylic acids, ketoacids, and α-dicarbonyls for
194 day- and night-time are shown in Figure 3. Their values are also summarized in Table 2. Oxalic
195 acid (C_2) was detected as the most abundant diacid species (1.43±0.99 μg m^{-3}), followed by
196 succinic (C_4) (0.30±0.16 μg m^{-3}) and malonic acids (C_3) (0.19±0.05 μg m^{-3}). On average,
197 concentrations of C_2-C_4 diacids accounted for 82% of TDCA. The average concentration of C_2
198 is substantially larger than that observed at other urban sites in Asia, such as Tokyo (0.27-1.35
199 μg m^{-3}) [*Kawamura and Ikushima*, 1993; *Sempere and Kawamura*, 1994], Beijing (0.22 μg
200 m^{-3}) [*Huang et al*., 2005], and Hong Kong (0.35-0.37 μg m^{-3}) [*Yao et al*., 2004]. Differences
201 between day- and night-time molecular distributions are discussed in section 3.4.

202 Concentrations of total ketoacids ranged from 0.03 to 0.64 μg m^{-3}, with an average of
203 0.16±0.09 μg m^{-3} (Table 2). Glyoxylic acid (ωC_2) (average: 0.12±0.08 μg m^{-3}) was found as a
204 dominant ketoacid, followed by 4-oxobutanoic acid (ωC_4) and pyruvic acid (Pyr). ωC_2 is an
205 intermediate in the OH oxidation of glyoxal, glycolate, methylglyoxal, and acetic acid
206 [*Kawamura et al*., 1996; *Ervens et al*., 2004], which results in oxalic acid [*Lim et al*., 2005;
207 *Carlton et al*., 2006]. The predominance of ωC_2 was also observed at the Chinese urban sites
208 [*Ho et al*., 2007]. The average concentration of ωC_2 in the New Delhi samples was
209 approximately three times greater than that observed in Chinese urban aerosols, whereas the
210 average ratio of ωC_2 to total diacids in New Delhi (0.040) was similar to that observed at the
211 Chinese urban sites (0.042). Concentrations of α-dicarbonyls ranged from 0.01 to 0.25 μg m^{-3}
212 with an average of 0.05±0.05 μg m^{-3}. These values are substantially higher than those at
213 Chinese urban sites in the winter of 2003 (av. 0.02 μg m^{-3}).

214 For inorganic species, average mass concentrations of sulfate (SO_4^{2-}) and nitrate (NO_3^-) in
215 winter were 13.6±6.1 μg m^{-3} and 18.9±4.7 μg m^{-3}, respectively (Figure 2c). These values are
216 similar to those (12.7±6.2 μg m^{-3} for SO_4^{2-} and 13.3±6.6 μg m^{-3} for NO_3^-) in winter observed at

217 Hisar (29.3ºN, 75.7ºE), an urban site ~150 km northwest of New Delhi [*Rengarajan et al*.,
218 2007]. This similarity of the concentrations suggests that the observed levels of aerosol
219 concentrations might be representative on a regional scale during the study period without any
220 significant influence from local sources.
221
222 **3.2. Mass Fractions of WSOC and Dicarboxylic Acids**
223 Figure 4 presents temporal variations in the mass fractions of the M-WSOC, L-WSOC,
224 and WIOC relative to OC. Daytime TWSOC/OC ratio was on average 37±9%, which is higher
225 than the nighttime ratio of 25±9%. The average M-WSOC/TWSOC ratio was substantially
226 greater in daytime (50±6%) than in nighttime (27±8%). TWSOC concentrations were similar
227 in daytime (av. 10.4 μgC m^{-3}) and nighttime (av. 9.6 μgC m^{-3}) (Table 2). In contrast,
228 concentration of M-WSOC was higher in daytime (9.4 μgC m^{-3}) than in nighttime (6.1 μgC
229 m^{-3}), leading to the higher mass fractions of M-WSOC in TWSOC in daytime. The higher
230 contribution of more water-soluble organic compounds in the daytime samples can be
231 explained by an enhanced photochemical production of polar compounds in daytime.
232 *Kawamura and Ikushima* [1993] suggested that the C_3/C_4 ratio can be used as an indicator
233 of enhanced photochemical production of diacids, because C_3 is produced by photochemical
234 oxidation of C_4 in the atmosphere. In the present study, C_3/C_4 ratios ranged between 0.40 and
235 1.06 with an average of 0.62, which are similar to those (0.59-0.63) reported for aerosols in
236 China in winter [*Ho et al*., 2007] and those (0.5-0.9) for winter aerosols in Tokyo [*Kawamura*
237 *and Ikushima*, 1993]. The average C_3/C_4 ratio was higher in daytime (0.66) than in nighttime
238 (0.58), being consistent with an enhanced photochemical processing as indicated by the higher
239 TWSOC/OC and M-WSOC/TWSOC ratios in daytime.
240 Figure 5 shows time series of mass fractions of C_2-C_4 in M-WSOC and relative

241 abundances of C_2-C_9 as well as C_2 relative to TDCA in day- and night-time. In contrast to the
242 diurnal difference in the M-WSOC/TWSOC ratios, the average mass fractions of C_2 in
243 M-WSOC exhibited larger contributions (av. 8%) in nighttime than in daytime (av. 3%). Much
244 higher contributions of C_2-C_4 to M-WSOC were found in nighttime largely due to the enhanced
245 concentrations of C_2 and decreased M-WSOC concentrations in nighttime. In fact, oxalic acid
246 (C_2) was found to account for 70±7% of total C_2-C_9 diacids in nighttime, which is greater than
247 that in daytime (59±9%). It is interesting to note that the average C_2 concentration in nighttime
248 (1.91 μg m^{-3}) is almost twice higher than in daytime (0.96 μg m^{-3}) (Figure 3). Increased relative
249 abundance of C_2 in nighttime suggests a possible formation of C_2 in nighttime, a point to be
250 discussed in section 3.4.

251 Figure 6 shows relations between total C_2-C_4 concentrations and M-WSOC contents in
252 day- and night-time. A linear relation was found between C_2-C_4 diacid concentrations and
253 M-WSOC both in daytime (r^2 = 0.94) and nighttime (r^2 = 0.81). This suggests that sources for
254 C_2-C_4 and M-WSOC (and their precursors) were similar. However, M-WSOC concentrations
255 are significantly higher (1.5 times) in daytime than in nighttime, whereas daytime
256 concentrations of C_2-C_4 diacids are lower by ~60% than in nighttime (see Figure 6). These
257 results suggest that more water-soluble organic compounds, except for C_2-C_4 diacids,
258 contribute more significantly to M-WSOC in daytime, and are likely produced by
259 photochemical processes. Photochemical production of C_2-C_4 diacids can also occur in
260 daytime, as indicated by the higher C_3/C_4 ratio in daytime.

261 Azelaic acid (C_9) has been proposed as oxidation product of unsaturated fatty acids.
262 Unsaturated fatty acids are abundant in terrestrial plant leaves and contain a double bond
263 predominantly at C-9 position [*Kawamura and Gagosian*, 1987; *Kawamura and Kaplan*,
264 1987]. In fact, C_9 was generally found as the fourth most abundant diacid species in both day-

and night-time samples, with average concentrations of 0.070 μg m^{-3} and 0.072 μg m^{-3}, respectively. These C_9 concentrations account for 4% and 3% of C_2-C_9 diacids in day- and night-time, respectively. Although it is difficult to quantitatively determine the contribution of biogenic emissions to the observed aerosols from our data alone, the result suggests that biogenic emissions of unsaturated fatty acids contribute to the New Delhi aerosols to some extent.

3.3 Contribution of Biomass/Biofuel Burning Emissions

It has been well recognized that the emissions from biomass/biofuel burning include a large portion of water soluble organic species [e.g., *Mayol-Bracero et al*., 2002]. With regard to the molecular distributions of diacids shown in Figure 3, a similar distribution of diacids dominated by C_2, followed by C_4, was also reported for biomass burning plumes [e.g., *Narukawa et al*., 1999; *Legrand et al*., 2007]. *Chowdhury et al*. [2007] inferred that 20% of $PM_{2.5}$ aerosol sampled in Delhi in winter was attributable to primary emissions from biomass burning using various molecular markers (e.g., levoglucosan). To investigate the effects of biomass/biofuel burning on the sources of observed organic compounds, water soluble potassium (K^+) is used as a tracer of wood burning [e.g., *Andreae*, 1983]. Scatter plots of C_2, M-WSOC, and L-WSOC with K^+ are shown in Figure 7. C_2 and M-WSOC showed moderate positive correlations with K^+ in nighttime (r^2 = 0.34-0.43), suggesting that emissions from biomass/biofuel burning are apparently an important source for C_2 and M-WSOC in nighttime at the sampling site. More pronounced linear relation was observed between L-WSOC and K^+ in both daytime (r^2 = 0.70) and nighttime (r^2 = 0.75). These tight, positive correlations suggest that majority of L-WSOC originated from biomass/biofuel combustion.

Gao et al. [2003] used the data set of C_2-C_5 diacids in wood burning on savanna fires,

showing an increase in the C_2 to K^+ ratio, ranging from 0.02 over the fire to 0.25 after 40 min of aging. This C_2 formation during the initial aging of aerosols was attributed to fast secondary production in the smoke plumes [*Gao et al*., 2003]. Even higher enrichment of C_2 with respect to K^+ was observed for the nighttime samples of the present study with an average C_2/K^+ ratio of 0.61±0.35, indicating a fast secondary production of C_2 in the burning plume. On the other hand, an average C_2/K^+ ratio in daytime was 0.30±0.15, indicating less significant production of C_2 in the burning plume compared to that in nighttime (a point to be discussed in the following section).

Contribution of wood burning to secondary formation of diacids is also supported by a positive correlation between C_9 and K^+ in the nighttime samples ($r^2 = 0.28$). Unsaturated fatty acids such as oleic acid mainly emitted by wood burning and meat cooking [*Rogge et al*., 1991] can be the source of diacids. Oleic acid, once emitted, is rapidly oxidized to result in C_9 and other products, which can be further decomposed to C_5 and C_4. The moderate correlation between C_9 and K^+ in the nighttime samples suggests a certain contribution of wood burning to C_9.

3.4 Potential Processes for the Net Production of Oxalic Acid

As shown in section 3.2, relative increases in the concentrations and fractions of C_2 were observed for the nighttime samples. One might argue that descending of boundary layer (BL) height in nighttime could accumulate air pollutants within the BL, leading to the enhancement of the mass concentrations of C_2 and other species in the atmosphere near the ground surface in nighttime. However, the difference in the C_2 mass fractions between day- and night-time cannot be explained by the changes in the BL heights. Although time-resolved measurements of C_2 precursors (i.e., VOCs) are not available, no significant difference was found between the

313 day- and night-time concentrations of primary emission tracers such as EC and K^+. In fact, the
314 average concentrations of EC (40.4 μgC m^{-3}) and K^+ (3.55 μg m^{-3}) in daytime were similar to
315 those in nighttime (EC: 36.6 μgC m^{-3}, K^+: 3.65 μg m^{-3}). This suggests that the difference in the
316 C_2 mass fractions has been unlikely due to differences in the amount of precursor emissions in
317 day- and night-time. Rather, the relative increases in the concentrations and fractions of C_2 in
318 nighttime can be explained by the differences in net production processes of C_2 between day
319 and night: (1) more significant production of C_2 and/or (2) less significant loss of C_2 in
320 nighttime than in daytime.

321 Previous studies have reported that low molecular weight dicarboxylic acids were
322 sometimes found in the coarse mode, such as sea salt, dust, and soil particles [*Neusüss et al*.,
323 2000; *Mochida et al*., 2003; *Huang et al*., 2006]. C_2 and other acidic species in the gas phase
324 could be adsorbed onto alkaline coarse particles [e.g., *Neusüss et al*., 2000]. In this study,
325 however, concentration of sodium (Na^+) was considerably low with no significant difference
326 between day and night. Moreover, a correlation between C_2 and Na^+ was poor ($r^2 < 0.09$) for
327 both day- and night-time samples, suggesting that the production of C_2 associated with sea salt
328 particles is unlikely in the New Delhi samples studied. Similarly, C_2 also showed poor
329 correlation with calcium (Ca^{2+}) ($r^2 < 0.09$), indicating an insignificant contribution of dust
330 particles to C_2 production.

331 One possible explanation for the increases in the relative concentrations and fractions of
332 C_2 in nighttime is a secondary formation of C_2 in aqueous phase of aerosols. In aqueous phase,
333 C_2 is known to be formed by OH oxidation of various precursors including longer-chain
334 dicarboxylic acids and glyoxylic acid (ωC_2). *Kawamura and Ikushima* [1993] proposed that
335 malic acid (hC_4), hydroxylated dicarboxylic acid, can be formed by hydration of maleic acid
336 and/or hydroxylation of succinic acid, and is further oxidized to produce C_2. Evidence for

337 hydroxylation is based on detection of C_5 and C_6 hydoxyacids [*Appel et al.*, 1979]. Since no
338 direct emissions have been identified for hC_4, the change in the relative abundances of C_4 and
339 hC_4 may indicate a progress of photochemical aging of aerosols probably in aqueous phase.
340 Figure 8 shows concentrations of C_2 and its contribution to total diacids (C_2/TDCA) as a
341 function of $hC_4/(hC_4+C_4)$ in day- and night-time. Although hC_4 showed lower concentrations
342 than C_4, their relative abundance fluctuated significantly. The C_2 concentrations and
343 $hC_4/(hC_4+C_4)$ ratios showed a strong correlation (r^2 = 0.86) for the nighttime samples.
344 Similarly, the C_2/TDCA ratios also showed an increase with increasing ratios of $hC_4/(hC_4+C_4)$,
345 with a moderate correlation (r^2 = 0.21) in nighttime. Furthermore, C_2 was well correlated with
346 C_3 (r^2 = 0.66), C_4 (r^2 = 0.82), and glutaric acid (C_5; r^2 = 0.80). These results provide further
347 evidence for aqueous phase secondary production of C_2 in aerosols via decays of longer-chain
348 dicarboxylic acids, rather than primary emission. However, such a correlation between C_2 and
349 $hC_4/(hC_4+C_4)$ was not found for the daytime samples. This may suggest that heterogeneous
350 reactions to produce C_2 in aerosol phase are not important in daytime because relative humidity
351 was low (av. ~55%, see Figure 2d), possibly leading to lower water contents of aerosols in
352 daytime. On the other hand, higher relative humidity (av. ~70%) in nighttime is favorable for
353 aqueous phase chemistry in ambient aerosols. In fact, time-resolved meteorological data
354 showed that the maximum relative humidity reached nearly 100% during most of the nighttime
355 samplings (Figure 2d), supporting an occurrence of fogs or clouds near the ground surface. On
356 the other hand, the relative humidity in daytime never reached 100%. Moreover, a positive
357 correlation between C_2/TDCA and the average relative humidity (r^2 =0.49) was found for all
358 the data sets, which also suggests aqueous phase production of C_2 in aerosols.
359 The above results support the possibility of aqueous phase productions of C_2 via decays of
360 longer-chain dicarboxylic acids, but do not exclude an importance of other pathways such as

361 production of C_2 via the oxidation of ωC_2. *Sorooshian et al*. [2007] conducted aircraft
362 measurements and found that relative abundance of ωC_2 in total organic acids detected is
363 higher in cloud droplets than in the aerosol samples collected below clouds. Their result
364 suggests the formation of ωC_2 in an early stage of liquid phase reaction. In the present study,
365 average concentrations of ωC_2 were as high as 0.11 and 0.13 μg m^{-3} in day- and night-time,
366 respectively, which correspond to 7-12% of the C_2 mass concentrations. Concentrations of C_2
367 and ωC_2 highly correlate ($r^2 = 0.65$) in nighttime, whereas the correlation is low ($r^2 = 0.35$) in
368 daytime (data not shown as a figure). Pyr has also been reported to act as a precursor of C_2 via
369 ωC_2 in aqueous phase [*Lim et al*., 2005; *Carlton et al*., 2006] and shown to originate from
370 aromatic hydrocarbons and isoprene [*Talbot et al*., 1995]. In fact, Pyr and ωC_2 were positively
371 correlated ($r^2 = 0.53$) in our nighttime samples. These results further support the secondary
372 production of C_2 in nighttime. It is noted that a $\omega C_2/C_2$ ratio in nighttime (av. 0.067) is lower
373 than that in daytime (av. 0.119), possibly due to the production of C_2 from ωC_2 in nighttime.
374 This process may proceed within a time scale of the aerosol sampling (~12 h), leading to the
375 lower $\omega C_2/C_2$ ratio in nighttime.

376 It has been shown that ωC_2 produces C_2 more efficiently at higher pH, because the rate
377 constant of the oxidation of anion (glyoxylate) is an order of magnitude greater than that of ωC_2
378 [*Ervens et al*., 2003a]. In the present study, however, no significant difference in the imbalance
379 between anions and cations was seen for the day- and night-time samples (data not shown). The
380 ammonium (NH_4^+)-to-SO_4^{2-} molar ratios were 4.3 and 4.9 in day- and night-time, respectively.
381 Although these ratios indicate that there was sufficient NH_4^+ to neutralize SO_4^{2-}, there was no
382 significant difference in the NH_4^+/SO_4^{2-} molar ratios between day- and night-time. These
383 results suggest insignificant effects of pH on the enhanced C_2 levels in nighttime. Hence,
384 secondary production of C_2 likely takes place in aqueous phase, in which oxidation of both

longer-chain diacids and ωC_2 are important sources of C_2 in the nighttime atmosphere in New Delhi.

A correlation between C_2 and sulfate and its linear regression slope have been used to investigate production processes of C_2 mainly via aqueous phase reactions [e.g., *Yu et al*, 2005], because secondary production pathways of sulfate are well established. Figure 9 shows a scatter plot between C_2 and sulfate in daytime and nighttime. The C_2 and sulfate concentrations showed a strong correlation for the daytime ($r^2 = 0.67$) and nighttime samples ($r^2 = 0.65$). Previous studies have noted similarity in size distributions of these two species, suggesting a common source [*Yao et al*., 2003 *Crahan et al*., 2004; *Huang et al*., 2006]. Although the size distributions of C_2 and sulfate are not available and the sample number is limited in the present study, a linear regression slope of the C_2 to sulfate ratio in the nighttime samples (0.198) is about 60% higher than in the daytime samples (0.125). The difference in the slopes together with higher concentrations of C_2 in nighttime suggests more efficient production of C_2 in nighttime, which can be a consequence of multistep aqueous-phase production of C_2: several subsequent oxidation steps are necessary to form C_2 from VOC precursors, whereas fewer steps are required for production of sulfate [*Warneck*, 2003; *Ervens et al*., 2004; *Yu et al*., 2005]. The *x*-intercept of the linear relationship between C_2 and SO_4^{2-} found in day- and night-time suggests that an additional source of SO_4^{2-} exists in addition to the different formation and loss processes of both species. One possible explanation for the intercept is that New Delhi and surrounding regions have more power plants and coal combustion [*Prasad et al*., 2006; *Zhang et al*., 2009], which produce much greater emissions of SO_2 but emit fewer VOCs.

In addition to the production processes, various loss processes of C_2 could affect the concentrations and fractions of C_2. C_2 is lost by its oxidation to result in CO_2 [e.g., *Warneck*, 2003]. *Zuo and Holgné* [1992] suggested that photolysis of iron(III)/iron(II)−oxalato

409 complexes can be an important sink of C_2 in the aqueous phase in the atmosphere. *Ervens et al*.
410 [2003b] used a photochemical box model and showed that photolysis of the
411 iron-dioxalato-complex is much more effective than oxidation of C_2 by OH. On the other hand,
412 evaporation of C_2 is negligible because the vapor pressure of C_2 is sufficiently low ($<10^{-4}$ mm
413 Hg) [*Saxena and Hildemann*, 1996]. It is possible that rates of production and loss of C_2 are
414 different for daytime and nighttime. During nighttime, oxidation of organics may be less
415 efficient due to the absence of sunlight and lack of H_2O_2 and OH production, and NO_3 may act
416 as a major oxidant. In summary, the relative increases in the concentrations and fractions of C_2
417 in nighttime are probably due to more significant production of C_2 and/or less significant loss
418 of C_2 than in daytime.
419

420 **4. Conclusions**

421 Measurements of water-soluble dicarboxylic acids and bulk water soluble organic carbon
422 (WSOC) were made for the first time in ambient aerosol samples collected at an urban site in
423 New Delhi, India, in winter 2006-2007. A macroporous nonionic (DAX-8) resin was used to
424 quantify more- and less-WSOC, which are defined as the fractions of WSOC that passed
425 through and retained on the DAX-8 column, respectively. On average, total WSOC
426 (TWSOC)/OC ratio was found to be significantly higher in daytime (37%) than in nighttime
427 (25%). In particular, more water-soluble OC (M-WSOC) accounted for 50% of TWSOC in
428 daytime, which is substantially higher than that (27%) in nighttime. These differences can be
429 interpreted by more pronounced photochemical production of polar organic compounds in
430 daytime even in winter.

431 Average concentrations of oxalic acid (C_2) (1.43 μg m^{-3}) and total diacids (2.33 μg m^{-3})
432 were substantially high compared to those previously reported at other urban sites in Asia. C_2

433 was the most abundant dicarboxylic acid, followed by succinic (C_4) and malonic (C_3) acid. We
434 found that the average mass fractions of C_2 in nighttime were greater (8% of M-WSOC and
435 66% of total diacids) than daytime (3% of M-WSOC and 54% of total diacids). Positive
436 correlations of C_2 with malic acid (hC_4), glyoxylic acid (ωC_2), and relative humidity suggest
437 that secondary production of C_2 occurs possibly in aqueous aerosol phase via the oxidation of
438 longer-chain diacids and ωC_2 in the New Delhi atmosphere during nighttime. C_2 also showed a
439 positive correlation with potassium in nighttime, suggesting that the enhanced net production
440 of C_2 is associated with fast secondary reactions in biomass/biofuel burning plumes.

441 Less water soluble OC (L-WSOC), composed of longer carbon chain and higher
442 molecular weight organic compounds, was highly correlated with potassium in both day- and
443 night-time samples, demonstrating that majority of L-WSOC was affected by primary
444 emissions from biomass/biofuel burning.

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References

Altieri, K. E., A. G. Carlton, H.-J. Lim, B. J. Turpin, and S. P. Seitzinger (2006), Evidence for oligomer formation in clouds: reactions of isoprene oxidation products, *Environ. Sci. Technol*, 40. 4956-4960.

Andreae, M. O. (1983), Soot carbon and excess fine potassium: Long-range transport of combustion-derived aerosols, *Science*, 220, 1148–1151.

Appel, B. R., E. M. Hoffer, E. L. Kothny, S. M. Wall, M. Haik, and R. L. Knights (1979), Analysis of carbonaceous material in Southern-California atmospheric aerosols. 2., *Environ. Sci. Technol*., 13, 98–104.

Birch, M. E., and R. A. Cary (1996), Elemental carbon-based method for monitoring occupational exposures to particulate diesel exhaust, *Aerosol Sci. Technol*., *25*, 221– 241.

Carlton, A. G., B. J. Turpin, H. J. Lim, K. E. Altieri, and S. Seitzinger (2006), Link between isoprene and secondary organic aerosol (SOA): Pyruvic acid oxidation yields low volatility organic acids in clouds, *Geophys. Res. Lett.*, 33, L06822, doi:10.1029/2005GL025374.

Chowdhury, Z., M. Zheng, J. J. Schauer, R. J. Sheesley, L. G. Salmon, G. R. Cass, and A. G. Russell (2007), Speciation of ambient fine organic carbon particles and source apportionment of PM2.5 in Indian cities, *J. Geophys. Res*., *112*, D15303, doi:10.1029/2007JD008386.

Crahan, K. K., D. Hegg, D. S. Covert, and H. Jonsson (2004), An exploration of aqueous oxalic acid production in the coastal marine atmosphere, *Atmos. Environ.*, *38*, 3757–3764.

Duarte, R. M. B., and Duarte, A. C. (2005), Application of non-ionic solid sorbents (XAD Resins) for the isolation and fractionation of water-soluble organic compounds from atmospheric aerosols, *J. Atmos. Chem*., 51, 79-93.

Ervens, B., S. Gligorovski, and H.Herrmann (2003a), Temperature-dependent rate constants for hydroxyl radical reactions with organic compounds in aqueous phase, *Phys. Chem. Chem. Phys*., 5, 1811–1824.

Ervens, B., et al. (2003b), CAPRAM 2.4 (MODAC mechanism): An extended and condensed tropospheric aqueous phase mechanism and its application, *J. Geophys. Res.*, 108(D14), 4426, doi:10.1029/2002JD002202.

Ervens, B., G. Feingold, G. J. Frost, and S. M. Kreidenweis (2004), A modeling study of aqueous production of dicarboxylic acids: 1. Chemical pathways and speciated organic mass production, *J. Geophys. Res*., 109, D15205, doi:10.1029/2003JD004387.

Facchini, M. C., M. Mircea, S. Fuzzi, and R. J. Charlson (1999), Cloud albedo enhancement by surface-active organic solutes in growing droplets, *Nature*, 401, 257– 259.

Gao, S., D. A. Hegg, P. V. Hobbs, T. W. Kirchstetter, B. I. Magi, and M. Sadilek (2003), Water-soluble organic components in aerosols associated with savanna fires in southern

Africa: Identification, evolution, and distribution, *J. Geophys. Res*., 108(D13), 8491, doi:10.1029/2002JD002324.

Ho, K. F., J. J. Cao, S. C. Lee, K. Kawamura, R.-J. Zhang, J.-C. Chow, and J.-G. Watson (2007), Dicarboxylic acids, ketocarboxylic acids and dicarbonyls in urban atmosphere of China, *J. Geophys. Res.*, 112, D22S27, doi:10.1029/2006JD008011.

Huang, X. F., M. Hu, L. Y. He, and X. Y. Tang (2005), Chemical characterization of water-soluble organic acids in PM2.5 in Beijing, China, *Atmos. Environ.*, *39*, 2819–2827.

Huang, X.-F., J. Z. Yu, L.-Y. He, and Z. Yuan (2006), Water-soluble organic carbon and oxalate in aerosols at a coastal urban site in China: Size distribution characteristics, sources, and formation mechanisms, *J. Geophys. Res.*, *111*, D22212, doi:10.1029/2006JD007408.

Jacobson, M. Z. (2005), Fundamentals of Atmospheric Modeling, Second Edition, Cambridge University Press, New York, pp. 813.

Jayaraman, A., D. Lubin, S. Ramachandran, V. Ramanathan, E. Woodbridge, W. D. Collins, and K. S. Zalpuri (1998), Direct observations of aerosol radiative forcing over the tropical Indian Ocean during the January-February 1996 pre-INDOEX cruise, *J. Geophys. Res.*, *103*(D12), 13,827–13,836.

Kawamura, K., and R. B. Gagosian (1987), Implications of ω-oxocarboxylic acids in the remote marine atmosphere for photo-oxidation of unsaturated fatty acids, *Nature*, 325, 330–332.

Kawamura, K., and I. R. Kaplan (1987), Motor-exhaust emissions as a primary source for dicarboxylic acids in Los Angeles, *Environ. Sci. and Technol*. 21, 105-110.

Kawamura, K. (1993), Identification of C_2-C_{10} ω-oxocarboxylic acids, pyruvic acid, and C_2-C_3 a-dicarbonyls in wet precipitation and aerosol samples by capillary GC and GC/MS, *Anal. Chem.*, 65, 3505– 3511.

Kawamura, K., and K. Ikushima (1993), Seasonal changes in the distribution of dicarboxylic acids in the urban atmosphere. *Environ. Sci. Technol. 27*, 2227-2235.

Kawamura, K., H. Kasukabe, and L. A. Barrie (1996), Source and reaction pathways of dicarboxylic acids, ketoacids and dicarbonyls in Arctic aerosols: One year of observations, *Atmos. Environ*., 30, 1709-1722.

Kawamura, K., and F. Sakaguchi (1999), Molecular distributions of water-soluble dicarboxylic acids in marine aerosols over the Pacific Ocean including tropics, *J. Geophys. Res*., 104(D3), 3501-3509.

Kawamura, K., and O. Yasui (2005), Diurnal changes in the distribution of dicarboxylic acids, ketocarboxylic acids and dicarbonyls in the urban atmosphere, *Atmos. Environ*, 39, 1945-1960.

Legrand, M., S. Preunkert, T. Oliveira, C. A. Pio, S. Hammer, A. Gelencse´r, A. Kasper-Giebl,

and P. Laj (2007), Origin of C_2–C_5 dicarboxylic acids in the European atmosphere inferred from year-round aerosol study conducted at a west-east transect, *J. Geophys. Res*., 112, D23S07, doi:10.1029/2006JD008019.

Lim, H. J., A. G. Carlton, and B. J. Turpin (2005), Isoprene forms secondary organic aerosol through cloud processing: Model simulations, *Environ. Sci. Technol*., 39, 4441–4446.

Mayol-Bracero, O. L., et al. (2002), Water-soluble organic compounds in biomass burning aerosols over Amazonia 2. Apportionment of the chemical composition and importance of the polyacidic fraction, *J. Geophys. Res*., 107(D20), 8091, doi:10.1029/2001JD000522.

Miyazaki, Y., Y. Kondo, N. Takegawa, Y. Komazaki, K. Kawamura, M. Mochida, K. Okuzawa, and R. J. Weber (2006), Time-resolved measurements of water-soluble organic carbon in Tokyo, *J. Geophys. Res*., *111*, D23206, doi:10.1029/2006JD007125.

Miyazaki, Y., Y. Kondo, S. Han, M. Koike, D. Kodama, Y. Komazaki, H. Tanimoto, and H. Matsueda (2007), Chemical characteristics of water-soluble organic carbon in the Asian outflow, *J. Geophys. Res.*, *112*, D13209, doi:10.1029/2006JD007896.

Miyazaki, Y., et al. (2009), Chemical characterization of water-soluble organic carbon aerosols at a rural site in the Pearl River Delta, China, in the summer of 2006, *J. Geophys. Res.*, doi:10.1029/2009JD011736, *114*, D14208, doi:10.1029/2009JD011736.

Mochida, M., N. Umemoto, K. Kawamura, and M. Uematsu, Bimodal size distribution of C_2–C_4 dicarboxylic acids in the marine aerosols (2003), *Geophys. Res. Lett*., *30*(13), 1672, doi:10.1029/2003GL017451.

Narukawa, K. Kawamura, N. Takeuchi and T. Nakajima (1999), Distribution of dicarboxylic acids and carbon isotopic compositions in aerosols from 1997 Indonesian forest fires. *Geophys. Res. Lett.* 26, 3101-3104.

Neusüss, C., M. Pelzing, A. Plewka, and H. Herrmann (2000), A new analytical approach for size-resolved speciation of organic compounds in atmospheric aerosol particles: Method and first results, *J. Geophys. Res.*, *105*, 4512–4527.

Neusüß, C., T. Gnauk, A. Plewka, H. Herrmann, and P. K. Quinn (2002), Carbonaceous aerosol over the Indian Ocean: OC/EC fractions and selected specifications from size-segregated onboard samples, *J. Geophys. Res.*, *107*(D19), 8031, doi:10.1029/2001JD000327.

Norton, R. B., J. M. Roberts, and B. J. Huebert, Tropospheric oxalate, *Geophys. Res. Lett*., 10, 517–520, 1983.

Novakov, T., and J. E. Penner (1993), Large contribution of organic aerosols to cloud-condensation-nuclei concentrations, *Nature*, 365, 823–826.

Peuravuori, J., P. Ingman, K. Pihlaja, and R. Koivikko (2001), Comparisons of aquatic humic matter by DAX -8 and XAD-8 resins from solid-state 13C NMR spectroscopy's point of view, *Talanta*, *55*, 733-742.

Prasad, A. K., R. P. Singh, and M. Kafatos (2006), Influence of coal based thermal power plants on aerosol optical properties in the Indo-Gangetic basin, *Geophys. Res. Lett.*, *33*, L05805, doi:10.1029/2005GL023801.

Ramanathan, V., et al. (2002), The Indian Ocean Experiment and the Asian brown cloud, *Curr. Sci.*, *83(8)*, 947-955.

Ramanathan, V., et al. (2007), Atmospheric brown clouds: Hemispherical and regional variations in long-range transport, absorption, and radiative forcing, *J. Geophys. Res.*, *112*, D22S21, doi:10.1029/2006JD008124.

Rengarajan, R., M. M. Sarin, and A. K. Sudheer (2007), Carbonaceous and inorganic species in atmospheric aerosols during wintertime over urban and high-altitude sites in North India, *J. Geophys. Res*., *112*, D21307, doi:10.1029/2006JD008150.

Rogge, W. F., L. M. Hildemann, M. A. Mazurek, G. R. Cass, and B. R. T. Simoneit (1991), Sources of fine organic aerosol, 1, Charbroilers and meat cooking operations, *Environ. Sci. Technol*., 25, 1112–1125.

Saxena, P., L. M. Hildemann, P. H. McMurry, and J. H. Seinfeld (1995), Organics alter hygroscopic behavior of atmospheric particles, *J. Geophys. Res*., *100*, 18,755– 18,770.

Saxena, P., and L. M. Hildemann (1996), Water-soluble organics in atmospheric particles: A critical review of the literature and application of thermodynamics to identify candidate compounds, *J. Atmos. Chem*., *24*, 57– 109.

Sempere, R., and K. Kawamura (1994), Comparative distributions of dicarboxylic-acids and related polar compounds in snow rain and aerosols from urban atmosphere, *Atmos. Environ.*, *28*, 449–459.

Sharma, D. N., A. A. Sawant, R. Uma, and D. R. Cocker III (2003), Preliminary chemical characterization of particle-phase organic compounds in New Delhi, India, *Atmos. Environ.*, *37*, 4317–4323.

Sorooshian, A., et al. (2006), Oxalic acid in clear and cloudy atmospheres: Analysis of data from International Consortium for Atmospheric Research on Transport and Transformation 2004, *J. Geophys. Res*., 111, D23S45, doi:10.1029/2005JD006880.

Sorooshian, A., N. L. Ng, A. W. H. Chan, G. Feingold, R. C. Flagan, and J. H. Seinfeld (2007), Particulate organic acids and overall water-soluble aerosol composition measurements from the 2006 Gulf of Mexico Atmospheric Composition and Climate Study (GoMACCS), *J. Geophys. Res.*, *112*, D13201, doi:10.1029/2007JD008537.

Srivastava, M. K., S. Singh, A. Saha, U. C. Dumka, P. Hegde, R. Singh, and P. Pant (2006), Direct solar ultraviolet irradiance over Nainital, India, in the central Himalayas for clear-sky day conditions during December 2004, *J. Geophys. Res*., *111*, D08201, doi:10.1029/2005JD006141.

Stone, E. A., G. C. Lough, J. J. Schauer, P. S. Praveen, C. E. Corrigan, and V. Ramanathan (2007), Understanding the origin of black carbon in the atmospheric brown cloud over the Indian Ocean, *J. Geophys. Res*., *112*, D22S23, doi:10.1029/2006JD008118.

Sullivan, A. P., and R. J. Weber (2006), Chemical characterization of the ambient organic aerosol soluble in water: 1. Isolation of hydrophobic and hydrophilic fractions with a XAD-8 resin, *J. Geophys. Res.*, *111*, D05314, doi:10.1029/2005JD006485.

Talbot, R. W., B. W. Mosher, B. G. Heikes, D. J. Jacob, J. W. Munger, B. C. Daube, W. C. Keene, J. R. Maben, and R. S. Artz (1995), Carboxylic-acids in the rural continental atmosphere over the eastern United States during the Shenandoah Cloud and Photochemistry Experiment, *J. Geophys. Res*., 100, 9335–9343.

Venkataraman, C., G. Habib, A. Eiguren-Fernandez, A. H. Miguel, and S. K. Friedlander (2005), Residential biofuels in South Asia: Carbonaceous aerosol emissions and climate impacts, *Science*, *307*, 1454-1456.

Wang, G., K. Kawamura, T. Watanabe, S. Lee, K. Ho, and J. Cao (2006), High loadings and source strengths of organic aerosols in China, *Geophys. Res. Lett*., 33, L22801, doi:10.1029/2006GL027624.

Warneck, P. (2003), In-cloud chemistry opens pathway to the formation of oxalic acid in the marine atmosphere, *Atmos. Environ*., *37*, 2423–2427.

Yang, H., J. Z. Yu, S. S. H. Ho, J. Xu, W.-S. Wu, C. H. Wan, X. Wang, X. Wang, and L. Wang (2005), The chemical composition of inorganic and carbonaceous materials in PM2.5 in Nanjing, China, *Atmos. Environ*., *39*, 3735–3749.

Yao, X. H., A. P. S. Lau, M. Fang, C. K. Chan, and M. Hu (2003), Size distributions and formation of ionic species in atmospheric particulate pollutants in Beijing, China: 2-dicarboxylic acids, *Atmos. Environ*., 37, 3001–3007.

Yao, X. H., M. Fang, C. K. Chan, K. F. Ho, and S. C. Lee (2004), Characterization of dicarboxylic acids in PM2.5 in Hong Kong, *Atmos. Environ.*, *38*, 963–970.

Yu, J. Z., S. F. Huang, J. H. Xu, and M. Hu (2005), When aerosol sulfate goes up, so does oxalate: Implication for the formation mechanisms of oxalate, *Environ. Sci. Technol*., 39, 128–133.

Zhang, Q., D. G. Streets, G. R. Carmichael, K. He, H. Huo, A. Kannari, Z. Klimont, I. Park, S. Reddy, D. Chen, L. Duan, Y. Lei, L. Wang, and Z. Yao (2009), Asian emissions in 2006 for the NASA INTEX-B mission, *Atmos. Chem. Phys. Discuss*., 9, 4081-4139.

Zuo, Y., and J. Holgné (1992), Formation of hydrogen peroxide and depletion of oxalic acid in atmospheric water by photolysis of iron (III)−oxalato complexes, *Environ. Sci. Technol*., 26, 1014–1022.

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Figure Captions

Figure 1. A map of India with the location of New Delhi indicated by asterisk. Anthropogenic emission estimates of OC and VOC are also shown for year 2006 over India [*Zhang et al*., 2009], the data of which are available at the website http://www.cgrer.uiowa.edu/ EMISSION_DATA_new/data/ intex-b_emissions/.

Figure 2. Time series of speciated and identified organic aerosols, major inorganic aerosols collected in New Delhi from September 2006 to April 2007 together with meteorological data: (a) M-WSOC, L-WSOC, and WIOC, (b) total diacids, oxalic acid (C_2), malonic acid (C_3), and succinic acid (C_4), (c) sulfate and nitrate, (d) the maximum and average relative humidity (RH), and ambient temperature. "D" indicates data corrected in daytime, whereas "N" indicates that in nighttime.

Figure 3. Average molecular distributions of dicarboxylic acids, ketoacids, and α-dicarbonyls in daytime and nighttime. To highlight the concentrations of less abundant species, molecular distributions except C_2-C_4 are also shown with an enhanced y-axis scale.

Figure 4. Mass fractions of M-WSOC, L-WSOC, and WIOC in OC during daytime and nighttime.

Figure 5. Mass fractions of C_2-C_4 in M-WSOC, C_2 in total diacids, and relative abundances of C_2-C_9 in day- and night-time. To highlight the abundance of organic compounds shown, the y-axis begins at 40% in the bottom panels as C_2 contributes >40% in each category.

Figure 6. Relation between total C_2-C_4 and M-WSOC concentrations during daytime (open circle) and nighttime (solid circle).

Figure 7. Scatter plots of (a) C_2, (b) M-WSOC and (c) L-WSOC with K^+ in daytime (open circle) and nighttime (solid circle). The correlation coefficient is based on a regression line

that is not forced through 0.

Figure 8. Variations of (a) C_2 concentrations and (b) C_2/TDCA ratios as a function of $hC_4/(hC_4+C_4)$ in daytime (open circle) and nighttime (solid circle).

Figure 9. Relation between C_2 and sulfate concentrations during daytime (open circle) and nighttime (solid circle). “S” indicates a slope of a linear regression.

Table 1. Summary of the laboratory experiments for the DAX-8 resin using organic species standards. The classification as M-WSOC or L-WSOC for each compound is indicated by circles, based on their passing efficiency through the DAX-8 resin. For definition of M-WSOC and L-WSOC, see text.

Functional group	Compounds	MW,[a] g mol^{-1}	Solubility in water[b] (g solute per 100g water)	Concentrations, μgC L^{-1}	Passing Efficiency, %	Definition in this study	
						M-WSOC	L-WSOC
Dicarboxylic Acid	Oxalic acid,C_2	90	12	50, 260, 430	100	○	
	Succinic acid, C_4	118	8.8	380	100	○	
	Maleic acid, C_4	116	80.4	405	100	○	
	Fumaric acid, C_4	116	0.63	470	0		○
	Adipic acid, C_6	146	2.5	120	0		○
	Azelaic acid, C_9	188	0.21	280	0		○
Carbonyls	Glyoxal	58	≥ 67	120	100	○	
	Methyl glyoxal	72	≥ 67	90	100	○	
Amines	Ethanolamine	61	Infinity[c]	160	100	○	
Saccharides	Levoglucosan	162	N/A[d]	330	100	○	
	Sucrose	342	211.5	80, 290	100	○	
Phenols	Catechol	110	43	220	0		○
Aromatic acids	Phthalic acid	166	0.7	200	0		○
Cyclic acids	Pinic acid	186	N/A[d]	110	0		○
	cis-Pinonic acid	184	N/A[d]	320	0		○
Humic-like	Suwannee River fulvic	---	N/A[c]	70, 250, 520	0		○

[a] 'MW' denotes molecular weight.

[b] The values are those summarized in *Saxena and Hildemann* [1996].

[c] 'Infinity' denotes complete miscibility with water.

[d] 'N/A' denotes that solubility data is not readily available.

Table 2. Summary of the average concentrations of OC, WSOC, diacids, ketoacids, and α-dicarbonyls in the aerosol samples collected in New Delhi, India[a]

	Groups/Compounds/ Abbreviations	Concentrations	
		Daytime	Nighttime
OC ($\mu gC\ m^{-3}$)		60.1 (16.0-110.4)	76.7 (31.8-159.5)
TWSOC ($\mu gC\ m^{-3}$)	M-WSOC	9.4 (3.5-14.7)	6.1 (1.6-10.2)
	L-WSOC	11.4 (5.3-16.8)	13.1 (6.5-45.2)
	Total	20.8 (8.8-31.5)	19.2 (8.4-53.8)
Diacids ($ng\ m^{-3}$)	Oxalic acid, C_2	955 (396-1619)	1906 (517-4531)
	Malonic, C_3	181 (85-254)	194(101-242)
	Succinic acid, C_4	273 (150-636)	330 (165-605)
	Maleic acid, M	18 (6-77)	36(6-182)
	Fumaric acid, F	9 (3-46)	10 (3-19)
	Glutaric, C_5	51 (28-122)	66 (25-144)
	Adipic acid, C_6	29 (17-69)	32 (14-46)
	Pimelic, C_7	15 (6-37)	16 (5-28)
	Suberic, C_8	5 (1-11)	5 (1-11)
	Azelaic acid, C_9	70 (26-168)	72 (44-115)
	Sebacic, C_{10}	8 (3-23)	10 (4-15)
	Methylmalonic, iC_4	9 (4-12)	12 (6-23)
	Methylsuccinic, iC_5	39 (9-102)	53 (17-95)
	Methylglutaric, iC_6	7 (3-21)	8.5 (3-13)
	Methylmaleic, mM	21 (4-82)	24 (8-38)
	Phthalic, Ph	40 (22-95)	45 (14-76)
	Iso-phthalic, iPh	8 (3-19)	11 (4-24)
	Tere-phthalic, tPh	5 (3-9)	4 (1-11)
	Malic, hC_4	9 (3-30)	9(4-17)
	Sub total	1777 (835-2835)	2875 (979-6030)
Ketoacids ($ng\ m^{-3}$)	Pyruvic, Pyr	18 (1-80)	9 (1-32)
	Glyoxylic, ωC_2	114 (27-508)	128 (19-245)
	3-Oxopropanoic, ωC_3	5 (3-15)	6 (2-17)
	4-Oxobutanoic, ωC_4	19 (11-35)	15 (9-23)
	9-Oxononanoic, ωC_9	1 (0.4-4)	2 (0.4-7)
	Sub total	157 (42-642)	161 (31-324)
α-Dicarbonyls ($ng\ m^{-3}$)	Glyoxal, Gly	12 (1-41)	24 (4-72)
	Methylglyoxal, mGly	28 (7-68)	31(9-178)
	Sub total	40 (8-109)	55 (13-250)

[a] Numbers in parenthesis in the concentration rows are ranges of the concentrations.

Figure 1

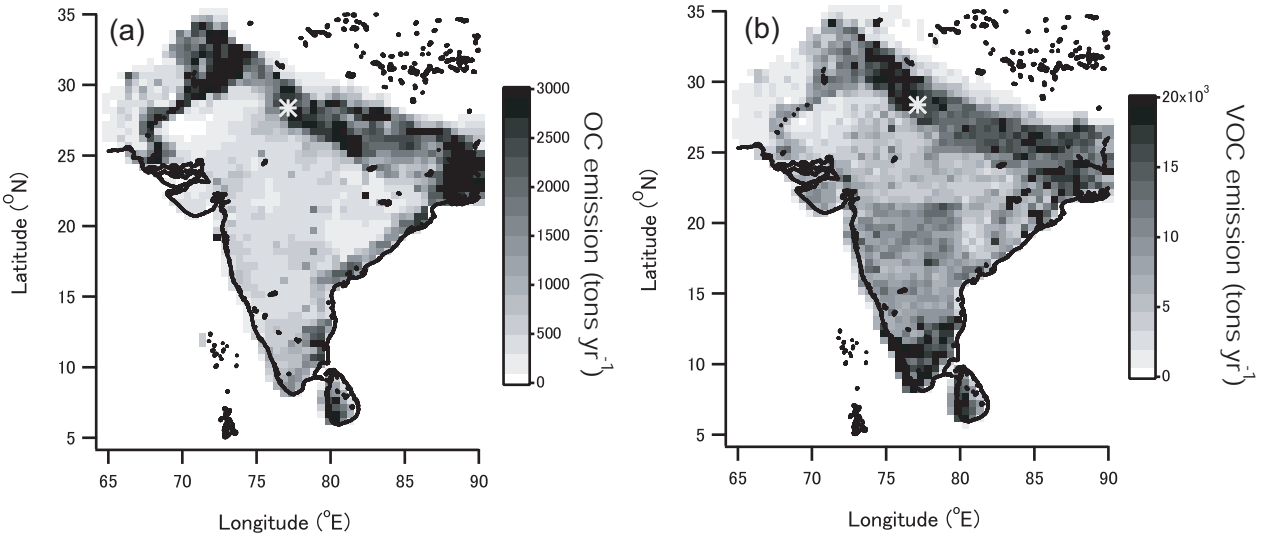

(a)
(b)
Latitude (°N)
Longitude (°E)
OC emission (tons yr-1)
VOC emission (tons yr-1)
35
30
25
20
15
10
5
65
70
75
80
85
90
3000
2500
2000
1500
1000
500
0
20x10^3
15
10
5
0

Figure 2

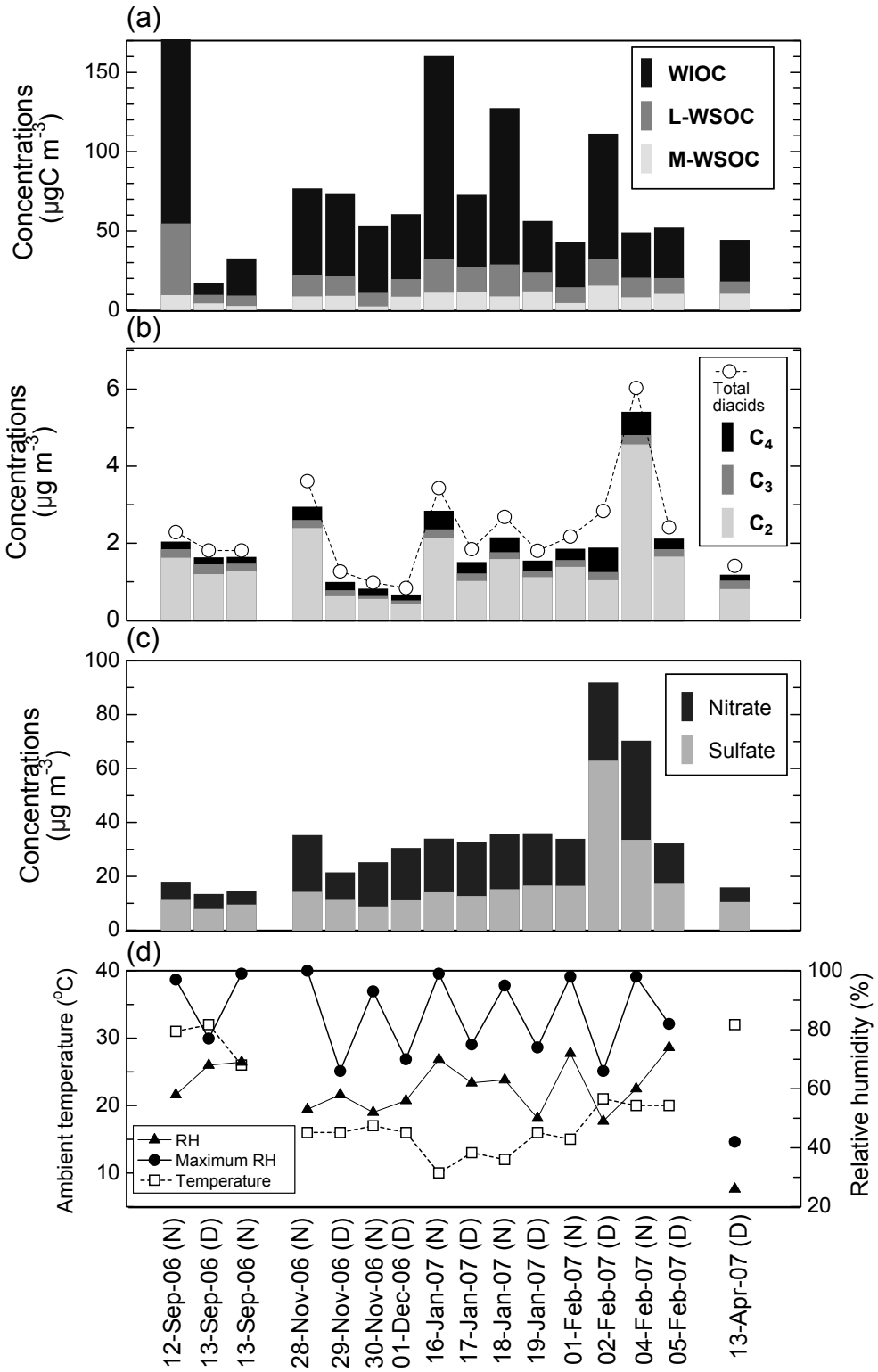

(a)
Concentrations (μgC m-3)
WIOC
L-WSOC
M-WSOC
(b)
Concentrations (μg m-3)
Total diacids
C4
C3
C2
(c)
Concentrations (μg m-3)
Nitrate
Sulfate
(d)
Ambient temperature (°C)
Relative humidity (%)
RH
Maximum RH
Temperature
12-Sep-06 (N)
13-Sep-06 (D)
13-Sep-06 (N)
28-Nov-06 (N)
29-Nov-06 (D)
30-Nov-06 (N)
01-Dec-06 (D)
16-Jan-07 (N)
17-Jan-07 (D)
18-Jan-07 (N)
19-Jan-07 (D)
01-Feb-07 (N)
02-Feb-07 (D)
04-Feb-07 (N)
05-Feb-07 (D)
13-Apr-07 (D)

Figure 3

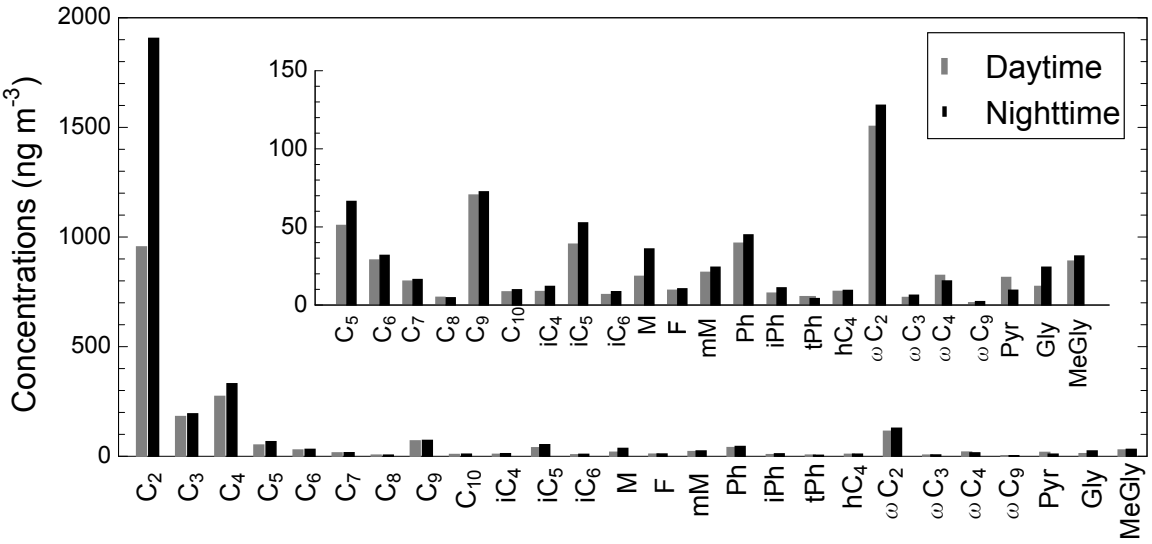

Concentrations (ng m-3)
2000
1500
1000
500
0
150
100
50
0
Daytime
Nighttime
C2
C3
C4
C5
C6
C7
C8
C9
C10
iC4
iC5
iC6
M
F
mM
Ph
iPh
tPh
hC4
ω C2
ω C3
ω C4
ω C9
Pyr
Gly
MeGly

Figure 4

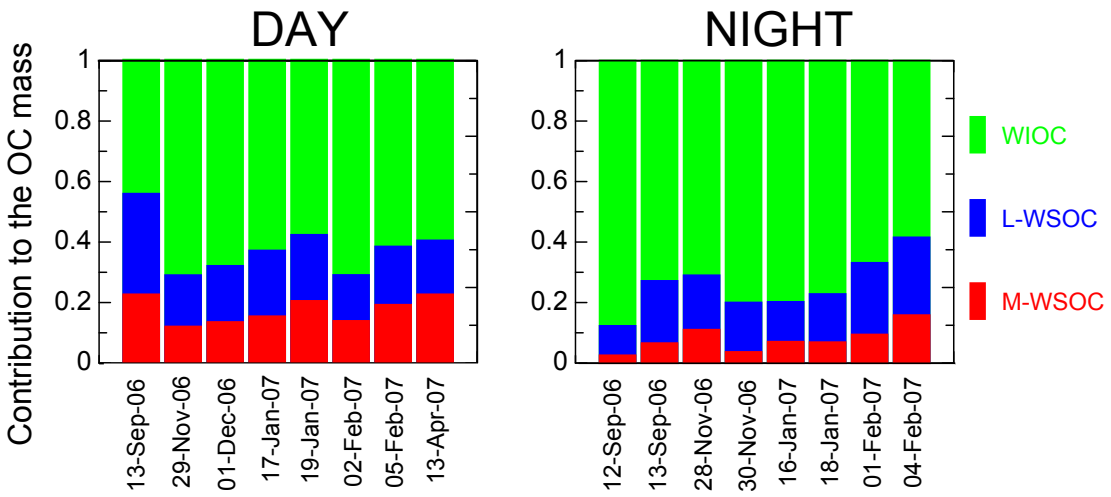

DAY
NIGHT
Contribution to the OC mass
1
0.8
0.6
0.4
0.2
0
13-Sep-06
29-Nov-06
01-Dec-06
17-Jan-07
19-Jan-07
02-Feb-07
05-Feb-07
13-Apr-07
12-Sep-06
13-Sep-06
28-Nov-06
30-Nov-06
16-Jan-07
18-Jan-07
01-Feb-07
04-Feb-07
WIOC
L-WSOC
M-WSOC

Figure 5

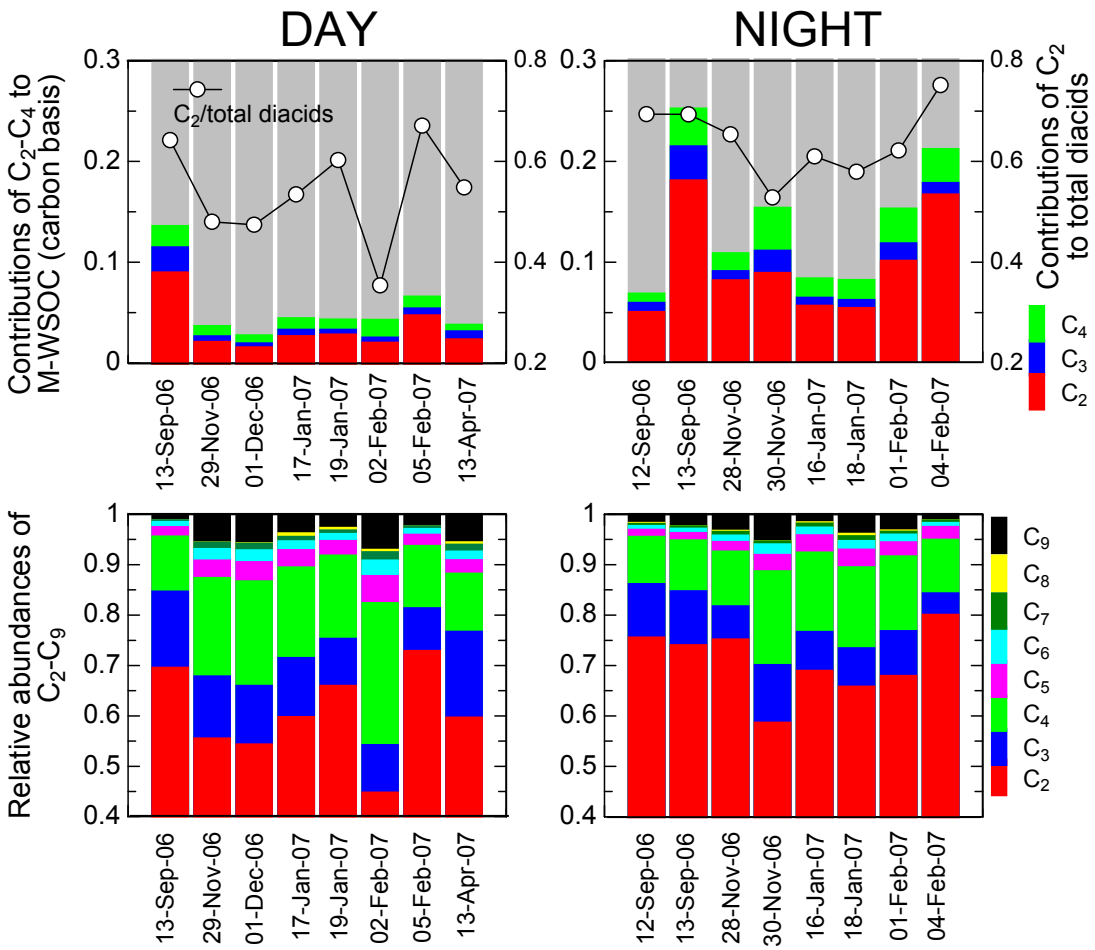

DAY
NIGHT
Contributions of C2-C4 to M-WSOC (carbon basis)
C2/total diacids
Contributions of C2 to total diacids
C4
C3
C2
Relative abundances of C2-C9
C9
C8
C7
C6
C5
C4
C3
C2
13-Sep-06
29-Nov-06
01-Dec-06
17-Jan-07
19-Jan-07
02-Feb-07
05-Feb-07
13-Apr-07
12-Sep-06
13-Sep-06
28-Nov-06
30-Nov-06
16-Jan-07
18-Jan-07
01-Feb-07
04-Feb-07

Figure 6

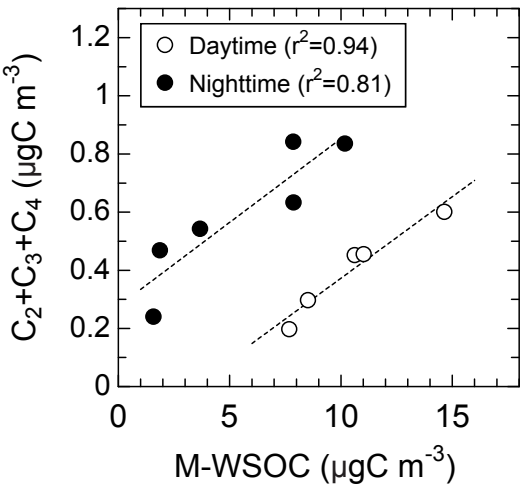

Daytime (r^2=0.94)
Nighttime (r^2=0.81)
$C_2+C_3+C_4$ (µgC m^{-3})
M-WSOC (µgC m^{-3})
0
0.2
0.4
0.6
0.8
1
1.2
0
5
10
15

Figure 7

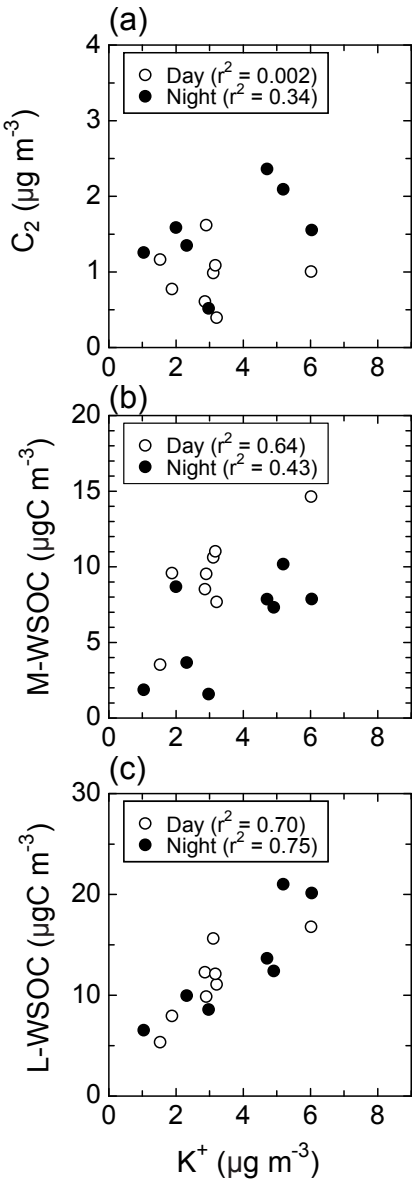

(a)
Day (r^2 = 0.002)
Night (r^2 = 0.34)
C_2 (μg m^{-3})
(b)
Day (r^2 = 0.64)
Night (r^2 = 0.43)
M-WSOC (μgC m^{-3})
(c)
Day (r^2 = 0.70)
Night (r^2 = 0.75)
L-WSOC (μgC m^{-3})
K^+ (μg m^{-3})

Figure 8

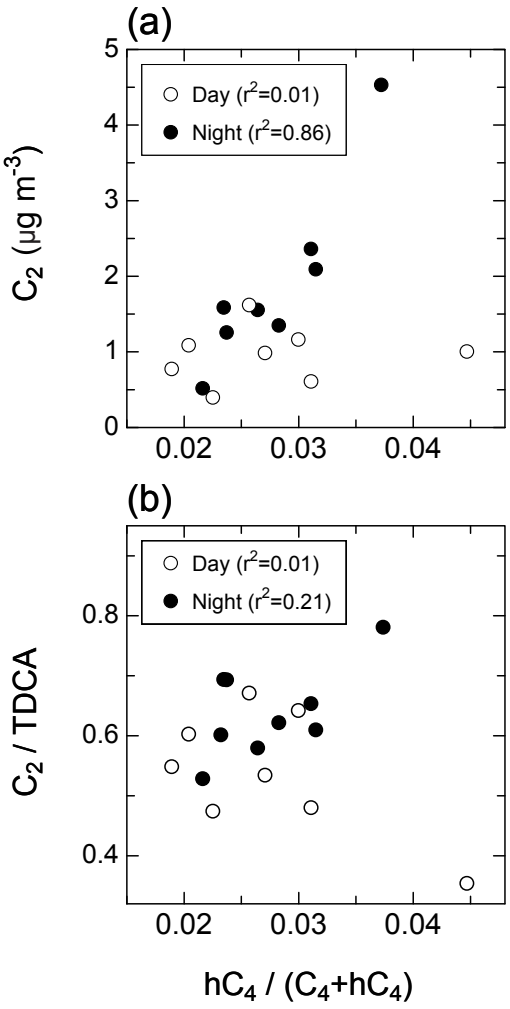
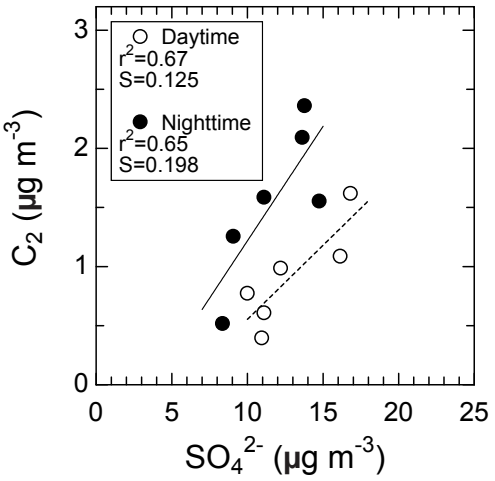

(a)
Day (r^2=0.01)
Night (r^2=0.86)
C_2 (μg m^{-3})
0
1
2
3
4
5
0.02
0.03
0.04
(b)
Day (r^2=0.01)
Night (r^2=0.21)
C_2 / TDCA
0.4
0.6
0.8
0.02
0.03
0.04
hC_4 / (C_4+hC_4)

Figure 9


Daytime
r²=0.67
S=0.125
Nighttime
r²=0.65
S=0.198
C2 (µg m-3)
SO4 2- (µg m-3)
0
1
2
3
0
5
10
15
20
25