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1 Perspectives on wild and hatchery salmon interactions at sea, potential climate effects on Japanese
2 chum salmon, and the need for sustainable salmon fishery management reform in Japan

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Abstract

Pacific salmon (*Oncorhynchus* spp.) play an important role as a keystone species and provider of ecosystem services in the North Pacific ecosystem. We review our studies on recent production trends, marine carrying capacity, climate effects and biological interactions between wild and hatchery origin populations of Pacific salmon in the open sea, with a particular focus on Japanese chum salmon (*O. keta*). Salmon catch data indicates that the abundance of Pacific salmon increased since the 1976/77 ocean regime shift. Chum and pink salmon (*O. gorbuscha*) maintained high abundances with a sharp increase in hatchery-released populations since the late 1980s. Since the 1990s, the biomass contribution of hatchery returns to the total catch amounts to 50% for chum salmon, more than 10% for pink salmon, and less than 10% for sockeye salmon (*O. nerka*). We show evidence of density-dependence of growth and survival at sea and how it might vary across spatial scales, and we provide some new information on foraging plasticity that may offer new insight into competitive interactions. The marine carrying capacity of these three species is synchronized with long-term patterns in climate change. At the present time, global warming has positively affected growth and survival of Hokkaido populations of chum salmon. In the future, however, global warming may decrease the marine carrying capacity and the area of suitable habitat for chum salmon in the North Pacific Ocean. We outline future challenges for salmon sustainable conservation management in Japan, and recommend fishery management reform to sustain the hatchery-supported salmon fishery while conserving natural spawning populations.

Keywords: Pacific salmon, ecological interaction, wild and hatchery-derived populations, carrying capacity, density-dependent effect, global warming, sustainable conservation management

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Introduction

The structure and function of ecosystems are influenced and disturbed not only by natural factors such as climate regimes, but also by anthropogenic impacts such as the global warming and overfishing. Pacific salmon (*Oncorhynchus* spp.) play an important role as a keystone species in the North Pacific ecosystems, and are useful for evaluation of ecosystem status given their high trophic level position in the ecosystem. According to results of carbon and nitrogen stable isotope analyses, they occupy the fourth and fifth trophic levels (Welch and Parsons 1993; Staterfield and Finney 2002; Kaeriyama et al. 2004; Johnson and Schindler 2008). Pacific salmon are also important for sustaining the biodiversity and productivity in the riparian freshwater ecosystem by supplying marine-derived material to the terrestrial ecosystem (Bilby et al. 1996; Hilderbrand et al. 1999; Kaeriyama & Minagawa 2008). However, in Japan Pacific salmon tend to be valued only as seafood and not for their important ecological role in the North Pacific (Kaeriyama 2008).

The biomass of Pacific salmon in the North Pacific Ocean increased from the late 1970s to the present owing to the favorable oceanic environment and increasing hatchery production (Beamish and Bouillon 1993; Mantua et al. 1997; Kaeriyama 1999a; Ruggerone et al. 2005). However, wild chum salmon (*O. keta*) populations since 1970s have decreased since the 1950s despite significant increases in hatchery stocking of this species in the North Pacific (Kaeriyama and Edpalina 2004; Ruggerone et al. 2010). One possible explanation for this is that hatchery fish have directly competed with and displaced wild salmon. Hilborn and Eggers (2000) reported that wild pink salmon (*O. gorbuscha*) were replaced with hatchery salmon in Prince William Sound due to higher survival rates and greater overall abundance than the wild fish, especially during their early marine life stages. Wertheimer et al (2001) later questioned the conclusions of Hilborn and Eggers (2000). In their rebuttal, Wertheimer et al. (2001) challenged their conclusion that hatchery pink salmon have replaced rather than enhanced pink salmon returns to Prince William Sound. They emphasized that there are a number of ways to interpret the data, and other equally plausible hypotheses to explain the observed trends, including hypotheses that address density-independent factors (e.g. zooplankton production) in the marine environment. They conclude that the situation is more complex than that implied by the analytical approach used by Hilborn and Eggers (2000). This debate underscores the need to develop better monitoring and assessment programs that would allow for a more direct assessment of effects hatchery salmon are having on wild salmon. This is becoming a more urgent need as hatchery salmon biomass in the North Pacific has grown along with interest in conservation of wild salmon populations (Kaeriyama and Edpalina 2004; Ruggerone et al. 2010).

While the focus of the 2010 State of the Salmon Conference focused primarily on

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3 71 ecological interactions between wild and hatchery salmon, it is important to acknowledge the genetic
4 72 risks to wild salmon posed by hatchery programs. Hatchery-origin salmon can introgress with wild
5 73 salmon and ultimately influence the fitness of wild Pacific salmon. Araki and Schmid (2010)
6 74 reviewed 266 peer-reviewed papers on effects of hatchery fish stocking on wild stocks and
7 75 consequences for stock enhancement, and concluded that negative effects of hatchery rearing on a
8 76 variety of fish species are common. Population and genetic structure of hatchery populations of
9 77 Hokkaido chum salmon has been studied in recent years (Okazaki 1982; Kaeriyama 1999a, Edpalina
10 78 et al. 2004; Yokotani et al. 2009). The hatchery-derived chum salmon have been determined to
11 79 exhibit lower genetic diversity than wild salmon. Okazaki (1982) noted that the average
12 80 heterozygosities of Japanese hatchery chum salmon were significantly lower than those of North
13 81 American populations based on allozyme analyses of populations. Heterozygosities estimated on the
14 82 basis of 22 loci ranged between 3% and 5%, which is relatively low compared with those of North
15 83 American chum salmon. Yokotani et al. (2009) surveyed population structure in the Yurappu River
16 84 using mitochondrial DNA (mtDNA) analysis. The Yurappu River chum salmon showed 8
17 85 haplotypes (Ht1-Ht8) in the 481bp 5' variable portion of the mtDNA control region. Pairwise
18 86 population F_{ST} estimates showed that the December-run population (YPD) differed significantly
19 87 from the October-run population (YPO) in the Yurappu River. The YPO population was closely
20 88 related to the other-river populations such as Chitose, Tokachi, and Nishibetsu river-populations on
21 89 the neighbor-joining tree. These results suggest that the Yurappu River chum salmon are genetically
22 90 different and perhaps reproductively isolated by run-timing. It is thought that the native population
23 91 persists as the late-run timing component, and that the early-run timing component represents an
24 92 introgressed population consisting of the native strain and out-of-basin transplants (Yokotani et al.
25 93 2009). Furthermore, hatchery programs can drastically lead to declines in the genetic variability of
26 94 wild Pacific salmon populations (Edpalina et al. 2004, Araki et al. 2007). Based on a review of 18
27 95 case studies, Berejikian and Ford (2004) concluded that steelhead (*O. mykiss*), coho (*O. kisutch*), and
28 96 Atlantic salmon (*Salmo salar*) populations have low lifetime relative fitness in the wild compared to
29 97 native, natural populations. Araki et al. (2007) indicate that hatchery salmonids have lower fitness in
30 98 natural environments than wild fish. This fitness decline can occur very quickly, sometimes
31 99 following only one or two generations of captive rearing. Outbreeding is a potential genetic risk in
32 100 Pacific salmon when aquaculture practices introduce nonnative domesticated fish to the wild
33 101 environments, making interbreeding with wild populations possible (McClelland et al. 2005). This
34 102 information suggests that we need to elucidate both ecological and genetic interactions between wild
35 103 and hatchery-derived salmon populations to help assure that both populations persist. Our rationale
36 104 to conserve wild and naturally reproducing salmon populations in Japan is based on their ability to
37 105 adapt readily to new environmental conditions.

38 106 We have reported on the population density-dependent effect of Hokkaido chum salmon

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3 107 with increasing population size since the 1980s (Kaeriyama 1989, 1998). This density-dependent
4 108 effect was also observed in North American populations of chum salmon in the same period (Helle
5 109 and Hoffman 1998), in wild sockeye salmon populations in Bristol Bay (Rogers and Ruggerone
6 110 1993), and in wild chum (Volobuev 2000) and pink salmon (Shuntov and Temnykh 2005) in the
7 111 Okhotsk Sea and in the Anadyr River (Zavolokin et al. 2009). While these studies have focused on
8 112 large scale ocean dynamics, we contend we can learn from experiences in small freshwater systems.
9 113 In the case of a sockeye salmon population in the oligotrophic Lake Shikotsu, Hokkaido, large
10 114 number of juveniles released from a hatchery increased the species density in this system and caused
11 115 density-dependent responses in growth and survival, leading to a crash of the population, a decline in
12 116 the effective population size, skewing of the sex ratio in the population, abnormal gonad
13 117 development, and an epidemic of fungal disease (Kaeriyama 1991). This system may serve as a
14 118 microcosm of the larger North Pacific ecosystem. While the scale and complexity of the North
15 119 Pacific ecosystem is considerably greater, there is an urgent need to accurately clarify mechanisms,
16 120 and describe the extent and intensity of density-dependent effects operating on salmon populations
17 121 in order to conserve Pacific salmon.

18 122 Global demand for wild and sea-ranched Pacific salmon has grown unabated, and this has
19 123 lead to growing economic interest in increasing hatchery production and improving survival rate of
20 124 released juveniles. This trend underscores the need to fully investigate potential negative effects of
21 125 hatchery fish on wild Pacific salmon. These biological interactions between wild and
22 126 hatchery-derived salmon populations should be given urgent attention in relation to their ecology,
23 127 behavior, genetics, and physiology. Additionally, chum salmon have been mass-produced by
24 128 hatchery programs in Japan, while numerous wild chum and masu salmon (*O. masou*) populations
25 129 have been put at risk from negative effects of hatchery programs operating in concert with a host of
26 130 other threats, including habitat loss and degradation, artificial river channelization, increases in
27 131 recreational fishing, and transplantation of exotic fishes into the river ecosystems in Japan
28 132 (Kaeriyama and Edpalina 2004).

29 133 In this paper, we review our studies on recent production trends, the relationship between
30 134 long-term climate change and carrying capacity, and biological interactions between wild and
31 135 hatchery-derived populations of Pacific salmon in the open sea. We add some new information
32 136 relating to marine salmon feeding ecology and we provide some new insights into how global
33 137 warming might affect Hokkaido chum salmon through the use of a path model. We offer our
34 138 perspective on these issues with a genuine interest in conserving wild salmon.

35 139 36 140 **Current situation of Pacific salmon biomass dynamics**

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38 142 Based on records of catch of wild and hatchery-origin Pacific salmon in the North Pacific (Fig. 1),

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3 143 the catches of pink, chum, and sockeye salmon (*O. nerka*) constituted more than 90% of the total
4 144 catch. The annual change in catch has roughly a 30- or 40-year periodicity, corresponding to the
5 145 long-term climate changes such as the Pacific Decadal Oscillation (PDO; Mantua et al. 1997) and
6 146 ocean regime shifts. It should be noted, however, that we did not see a marked shift in the North
7 147 Pacific salmon abundance (as reflected in catch) during the 1988/89 regime shift (Fig. 1, Kaeriyama
8 148 2008).

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11 149 We have been investigating trends in ocean salmon abundance and are developing
12 150 quantitative approaches to assess marine carrying capacity of sockeye, chum, and pink salmon
13 151 (Kaeriyama et al. 2009). North Pacific salmon abundance was estimated from catch data using an
14 152 expansion rate based on the fisheries exploitation rate (Eggers 2004; Kaeriyama et al. 2009).
15 153 Traditionally, the Ricker stock-recruitment function is assumed to account for density-dependence
16 154 operating on the spawning grounds or at some point during early life history in freshwater.
17 155 Alternatively, we think it is a useful exercise to explore the use of this model to explain potential
18 156 density-dependent processes operating during their marine life history (Yatsu and Kaeriyama 2005;
19 157 Yatsu et al. 2008). Kaeriyama et al. (2009) estimated the Ricker stock-recruitment equilibrium level
20 158 (the stock biomass at which recruits per spawner $(R/S) = 1$, i.e. the point at which the population is
21 159 at replacement) from brood tables based on expanded catch data. This equilibrium point is a measure
22 160 of production capacity of the species at the gross scale of the North Pacific, and provides an index of
23 161 carrying capacity (K) for each salmon species (Kaeriyama et al. 2009). We used a moving temporal
24 162 window of twenty years (projecting forward 10 generations of odd- and even-year of pink salmon
25 163 and 4-5 generations of chum and sockeye salmon) to resolve the potential influence of low
26 164 frequency ocean climate cycles (Fig. 2, Kaeriyama et al. 2009). While the causal mechanism and the
27 165 critical life history where recruitment success is determined are uncertain, Kaeriyama et al. (2009)
28 166 documented strong coherence between a large scale ocean climate signal (Aleutian Low Pressure
29 167 Index; Beamish and Bouillion 1993) and this Ricker stock equilibrium level (Fig. 2A). The carrying
30 168 capacity of three species (sockeye, chum, and pink salmon) decreased since the 1925 year-class, was
31 169 minimal during 1945-1955 year classes, increased during 1956-1975 year classes, and remained
32 170 constant during 1976-1997 year classes (Fig. 2B). These results suggest that there may be some
33 171 broad, density-dependent controls over salmon recruitment and productivity in the marine
34 172 environment.

35 173 36 174 **Intra-specific interactions of chum salmon**

37 175
38 176 Trends in abundance, which include catch and escapement, of wild and hatchery chum salmon
39 177 through the North Pacific region indicate that the mean population abundance of both wild and
40 178 hatchery chum salmon in the 1990s (132 million individuals) was roughly the same as that in the

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3 179 1930s (140 million individuals, Kaeriyama et al. 2009). However, the abundance of wild chum
4 180 salmon in the 1990s (67 million individuals) was only 50% of that in the 1930s despite a significant
5 181 increase in the abundance of hatchery populations across the North Pacific region (Fig. 3). The
6 182 1990s hatchery populations have contributed up to 50% to the total returns of chum salmon.
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8 183 Contributions of hatchery fish to returns of pink and sockeye salmon have been more modest (10%,
9 184 Fig. 3B, Kaeriyama et al. 2009). It is important here to recognize that the production potential of
10 185 wild salmon has been reduced during this period. In Japan, this is related to substantial habitat loss,
11 186 artificial river-channelization, and negative effects of hatchery programs (specifically river migration
12 187 impediments in the form of weirs and broodstock take by hatcheries). Although more difficult to
13 188 quantify, illegal take of salmon in Russia is also thought to be a significant threat to wild salmon
14 189 (Dronova and Spiridonov 2008). As previously noted, there is some evidence that wild populations
15 190 have been replaced with hatchery salmon due to competitive interactions during their early marine
16 191 life stages, such as pink salmon in the Prince William Sound (Hilborn and Eggers 2000). Ruggerson
17 192 et al. (2010) alleged that large and increasing abundance of hatchery salmon have affected
18 193 density-dependent processes of wild salmon.

194 Determining the nature and magnitude of density-dependence and ecological interactions
195 of salmon during their ocean life history is a very challenging area of research. We contend that
196 examining feeding habits of salmon in the ocean can provide novel insight into ecological processes
197 that may be regulating growth and survival during this important life stage. As a demonstration of
198 this, we surveyed the feeding habits of chum salmon collected by surface trawl nets in the Bering
199 Sea in the summer 2009, using the method of Kaeriyama et al. (2004). Chum salmon fed on diverse
200 prey-animals in the wider area, and predominantly consumed nekton in the coastal waters of the
201 Aleutian Islands, pteropods in the middle Bering Sea region, and a mixture of pteropods and
202 copepods in the northern part of the Bering Sea (Fig. 4). We found a positive correlation between an
203 index representing consumed food diversity (Shannon-Wiener Index) and CPUE (number of fish
204 captured per unit trawling effort) (Fig. 5). These results suggest that chum salmon show plasticity in
205 their feeding, shifting their diets from dominant, preferred prey species to more diverse prey with
206 increased salmon population density in order to minimize intra-specific competition. This
207 phenomenon has been observed in other studies. Tadokoro et al. (1996) found the diet of chum
208 salmon was correlated with the abundance of pink salmon owing to inter-specific competition. In
209 this study, the authors documented a shift in diets of chum salmon in the central Subarctic Pacific
210 related to the abundance of a resource competitor, pink salmon. In years with low pink salmon
211 abundance, chum salmon consumed a diet consisting mostly of nutritious assemblage of crustaceans.
212 In years of high pink salmon abundance, chum salmon consumed gelatinous zooplankton, a prey of
213 significantly lower quality. Thus, the increase of ocean salmon abundance caused by the increase in
214 hatchery populations could lead to elevated intra-specific interactions (e.g., ecological interactions

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215 between hatchery and wild salmon leading to a population density-dependent effect). On the other
216 hand, there is an opinion that Pacific salmon do not overpopulate under the high abundance because
217 of changing prey animals and expanding their distribution area (Shuntov and Temnykh 2005).

218 Insights can be gained by exploring relationships between size at age and age at maturity
219 of returning chum salmon to a simple index representing ocean production potential for chum
220 salmon. Here, we introduce a term, residual carrying capacity (RCC), which is meant to represent
221 the gap between potential ocean capacity to produce chum salmon (defined as carrying capacity
222 above) and that observed (total returns, summing catches and escapement). We present the following
223 simple equation here:

$$RCC = (\text{carrying capacity} - \text{abundance}) / (\text{carrying capacity}) \times 100,$$

224 where carrying capacity is derived from the Ricker stock equilibrium abundance (in return number
225 as defined above), and abundance is total returns (sum of catch and escapement) (Kaeriyama 2003).
226 We demonstrate the use of this by estimating RCC for a particular sex and age class of chum salmon
227 (age-4 females). We found a significant positive relationship between the RCC of Hokkaido chum
228 salmon and mean fork length (FL) of age-4 female adult chum salmon returning to 11 rivers in
229 Hokkaido ($r=0.979$, $F=753.8$, $P<0.001$). On the other hand, the mean age at maturity determined for
230 these 11 Hokkaido rivers showed a negative correlation ($r=-0.879$, $F=109.1$, $P<0.001$) with the RCC
231 (Fig. 6; Kaeriyama 2008). These results suggest that growth rates of these 11 Hokkaido chum
232 salmon populations respond to a population density-dependent effect, and growth reductions may
233 occur under conditions when RCC is low.

234 In an earlier paper we further explored how these chum salmon growth dynamics can be
235 scale-dependent. Seo et al. (2009) evaluated this by determining if variability in growth rates of
236 chum salmon observed in two separate river systems (Ishikari River in Japan and Namdae River in
237 Korea) could be explained by changes in abundance by evaluating dynamics at three different spatial
238 scales. These scales were: 1) Population level (sum of catches of chum salmon bound for Ishikari
239 and Namdae Rivers), 2) Asian population aggregate level (sum of catches of chum salmon bound for
240 Korea, Japan and Russia) and 3) North Pacific level (sum of all chum salmon captured across the
241 North Pacific). Growth rate (represented by FL at maturity of fish returning to the Ishikari and
242 Namdae Rivers) was found to vary only as a function of population abundance, and was not
243 significantly related to more aggregated abundance levels in the North Pacific. These results suggest
244 that growth of these two study populations (as reflected by changes in FL at maturity) responds
245 significantly to variability in their own abundance, but their dynamics do not appear to be strongly
246 influenced by competitive interactions occurring with other chum salmon stocks in the North Pacific.
247 This may be indicative of population-specific migration paths that lead to discrete partitioning of
248 ocean habitat occupied during their marine life history (e.g., Bugaev et al. 2009; Myers et al. 2009;
249 Urawa et al. 2009).

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251 At present, hatchery-derived chum salmon may not significantly affect wild salmon in
252 areas in the open sea such as the Bering Sea. The density-dependent effect seems to be limited to
253 body growth at the population level for chum salmon in the open sea. Ruggerone et al. (this volume),
254 however, provides some new, compelling evidence that productivity of western Alaska chum salmon
255 has been reduced as a result of increases in Hokkaido hatchery chum salmon in the Bering Sea over
256 the past several decades, leading to elevated intraspecific competitive interactions. In addition,
257 Ruggerone et al (2005) found a growth interaction between multiple populations of Bristol Bay
258 sockeye and an aggregate measure of Asian pink salmon. The density-dependent effects on Pacific
259 salmon may increase in the open sea in the near future, because global warming will likely decrease
260 their distribution areas and carrying capacity in the North Pacific Ocean.

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262 **Global warming effect on chum salmon**

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264 There is growing concern over potential impacts of climate change on Pacific salmon (Rand et al.
265 2006; Battin et al. 2007; Crozier et al. 2008, Schindler et al. 2008). Rand et al. (2006) simulated the
266 spawning migration of sockeye salmon in the Fraser River by accounting for energetic demands, and
267 results indicated that a long-term decline in the mean mass of adult sockeye salmon completing their
268 marine residency could erode their migratory fitness during the river migration and hence jeopardize
269 the sustainability of sockeye salmon and the fishery that targets them. Crozier et al. (2008) predicted
270 that global warming poses a direct threat to freshwater stages in Snake River Chinook salmon,
271 increasing their risk of extinction based on projections from several General Circulation Models.
272 Here we review some recent investigations on how climate variables, particularly sea surface
273 temperature, has influenced dynamics in Hokkaido chum salmon and how climate change may effect
274 this population and the North Pacific ecosystem in the future.

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276 We have investigated how SST explains significant variation in the annual changes in
277 survival rate and growth to age 1 (G1) of chum salmon returning to the Ishikari River from 1943 to
278 2005 (Kaeriyama et al. 2007; Seo et al. 2009). Ishikari River chum salmon live in the coastal waters
279 of Hokkaido and the Okhotsk Sea at age 1 (Urawa 2000), and they have responded positively
280 (increased survival and growth rate) to a trend in warming temperature observed in this region (Fig.
281 7). In a more recent paper (Seo et al. 2011) we extended these results by examining how past trends
282 in climate have directly and indirectly affected Hokkaido chum salmon. A path analysis was
283 conducted to assess both direct and indirect effects using a host of climate variables. The authors
284 found evidence of a direct negative effect between air temperature and sea ice cover in the Okhotsk
285 Sea region. The model also revealed a direct effect of SST on both growth and survival of chum
286 salmon (Fig. 8). Chum salmon growth, in particular was correlated both with surface air and sea
surface temperature. This study points to the importance of understanding the complexity of

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3 287 interactions leading to effects of climate dynamics on Pacific salmon.

4 288 By examining projected temperature isotherms and applying a optimum (8-12 °C;
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6 289 optimum growth rate and feeding habit), adaptable (5-13 °C; available habitat for swimming and
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8 290 feeding), and wintering temperature range (4-6 °C) of chum salmon (Kaeriyama 1986; Fukuwaka et
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10 291 al. 2007; Kaeriyama 2008), Kaeriyama (2008) projected a northward shift in the distribution of this
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12 292 species in the North Pacific into the Chukchi Sea by 2050, and a loss of a critical link in the
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14 293 migration route of chum salmon connecting Hokkaido with the Okhotsk Sea due to a thermal
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16 294 impediment (Fig. 9). A similar decrease in preferred habitat has also been projected for sockeye
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18 295 salmon in the North Pacific as a result of global warming (Welch et al. 1998). This projected loss of
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20 296 habitat would likely result in more intense ecological interactions occurring within and between
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22 297 salmon species occupying these habitats, and this could have important implications on growth,
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24 298 survival, production, fish condition and disease transmission. These effects may be exacerbated by
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26 299 growing contributions of hatchery releases to the salmon biomass in the North Pacific.

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28 300 Under a global warming scenario, therefore, interactions between wild and hatchery
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30 301 populations, including genetic, ecological, behavioral, and physiological dynamics, will likely
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32 302 become more intense in the near future. Japan, in particular, faces mounting challenges with
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34 303 sustaining wild and hatchery salmon populations. We outline some of these challenges and potential
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36 304 management responses in a companion paper in this special issue (Nagata et al., this issue). Efforts
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38 305 in Japan are now planned or underway to rehabilitate and restore freshwater salmon habitat, develop
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40 306 and implement a new management framework so wild salmon can be managed separately from
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42 307 hatchery populations, initiate new monitoring efforts to quantify natural escapement, and evaluate
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44 308 the effectiveness of different hatchery release strategies. Salmon are highly adaptable, and we
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46 309 believe these efforts will help Japanese chum salmon and other Japanese salmon populations and
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48 310 species cope with the broad impacts of global warming we expect in the future (Kaeriyama 1999b).

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51 312 **Concluding remarks**

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55 314 We outlined above some of our key research results relating to salmon abundance dynamics,
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57 315 intra-specific interactions operating in the marine environment, and how climate dynamics can affect
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59 316 Japanese chum salmon. The theme of this issue is focused on ecological interactions between wild
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61 317 and hatchery salmon. Understanding these interactions, and developing a management framework
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63 318 that can address this issue, is critical for the future of conservation and sustainable use of Japanese
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65 319 salmon. Therefore, these ecological interactions between hatchery and wild populations should be an
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67 320 important consideration in the sustainable conservation of Pacific salmon.

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69 321 To fully address this issue, we promote the use of adaptive management approaches and
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71 322 application of the precautionary principle to conserve salmon. Kaeriyama and Edpalina (2004)

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2
3 323 defined some of the important principles that need to guide future management of Japanese salmon.
4 324 Their action plan included: (1) the separation of wild and hatchery salmon populations in order to
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6 325 protect the genetic diversity and endemism of populations, (2) establishment of appropriate
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8 326 management zones (i.e., recreational fishing and hatchery program zones for commercial salmon
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10 327 fisheries) in rivers, (3) shifting the location of hatchery brood stock from river mouths to the outlet
11
12 328 of the hatchery facilities (thus allowing greater natural escapement into rivers where hatcheries
13
14 329 operate), and (4) regular biological monitoring of salmon body size, age composition, reproductive
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16 330 indices (fecundity and egg size), and dynamics of salmonid populations (Kaeriyama and Edpalina
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18 331 2004). Based on more recent research, we suggest two additional issues that must be addressed to
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20 332 ensure salmon are conserved and sustainably used. The first is recognition of limitations and
21
22 333 fluctuations of carrying capacity for Pacific salmon in the North Pacific. There is a growing body of
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24 334 scientific literature focusing on this issue, some of which we have referenced in the present paper.
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26 335 This issue cannot be ignored, and it is logical to conclude these interactions will likely become more
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28 336 intense in the future as demand for wild capture seafood continues to increase and ocean habitat for
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30 337 salmon will likely diminish. Finally, and perhaps most importantly, we need to focus on educating
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32 338 and training the next generation of fishery scientists, guided by a paradigm shift from the traditional
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34 339 fisheries science tailored to meet the needs of the fisheries industry to a new fisheries science based
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36 340 on ecological principles emphasizing aquatic ecosystem protection and sustainable fishing practices
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38 341 to provide seafood to future generations. In Figure 10, we provide an overview of the important
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40 342 characteristics of an adaptive management system that could be implemented to improve
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42 343 conservation management of Pacific salmon in Japan. New information on the status of natural
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44 344 reproduction of chum salmon is already being generated in Japan (see Miyakoshi et al., this issue). It
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46 345 will be important to develop mechanisms to allow our management system to adapt to this new
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48 346 information. Key questions will soon present themselves. For example, how can fishing pressure and
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50 347 effort be managed to allow increases in natural escapement in this region? How do we prioritize
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52 348 salmon river restoration and dam modification or removal to increase the contribution of natural
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54 349 spawners to overall juvenile salmon recruitment in this region? How can we ensure fishery
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56 350 management adaptively responds to emerging new information on trends and status of
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58 351 naturally-reproduced salmon? While these issues are complex and challenging, we are committed to
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60 352 working on salmon management reform to address these questions and work at sustaining our
61
62 353 productive hatchery-based salmon fishery while minimizing the future risks to wild salmon in Japan
63
64 354 and throughout the North Pacific region.
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54
55 356 **Acknowledgements**
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3 **552 Figure Captions**

4 553

5 554 Fig. 1. Annual changes in catch of Pacific salmon and PDO in the North Pacific Ocean. Catch data
6 includes both wild and hatchery-derived salmon (Kaeriyama 2008)
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10 557 Fig. 2. Temporal changes in ALPI and carrying capacity (K) of total (A) and each 3 species (B),
11 which are sockeye, chum, and pink salmon (Kaeriyama et al. 2009).
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13 558
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15 560 Fig. 3. Annual changes in abundance of wild and hatchery-derived populations of sockeye, chum,
16 and pink salmon in the North Pacific (modified Kaeriyama 2009).
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20 563 Fig. 4. Cluster analysis result of feeding habits of chum salmon in the Bering Sea in the summer
21 2009. Chum salmon fed on (1) diverse prey animals (red area), and predominantly consumed (2)
22 pteropods, (3) pteropods and copepods, and (4) nekton.
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27 567 Fig. 5. Relationship between the CPUE and food diversity of chum salmon collected by surface
28 trawl nets in the Bering Sea in the summer of 2009. The CPUE indicates number of fish per troll per
29 hour. The food diversity is calculated by the Shannon-Wiener's Index (H').
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33 571 Fig. 6. Relationships between the residual carrying capacity (RCC) and the fork length (A) or the age
34 at maturity (B) of chum salmon returning to 11 rivers in Hokkaido. The residual carrying capacity is
35 defined as “(carrying capacity – abundance) / carrying capacity” (Kaeriyama 2008).
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40 575 Fig. 7. Temporal changes in growth at age 1 (G1) and survival rate (SR) of Hokkaido chum salmon.
41 Growth anomaly was calculated for adult chum salmon returning to the Ishikari River using scale
42 analysis and back-calculation (modified Seo et al. 2011).
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47 579 Fig. 8. Path models of climate and oceanic environmental indexes, and growth at age 1 (G1) and
48 survival rate (SR), and population size (PS) of Hokkaido chum salmon. The growth at age 1 was
49 calculated for adult chum salmon returning to the Ishikari River using scale analysis and
50 back-calculation. Climate and oceanic environmental indexes are defined as follows. SAT: annual
51 change in global anomalies of Surface Air Temperature, PDO: Pacific Decadal Oscillation, ALPI:
52 Aleutian Low Pressure Index, SI: Siberian High, OH: Okhotsk High, AO: Arctic Oscillation, SST_O:
53 SST from June to October in the southern Okhotsk Sea, and ICE: Sea ice cover area (%) in the
54 Okhotsk Sea (Seo et al. 2011).
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588 Fig. 9. Prediction of the global warming effect for chum salmon in the North Pacific based on the
589 SRES-A1B scenario of IPCC (modified Kaeriyama 2008).

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591 Fig. 10. Conceptual diagram on the sustainable adaptive management of Pacific salmon.

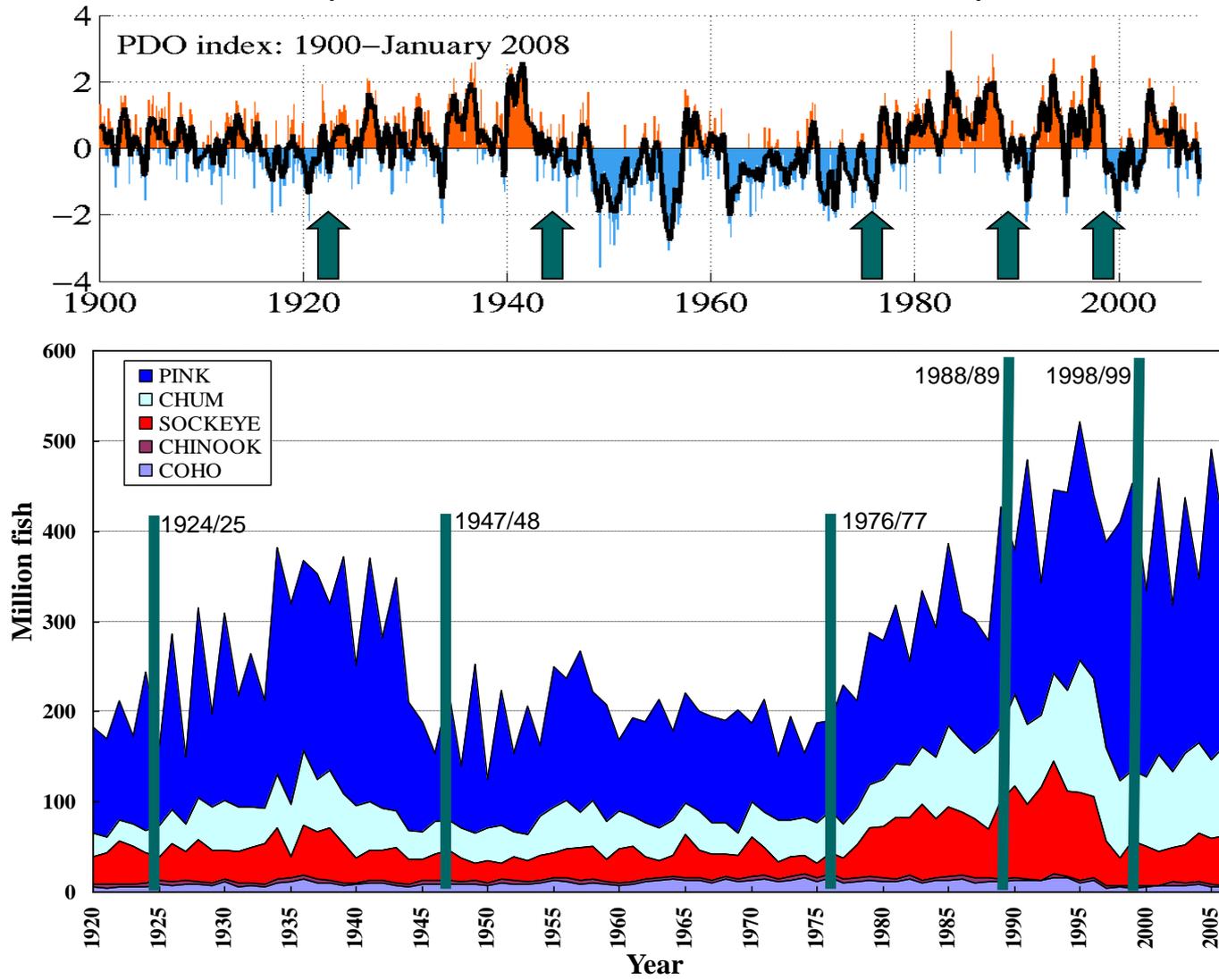


Fig. 1. (Kaeriyama et al.)

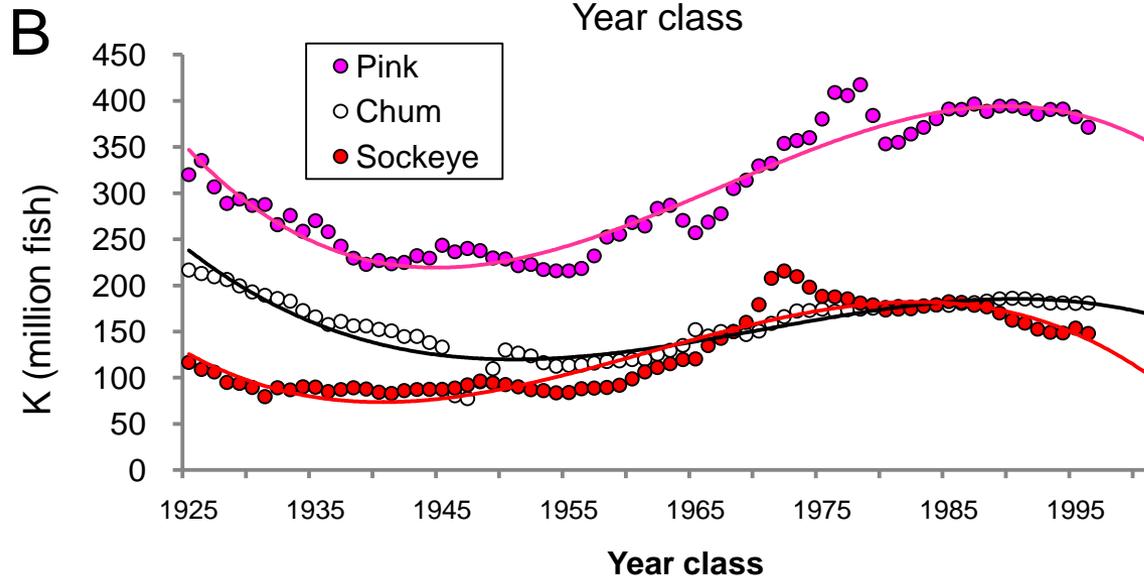
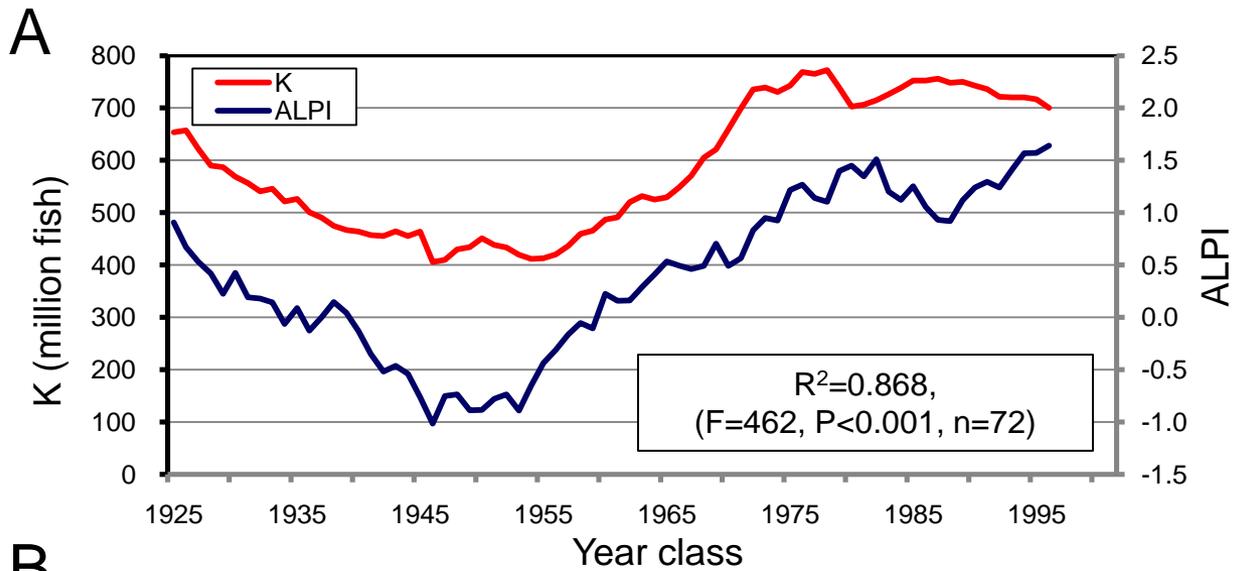


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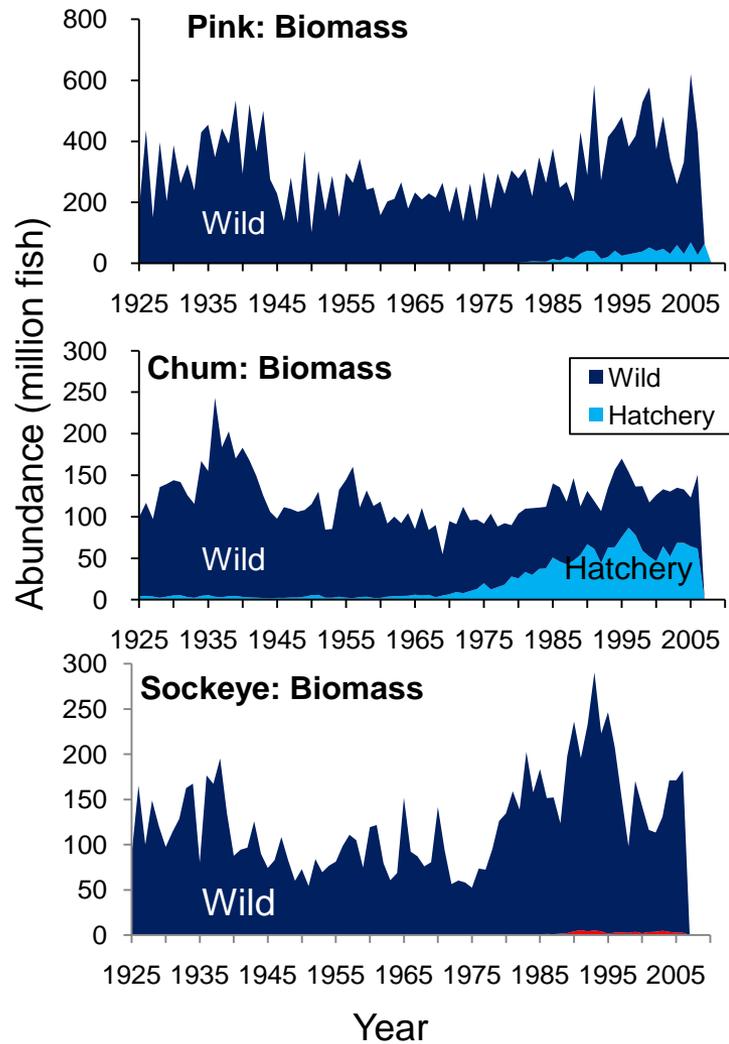
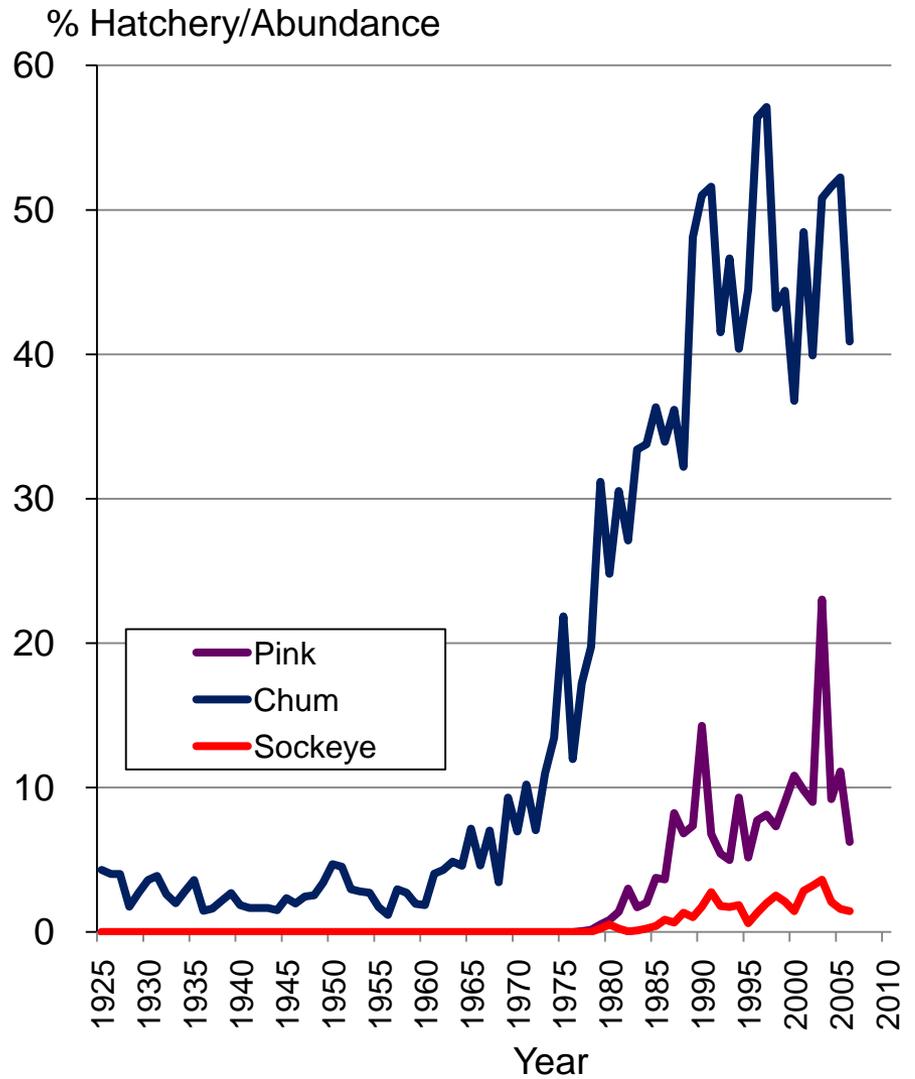
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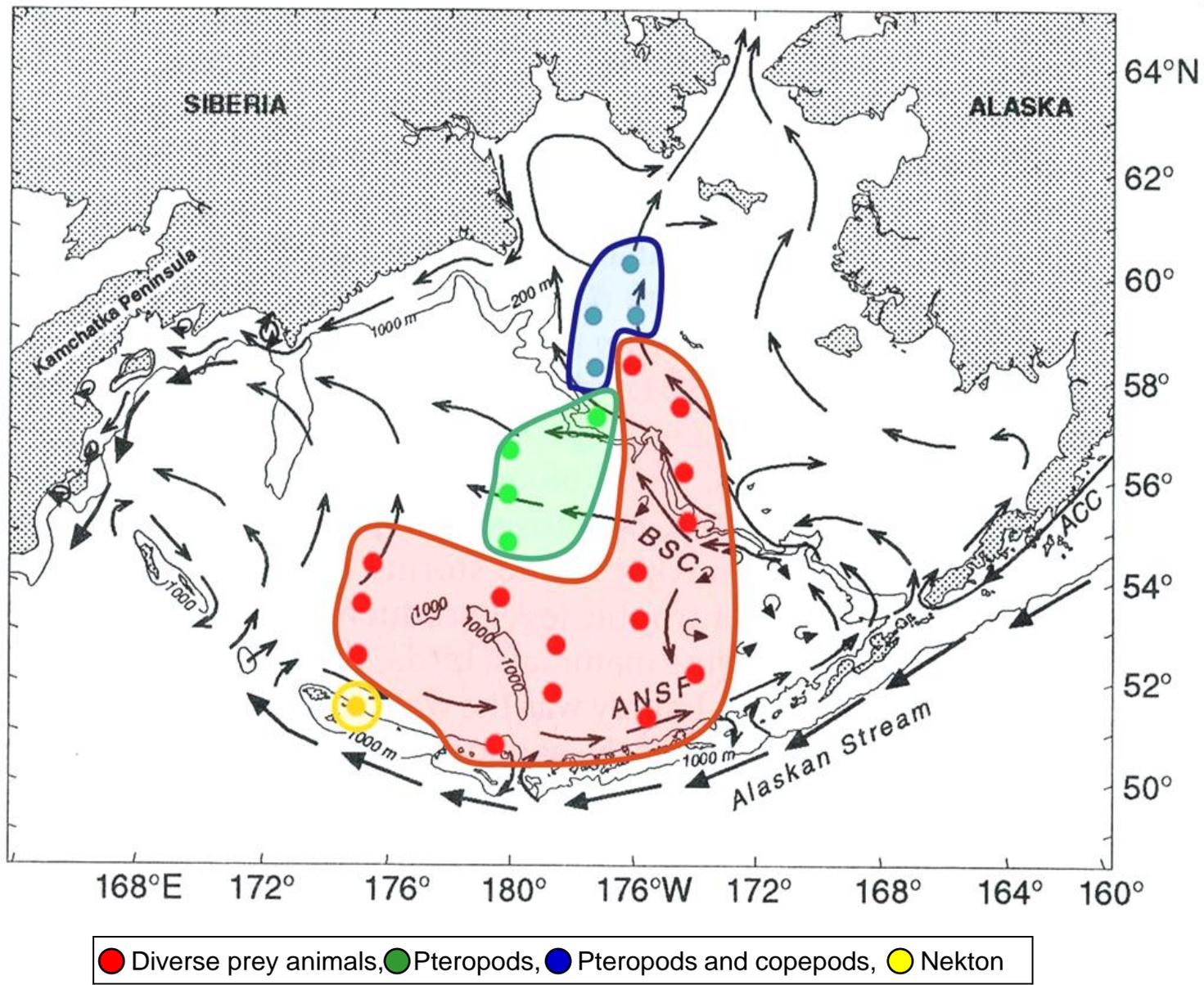


Fig. 4. (Kaeriyama et al.)

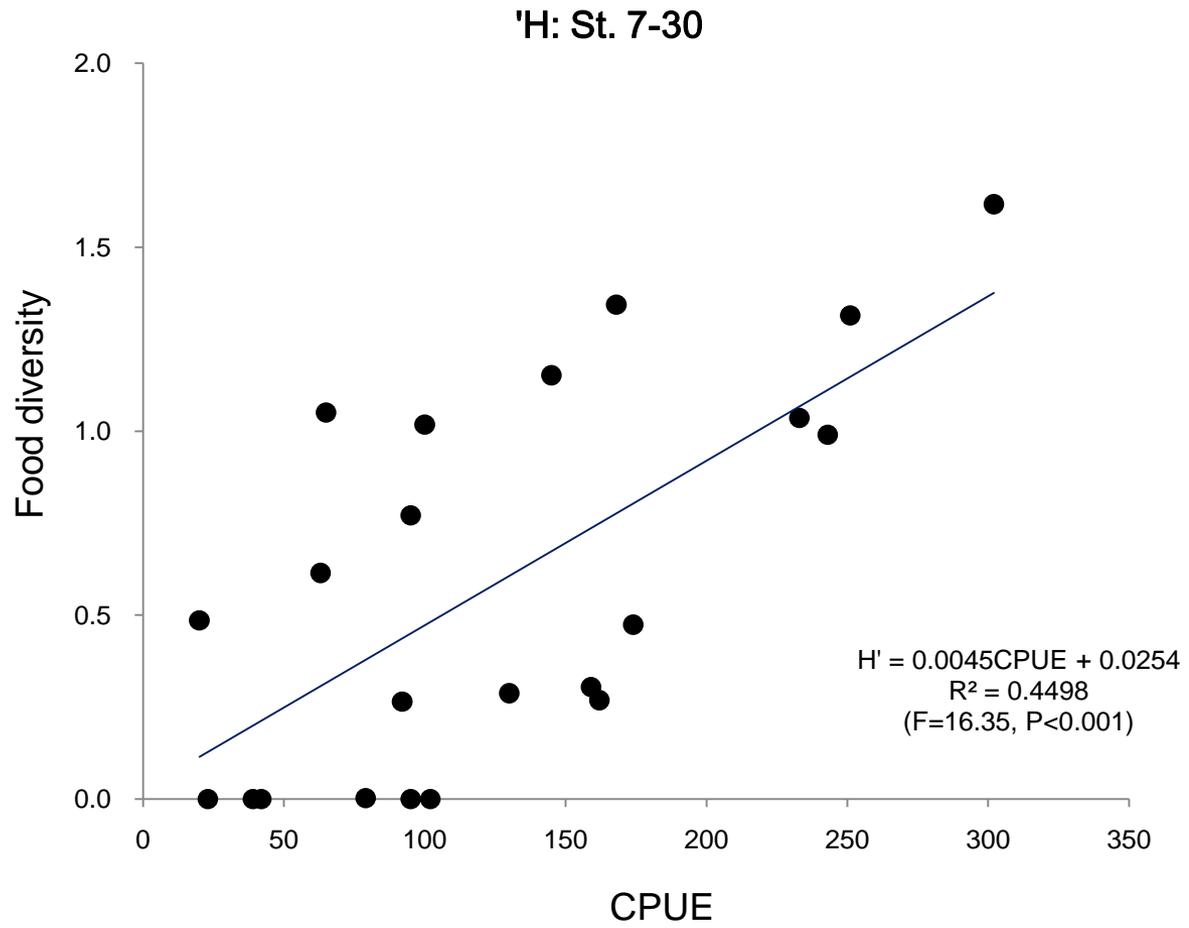


Fig. 5. (Kaeriyama et al.)

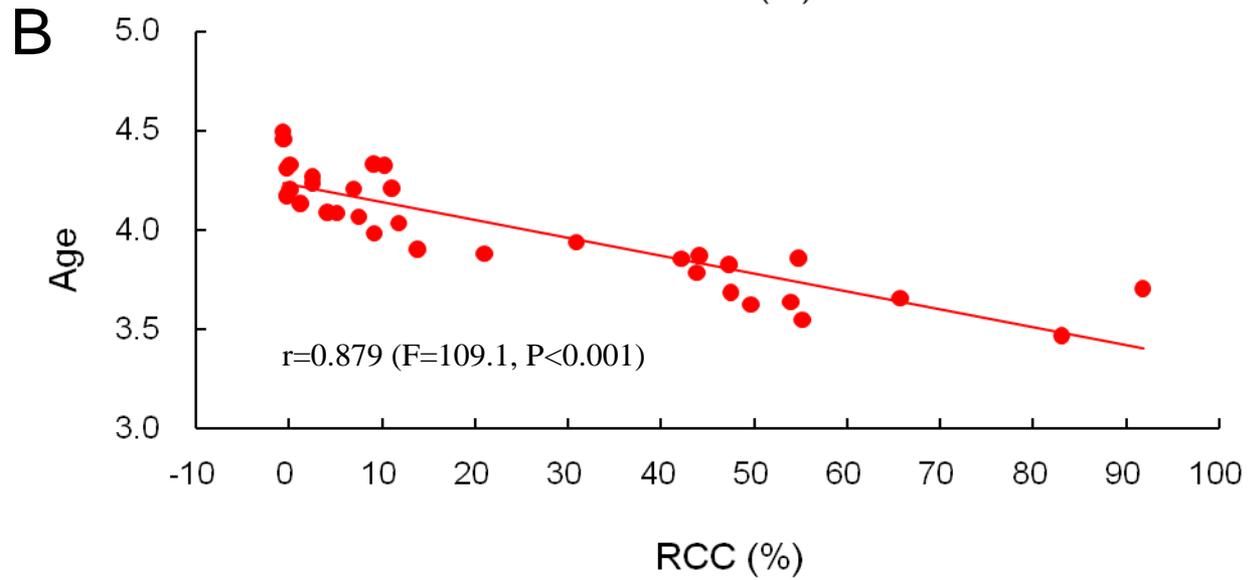
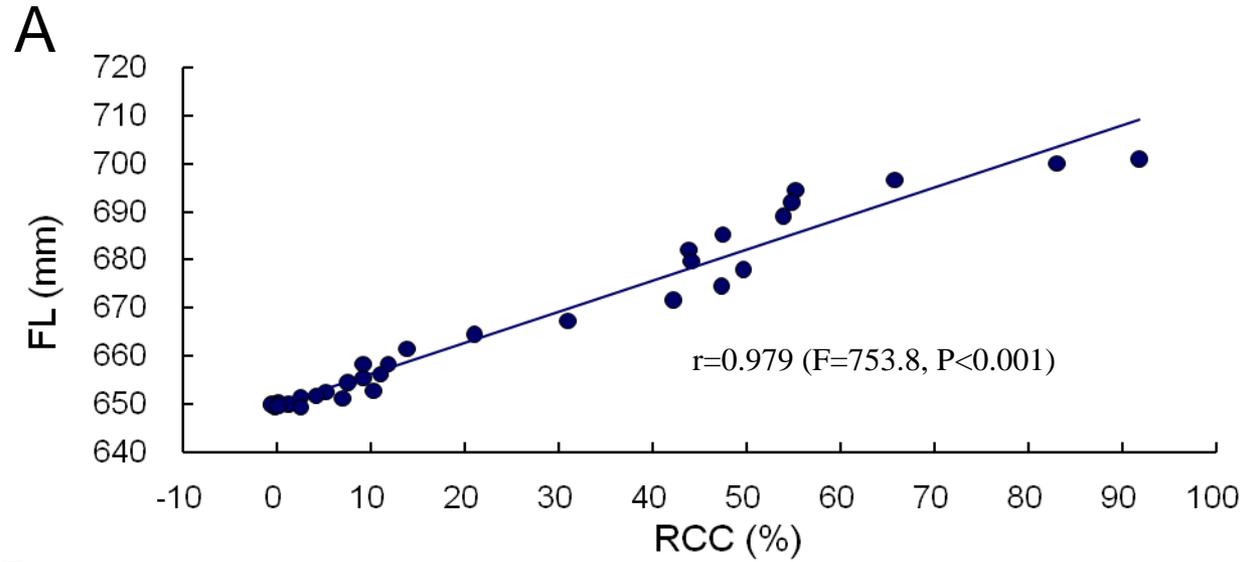


Fig. 6. (Kaeriyama et al.)

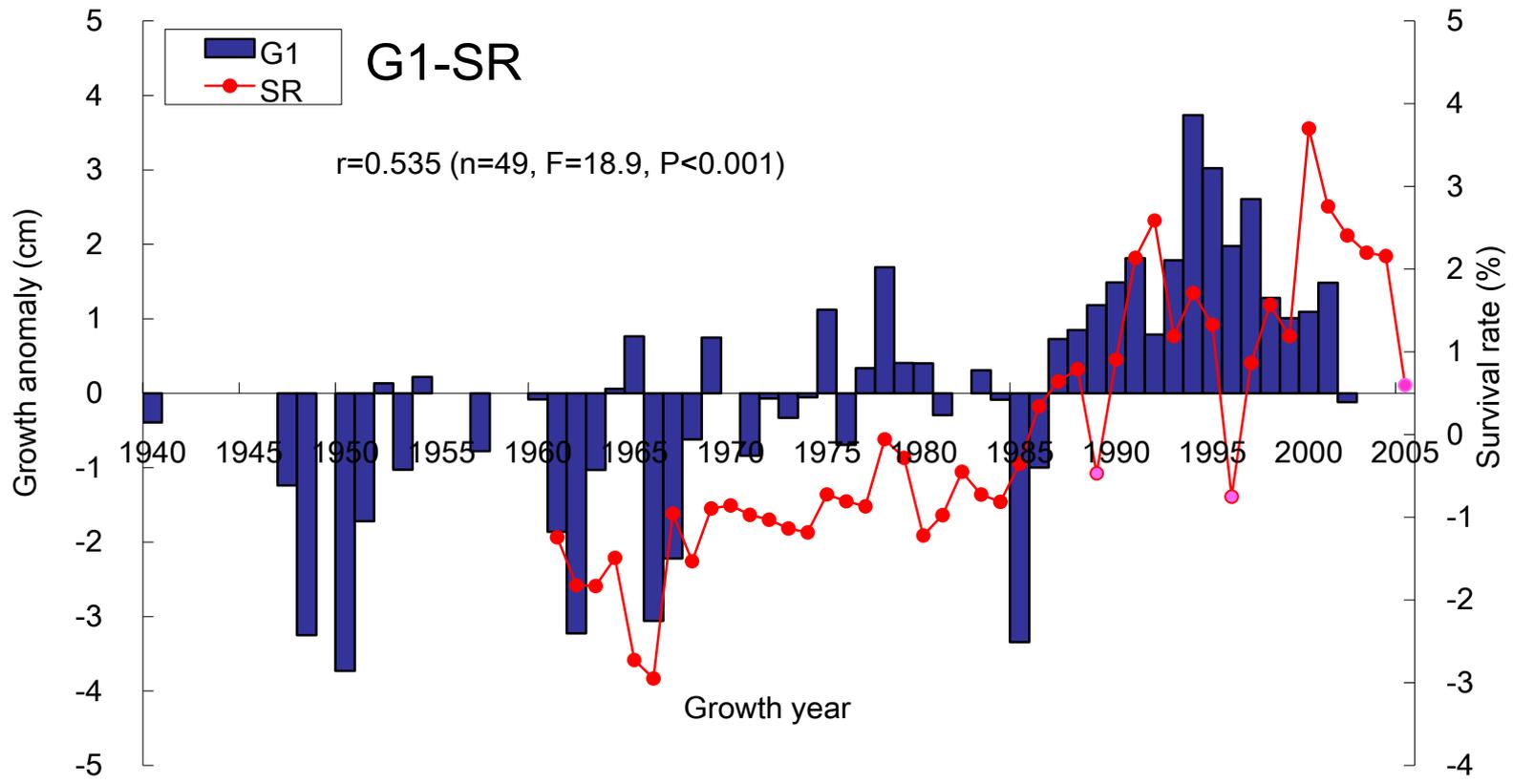


Fig. 7. (Kaeriyama et al.)

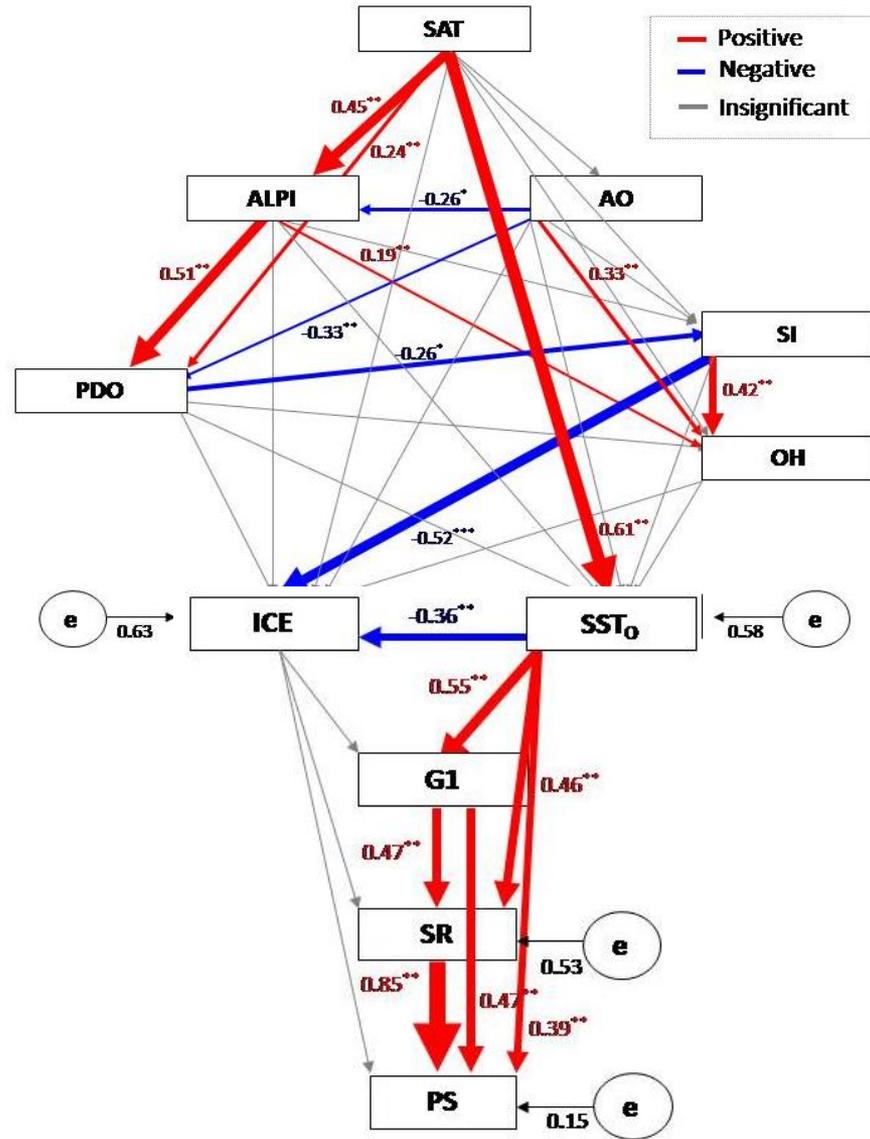


Fig.8. (Kaeriyama et al.)

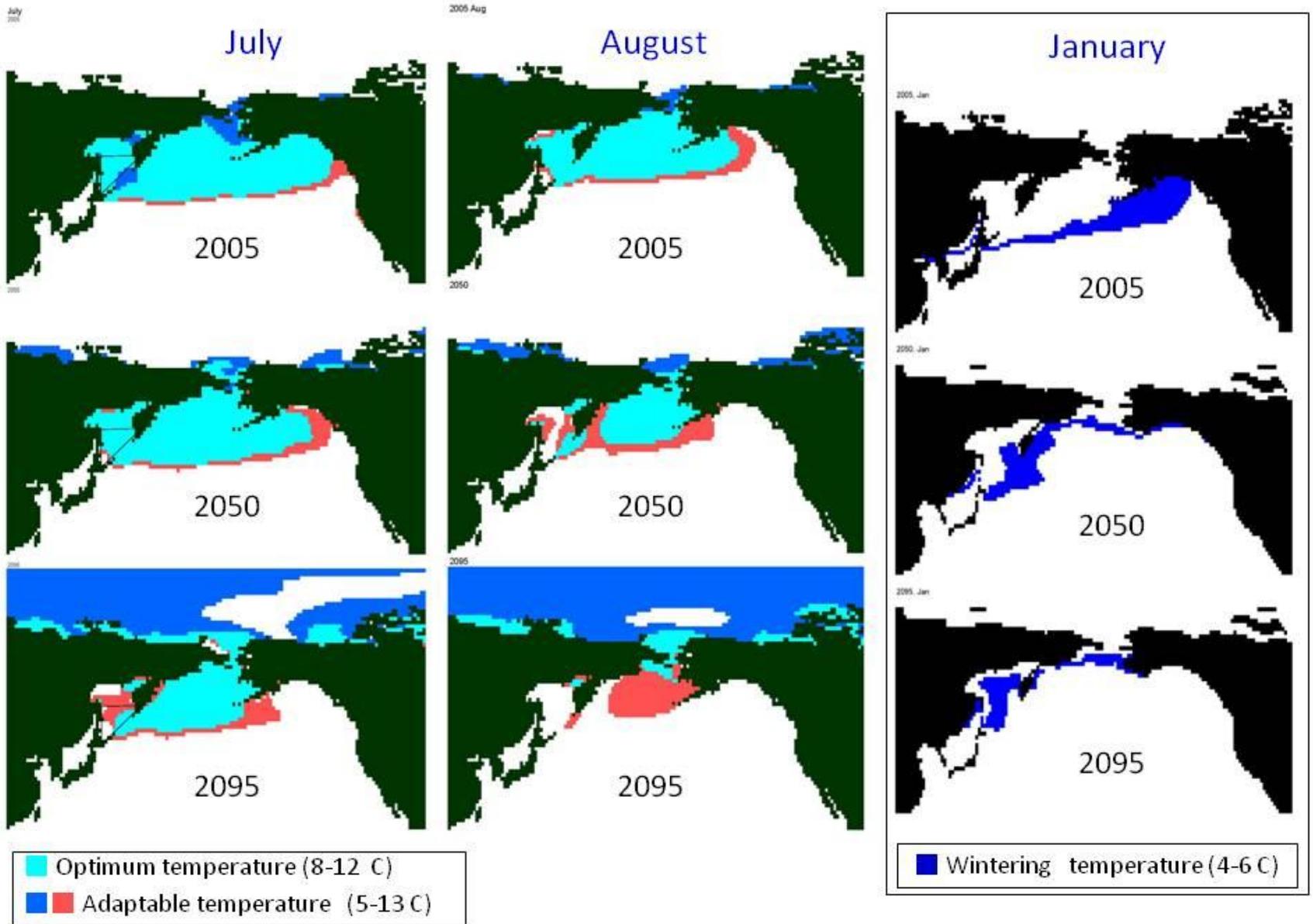


Fig. 9. (Kaeriyama et al.)

- Conservation: Natural ecosystems in the earth
- Paradigm shift of education: ecological fisheries science & human dietary
- Appropriate ecological interaction between wild & hatchery salmon populations

Concept

Evaluate

Plan & Design

Sustainable Adaptive Management of Pacific salmon

Implement & Actions

Monitoring

Analyze & Evaluate

- Conservation: Natural river ecosystems
- Protection: Wild salmon
- Establishment: Nursery & spawning areas in rivers
- Elimination: needless dams
- Separation: wild & hatchery salmon populations
- Discontinuance: salmon capture at river mouths
- Suitable operation: hatchery program

- Climate change
- Carrying capacity in the ocean
- Biological information of adult: body size, age composition, breeding (fecundity & egg size) & genetic characters
- Condition of river ecosystems

Fig. 10. (Kaeriyama et al.)