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1 Glutathione S-transferase gene polymorphisms in association with susceptibility to lead toxicity
2 in lead and cadmium exposed children near an abandoned lead-zinc mining area in Kabwe, Zambia
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42 **Abstract**

43 Interindividual genetic variations determine human's susceptibility to heavy metals-induced
44 toxicity. Thus, we analyzed blood concentrations of lead (Pb) and cadmium (Cd) in 140 lead-
45 exposed children. Genotyping of the glutathione S-transferase (GST) genes, *GSTM1*, *GSTT1*, and
46 *GSTP1* genes, was carried out to investigate their possible association with heavy metal
47 concentrations and the risk of susceptibility to Pb toxicity. Exposure to both heavy metals was
48 prevalent among the children. The blood Pb level ranged from 3.30 to 74.0 $\mu\text{g dL}^{-1}$ with an average
49 value of 26.8 $\mu\text{g dL}^{-1}$ that is five times above its reference level. The average Cd level (0.22 μg
50 L^{-1}) was below its reference level. The metal-gene interaction showed positive correlation between
51 *GSTT1* null genotype and Pb and Cd levels ($\beta = 0.11$; $p = 0.02$ and $\beta = 0.10$; $p = 0.01$, respectively).
52 More pronounced effects ($\beta = 0.19$; $p < 0.01$ and $\beta = 0.25$; $p = 0.04$) were found for the mixture
53 of the three putative genes with blood Pb concentration. The susceptibility analysis using 10 μg
54 dL^{-1} as blood Pb cut-off level showed a high risk of Pb toxicity (OR = 2.54; 95% CI: 1.02–6.32,
55 $p = 0.04$) for children carrying the *GSTP1 Ile/Val* genotype. Further, the combined effect of *GSTP1*
56 *Ile/Val* with *GSTT1* null genotype was more pronounced and showed an increased risk of
57 susceptibility to Pb toxicity (OR = 11.7; 95% CI: 1.36–102.1, $p = 0.02$). In summary, this study
58 suggests that *GSTT1* null and *GSTP1 Ile/Val* genotypes are the main genetic factors, and individual
59 and specific combinations of *GSTP1 Ile/Val* with *GSTM1* and *GSTT1* GST polymorphisms are
60 associated with susceptibility to Pb toxicity.

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63 **Keywords:** Blood, Children, Lead, Cadmium, *GST*, Polymorphism, Kabwe

64 **1. Introduction**

65 Heavy metals like lead (Pb) and cadmium (Cd) have no known biological roles in living organisms
66 and are considered as toxic metals. According to the International Agency for Research on Cancer
67 (IARC), Pb and Cd are classified as probable carcinogenic (group 2A) and known carcinogenic
68 (group 1) compounds, respectively (IARC 2021). Once inside the body, either from natural or
69 anthropogenic sources, they can interact metabolically with essential metals and displace them
70 from their specific cell constituent sites (Jan et al. 2015; Godwill et al. 2019). Generally, heavy
71 metal exposure can cause multiple organ dysfunction that led to toxicity and oxidative stress
72 (Tchounwou et al. 2012; Jan et al. 2015; Godwill et al. 2019). Owing to their developing bodies,
73 children are more prone to numerous adverse effects of heavy metals such as cognitive
74 impairments and central nervous system functional deficits that persist into older age (Jaishankar
75 et al. 2014).

76 Glutathione-S-transferases (GST), recognized as oxidative stress-related genes, are vital defense
77 enzymes engaged in metal biotransformation and detoxification of reactive oxygen species (Hayes
78 et al. 2005). It catalyzes the conjugations of hydrophobic and electrophilic compounds with
79 reduced glutathione (GSH) and excreted via feces and urine after GSH-metal conjugates formation
80 (Jozefczak et al. 2012). The GST superfamily comprises eight polymorphic genes, and of these,
81 polymorphic variants of the GST-mu 1 (*GSTM1*), theta 1 (*GSTT1*), and pi (*GSTP1*) are the most
82 studied and reported globally (Sharma et al. 2014; Saitou and Ishida 2015). Genetic variations in
83 these genes (deletion in *GSTM1* and *GSTT1* genes and polymorphism in *GSTP1*) contribute to
84 interindividual differences in susceptibility to xenobiotic toxicity (Hollman et al. 2016). Previous
85 studies reported an absence or decreased detoxification ability of GST enzymes and linked it with
86 oxidative stress and health outcomes in different polymorphic GST gene carriers. Individuals with

87 *GSTP1* variant alleles and double-null genotypes of *GSTM1* and *GSTT1* were associated with the
88 induction of oxidative stress (Sirivarasai et al. 2013). In another study, individuals with the *GSTM1*
89 null genotype and *GSTP1* wild allele had lower GSH levels and be at increased risk from exposure
90 to toxicants (Hunaiti and Soud 2000; Khansakorn et al. 2011). Pregnant mothers with combined
91 *GSTM1* and *GSTT1* genetic variants showed an inverse association of blood Pb and birth weight
92 (Lamichhane et al. 2018). However, polymorphisms in these functional genes vary by race and
93 ethnicity (Sharma et al. 2014; Saitou and Ishida 2015).

94 Kabwe, a capital town of the Central Province of Zambia, has a long history of Pb-zinc (Zn) mining
95 that operated from 1902 to 1994 without adequate pollution regulations. Thus, the town is one of
96 the most polluted places in the world. Even though the mine is inactive, the residue left a legacy
97 of contaminated soil with Pb, Zn, and other byproducts like cadmium (Cd). Moreover, illegal
98 mining activities are still ongoing. These create wide-contaminated surroundings and an
99 environment that poses a significant health risk. Recent studies involving a population-wide
100 screening on Pb poisoning in this town reported that approximately 75% of the residents had BLLs
101 higher than $5 \mu\text{g dL}^{-1}$ (Yabe et al. 2020; Yamada et al. 2020). Nevertheless, only one study
102 investigated gene polymorphism on candidate Pb biomarkers of effect genes delta-aminolevulinic
103 acid dehydratase (ALAD) and vitamin D receptor (VDR) genes in Kabwe (Yohannes et al. 2021).
104 Thus, this study (i) assessed the burden of Pb and Cd concentrations in children blood, (ii)
105 investigated the genetic polymorphism in the GST family; *GSTM1*, *GSTT1*, and *GSTP1*
106 polymorphisms, and (iii) examined the effect of the GST polymorphic variants on blood heavy
107 metal levels and the influence on the risk of susceptibility to Pb toxicity in children living near an
108 abandoned Pb-Zn mining area. We also reported Zn concentration. Thus, the data generated here
109 provide reference data for future studies.

110 2. Materials and methods

111 2.1. Study subjects and sampling

112 A total of 140 healthy children, aged 2–10 years, were recruited from five townships (Chowa,
113 Kasanda, Makululu, Bwacha, and Nakoli townships) in Kabwe, Zambia. Among these children,
114 102 children live in three townships near the abandoned mine and 38 children from two townships
115 live far away from the mine (Fig. 1). Children were invited to health centers, and parents were
116 asked a short questionnaire on demography and Pb-related clinical symptoms. No evidence or any
117 sign and symptoms of neurotoxicity among the children. Venous blood samples (≈ 3 mL) were
118 obtained from each child by trained nurses into heparinized blood collection tubes in July 2016
119 (Yohannes et al. 2020). A 500 μ L blood sample was kept in a 1.5 mL tube for heavy metal analysis
120 and stored all samples at -20 °C. All analyses were done in the Laboratory of Toxicology, Faculty
121 of Veterinary Medicine, Hokkaido University, Japan.

122 Fig. 1

123 2.2. Heavy metals analysis

124 Blood sample digestion and heavy metal analyses were done as described by Yohannes et al.
125 (2017). Briefly, the 500 μ L blood sample was digested with 5 mL 61% HNO₃ and 1 mL 30% H₂O₂
126 in a speedwave MWS-2 microwave system. Then, samples were made to a final volume of 10 mL
127 using Milli-Q water in a 15 mL tube. Blood Pb, Cd, Zn concentrations were measured using
128 inductively coupled plasma mass spectrometry (ICP-MS 7700 series). Seronorm™ Trace
129 Elements Whole Blood L-2 certified reference material (Sero AS, Norway) was used as quality
130 control. Replicate analyses of this reference material showed recovery rates ranged from 90% to
131 105% with a detection limit of 0.01 μ g L⁻¹ for the measured heavy metals.

132 2.3. *GSTs* genotyping

133 Genomic DNA was isolated from whole blood using a NucleoSpin Blood Kit (Macherey-Nagel)
134 and kept at -25°C until genotyping analysis. A multiplex polymerase chain reaction (PCR) was
135 used for *GSTMI* and *GSTT1* genotyping to detect the presence or absence of the genes with the
136 *CYP 1A1* gene as a positive control. A PCR-restriction fragment length polymorphism (PCR–
137 RFLP) method was carried out for *GSTP1 Ile105Val* (rs1695; A > G) polymorphism. PCR was
138 done to a mixture containing 1 μL DNA (>20 ng), 1x PCR buffer, 2.0 mM MgCl_2 , 200 μM of each
139 dNTP, 0.25 μM of each primer, and 1 U Taq polymerase in a total volume of 10 μL . The sequences
140 of primers and the PCR amplification conditions are displayed in Table 1.

141 Table 1

142 The presence of *GSTMI*, *GSTT1*, and *CYP 1A1* genes were identified by the bands at 219 bp, 480
143 bp, and 312 bp, respectively. The absence of the band describes the gene deletion referred to as
144 null genotype. For *GSTP1* genotyping, the resulting 433 bp PCR product was digested at 55°C
145 for 4 h with the restriction enzyme *BsmAI* (5 U). The corresponding fragmented product bands
146 were visible with 328 bp and 105 bp fragments (*Ile/Ile* genotype), 328 bp, 222 bp, and 105 bp
147 fragments (*Ile/Val* genotype), and 222 bp and 105 bp fragments (*Val/Val* genotype). The PCR
148 fragments were electrophoresed in a 2% agarose gel with Midori Green direct and visualized under
149 a Blue-LED transilluminator system. Representative gel electrophoresis showed the genotypes of
150 *GSTMI*, *GSTT1*, and *GSTP1* are presented in Fig. 2. For ensuring quality control, random samples
151 (10%) were genotyped twice, and no contamination was confirmed using no-template control
152 samples.

153 Fig. 2

154 2.4. Statistical analysis

155 To improve normality and homogeneity of variance of heavy metal concentrations, data were log-
156 transformed before statistical analysis. Blood heavy metals were expressed as mean \pm standard
157 deviation and analyzed by the non-parametric Kruskal-Wallis test. Deviation of *GSTP1* genotype
158 distribution from Hardy–Weinberg equilibrium (HWE) was assessed using the Pearson’s Chi-
159 squared (χ^2) test. Comparisons of the observed *GSTM1* and *GSTT1* null genotypes and *GSTP1*
160 *Val* allele frequencies between this study and global studies were examined using the χ^2 test. We
161 checked the effect of covariates (age and sex) on the statistical analysis, and no significant
162 association was found. Thus, linear regression analysis was performed for assessing the association
163 between polymorphic variants of GST genes and log-transformed blood heavy metal
164 concentrations. Next, crude odds ratios (ORs) and 95% confidence intervals (CIs) were calculated
165 to examine associations between the GST genes genotypes and susceptibility to Pb toxicity using
166 10 $\mu\text{g/dL}$ as a cutoff level. A $p \leq 0.05$ was considered statistically significant, and a p between
167 0.051 and 0.100 was marginally significant (Amrhein et al. 2019). The JMP 14.0 software (SAS
168 Institute, Cary, NC, USA) was used to perform all analyses.

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175 **3. Results and discussion**

176 *3.1. Blood Pb, Cd, and Zn concentrations*

177 The subjects of the current study were 65 females and 75 male children with mean ages of 5.72
178 and 5.69 years, respectively. The majority of children have lived in their respective townships for
179 most of their lifetime. The descriptive statistics of heavy metal concentrations are given in Table
180 2. The blood Pb concentrations among participants ranged from 3.30 to 74.0 $\mu\text{g dL}^{-1}$ with the mean
181 and median concentrations of 26.8 and 26.3 $\mu\text{g dL}^{-1}$ (interquartile range: 13.6–36.6 $\mu\text{g dL}^{-1}$),
182 respectively. These mean and median Pb levels were 5 times higher than the blood Pb reference
183 value of 5 $\mu\text{g dL}^{-1}$ (CDC, 2012). Among the 140 participants, 131 (93.5%) children exceeded this
184 blood Pb reference value. The mean/median concentrations of Cd and Zn were 0.22/0.18 $\mu\text{g L}^{-1}$
185 and 4135/4167 $\mu\text{g L}^{-1}$, respectively. These concentration values were below their respective
186 reference values (Table 2).

187 In addition, the measured heavy metals demonstrated some correlations among them and showed
188 Spearman's correlation coefficients $\rho = 0.497$ for Pb–Cd and $\rho = 0.400$ for Pb–Zn. These
189 correlations imply that these elements share the same sources of exposure and children are exposed
190 to both toxic metals. The closed mine is still a source of pollution. As shown in table 2, the
191 concentrations of the measured heavy metals near the mine site were significantly higher than far
192 from the mine site. There were no significant differences between blood heavy metal
193 concentrations and sex, showing equal exposure regardless of sex (data not shown). Owing to their
194 immature defense mechanisms, children are more vulnerable to heavy metals toxicity. Thus,
195 appropriate measures should be taken to prevent or minimize exposure for reducing the toxic
196 effects of Pb and Cd on the community.

197 3.2. *GSTs* polymorphism and blood Pb, Cd, and Zn concentrations by *GST* polymorphic variants

198 Table 3 shows the distribution rate of *GSTM1*, *GSTT1*, and *GSTP1* genotypes. We found *GSTM1*
199 present genotype as the most prevalent genotype with a frequency of 76.4%, followed by the
200 *GSTT1* present genotype (65.0%) and then the *GSTP1 Ile/Val* genotype (59.3%). The null
201 genotype prevalence was 23.6% and 35.0% for *GSTM1* and *GSTT1*, respectively. The combined
202 polymorphic variants of *GSTM1* and *GSTT1* were 8.6% for null/null, 41.4% for null/present, and
203 50.0% for present/present genotypes. Regarding the *GSTP1* polymorphism, 40 children (28.6%)
204 were *Ile/Ile* homozygous, 83 children (59.3%) were heterozygous *Ile/Val*, and 17 children (12.1%)
205 were *Val/Val* homozygous genotypes (Table 3). The distribution of *GSTP1 Ile105Val* showed
206 significant deviation from HWE ($\chi^2 = 6.69$; $p < 0.05$). The analysis revealed an excess number of
207 heterozygous *Ile/Val* genotypes (observed = 83 children vs expected = 68 children). The
208 prevalence of *GSTP1 Ile* and *Val* allele frequencies were 58% and 42%, respectively.

209 The blood heavy metal concentrations per *GSTM1*, *GSTT1*, and *GSTP1* genotypes are depicted in
210 Table 3. None of the measured heavy metals showed significant differences between the genotypes
211 of *GSTM1* and *GSTP1* genes. Whereas the *GSTT1* null group showed significantly higher blood
212 levels for Pb and Cd (Table 3).

213 Table 3

214 *GST* gene polymorphisms have been studied in various human races with pervasive existence (Fig.
215 3). Globally, the *GSTM1* null genotype frequency ranged from 20–46% in Africans, 42–55% in
216 Europeans, 36–52% in Americans, and 41–60% in Asians (Fig. 3a) (Palma-Cano et al. 2017). The
217 distribution frequency of *GSTT1* null genotype ranged from 30–50% for Asians followed by
218 Africans (20–47%), then Europeans and Americans (11–26%) (Fig. 3a) (Palma-Cano et al. 2017).

219 Thus, our study *GSTMI* null genotype prevalence rate of 23.6% was found in the lower range in
220 the Africans and lower than the levels reported from other continents. The 35% prevalence rate for
221 the *GSTTI* null genotype in the current study was significantly higher than those reported from
222 European and American countries. Comparing our study *GSTP1 Val* allele frequency (42%) with
223 global *Val* allele frequency based on the new Allele Frequency Aggregator (ALFA) dataset
224 (https://www.ncbi.nlm.nih.gov/snp/rs1695#frequency_tab), we observed apparent variation of the
225 *Val* allele frequencies between Africans and Asians (Fig. 3b). The *Val* allele was more frequent in
226 the Latin American individuals with mostly European and Native American Ancestry (49.7%),
227 followed by Africans (42%), Europeans (32.6%), and then Asians (18.9–28.9%). The chi-squared
228 statistical analysis showed clear out variation between this study and Asians ($p < 0.05$).

229 Fig. 3

230 In respect to metal–gene interactions, the effect of the three putative genes on blood heavy metal
231 levels showed both direct and inverse significant effects. The regression analysis coefficient for
232 genetic variants of GST genes in association with blood heavy metals is shown in Table 4. Logistic
233 regression analysis using log-transformed concentration as a dependent variable confirmed a
234 positive significant association for the *GSTTI* null genotype with blood Pb ($\beta = 0.11, p = 0.02$)
235 and Cd ($\beta = 0.10, p = 0.01$) levels. In the combined analysis, the result showed direct association
236 of blood Pb level with *GSTTI* null / *GSTP1 Ile/Val* genotypes ($\beta = 0.19, p = 0.002$), with *GSTMI*
237 present / *GSTTI* null / *GSTP1 Ile/Val* genotypes ($\beta = 0.16, p = 0.03$), and with *GSTMI* null /
238 *GSTTI* null / *GSTP1 Ile/Val* genotypes ($\beta = 0.25, p = 0.04$). For Cd, the double null genotypes
239 showed a significant higher association with beta coefficient of 0.11 ($p = 0.05$). On the other hand,
240 the *GSTP1 Ile/Ile* genotype showed an inverse association with blood Pb level ($\beta = -0.10, p =$

241 0.04), suggesting a protective role of this genotype. The combined effect of this genotype with
242 others also showed negative beta estimates; *GSTP1 Ile/Ile / GSTT1* null ($\beta = -0.17, p = 0.03$),
243 *GSTP1 Ile/Ile / GSTM1* present / *GSTT1* null ($\beta = -0.28, p = 0.01$), and *GSTP1 Ile/Ile / GSTM1*
244 null / *GSTT1* present ($\beta = -0.32, p = 0.03$) in relation with blood Pb levels.

245 Table 4

246 To our knowledge, most of the studies on the impact of GST polymorphism are based on
247 case-control studies and only a few had available on healthy people (Khansakorn et al. 2011,
248 2012). In nonoccupationally exposed populations, Khansakorn et al. (2011) reported an association
249 between GST polymorphism and the level of Cd in blood. The result revealed higher Cd levels for
250 subjects with the *GSTP1 Val/Val* genotype than their counterparts. The combination of *GSTP1*
251 with *GSTM1* and *GSTT1* also showed an association with increased blood Cd levels. In our study,
252 the metal-to-gene interaction showed significant associations between GSTs polymorphism and
253 blood heavy metal concentrations. The study revealed that individual and combined GST gene
254 variants could play substantial roles in heavy metal accumulation in our bodies. The result showed
255 direct positive associations between gene deletion of *GSTT1* and blood Pb and Cd concentration.
256 Moreover, *GSTP1* polymorphism (substitution of *Ile* \rightarrow *Val*) coupled with the *GSTT1* null genotype
257 showed a higher association with blood Pb concentration. This incidence can reduce the GST
258 enzyme activity and as a result reduce Pb- and Cd-GSH conjugates and excretion. Thus, these
259 toxic metals can accumulate in blood and tissues (Jozefczak et al. 2012). It is known that Pb and
260 Cd can replace essential elements such as calcium, copper, iron, and zinc in several major
261 biological processes, and mimic the actions of essential elements (Tchounwou et al. 2012). Overall,
262 the combined effect of *GSTP1* polymorphism with *GSTM1* and *GSTT1* gene deletions could affect
263 the levels of Pb and Cd in children living around the closed mine.

264 3.3. The effect of GST genotypes on susceptibility to lead toxicity

265 The GST genetic variant's impact on susceptibility to Pb toxicity was investigated using a blood
266 Pb level of $10 \mu\text{g dL}^{-1}$ as a cutoff point. Thus, we categorized the subjects as low exposed (≤ 10
267 $\mu\text{g dL}^{-1}$) and high exposed ($> 10 \mu\text{g dL}^{-1}$) children. The frequencies of the GST genotypes in these
268 groups are depicted in Table 5. The prevalence of *GSTMI* null in high exposed (23.7%) and low
269 exposed (23.1%) groups were similar and susceptibility analysis showed no association between
270 *GSTMI* polymorphism and susceptibility to Pb toxicity (OR = 1.03; 95% CI: 0.38–2.84, $p = 0.94$).
271 For the *GSTTI* polymorphism, the null genotype was marginally higher in the highly exposed
272 group (35.6%) than the low exposed group (19.2%) ($\chi^2 = 3.49$; $p = 0.06$). Analysis of susceptibility
273 to Pb toxicity also showed a marginal association between the *GSTTI* null genotype and risk of
274 lead toxicity (OR = 2.64; 95% CI: 0.92–7.51, $p = 0.06$), suggesting that individuals with *GSTTI*
275 null genotype are at higher risk of susceptibility to Pb toxicity. The combined effects of *GSTMI*
276 present / *GSTTI* null, *GSTMI* null / *GSTTI* present, and *GSTMI* null / *GSTTI* null genotypes were
277 not associated with Pb toxicity ($p = 0.12$, 0.56, and 0.23, respectively).

278 The *GSTPI* polymorphism deviated from HWE for the high exposed group ($\chi^2 = 7.83$; $p = 0.005$)
279 (Table 5). The *Ile/Val* genotypes showed a higher prevalence of 62.3% in the high exposed group
280 compared with 46.2% in the low exposed group. The *Ile* and *Val* allele percentages were 56% and
281 44% in the high exposed group, while 67.7% and 32.3% in the low exposed group. These allele
282 frequency differences showed marginal differences ($\chi^2 = 3.18$, $p = 0.07$) between the groups. This
283 result implies that living in precarious conditions might affect the genetic variants. The genetic
284 impact of *GSTPI* on susceptibility to Pb toxicity revealed a significant association of the *GSTPI*
285 *Ile/Val* genotype with the susceptibility to Pb toxicity. Individuals with this genotype were at a
286 2.54-fold higher risk of Pb toxicity (OR = 2.54; 95% CI: 1.02–6.32, $p = 0.04$) than individuals

287 carrying the *Ile/Ile* genotype. As for the wild *Val/Val* genotype, the OR ratio was greater than one
288 (OR = 3.21) but not significant ($p = 0.15$). The combination of the heterozygous *Ile/Val* and mutant
289 homozygous *Val/Val* genotypes showed a risk of susceptibility to Pb toxicity increased up to 2.63
290 ($p = 0.02$) times compared with the wild homozygous *Ile/Ile* genotype. The combined effect of all
291 three putative genes showed associations on susceptibility to Pb toxicity. The combination of the
292 *GSTP1 Ile/Val* with *GSTT1* null showed a risk factor of 11.7 (95% CI: 1.36–102.1, $p = 0.02$), while
293 *GSTP1 Ile/Val* with *GSTM1* present and the combination of *GSTP1 Ile/Val* / *GSTM1* present /
294 *GSTT1* null genotypes showed marginal associations with Pb susceptibility (OR = 2.62, 95% CI:
295 0.91–7.50, $p = 0.07$ and OR = 8.12, 95% CI: 0.89–73.8, $p = 0.06$, respectively) compared to their
296 non-risk genotype counterparts (Table 5). In general, the high prevalence of the *Ile/Val* genotype
297 might play the main role in Pb susceptibility.

298 Table 5

299 Studies have reported an increased health risk with occupational or environmental Pb exposure
300 among adults with GST polymorphisms. The *GSTM1* and *GSTT1* null genotypes with increased
301 levels on oxidative stress biomarkers such as malondialdehyde and high-sensitivity C-reactive
302 protein (Khansakorn et al. 2011; Sirivarasai et al. 2013), *GSTT1* present genotype with increased
303 Pb-related hypertension (Lee et al. 2012), and *GSTP1* risk genotypes (*Ile/Val* and *Val/Val*) with
304 the increasing effect of Pb on amyotrophic lateral sclerosis disease (Eum et al. 2015) were reported.
305 In another study, a $15 \mu\text{g g}^{-1}$ tibia Pb concentration was associated with weak cognitive function
306 (a decrement of Mini-Mental State Exam score by 0.24 point) on men with *GSTP1* risk genotype
307 (Eum et al. 2013). In our study, the genetic variants on susceptibility to Pb toxicity showed
308 significant associations with *GSTT1* null and *GSTP1* risk genotypes (*Ile/Val* and *Val/Val*
309 genotypes). The logistic regression analysis showed a risk of > 1 for *GSTT1* null, *GSTP1* risk

310 genotypes. Moreover, the combined effect of *GSTP1 Ile/Val* and *GSTT1* null showed a more
311 pronounced risk of 11.7 times compared with the non-risk genotypes. Overall, these results
312 indicated that children with deletions of the *GSTT1* gene and *GSTP1 Ile/Val* and *Val/Val* genotypes
313 are at higher risk of Pb susceptibility due to high blood Pb levels. Our findings stress continued
314 prevention programs on lead exposure in younger populations as lead exposure at early ages has
315 harmful impacts on cognitive functions and impair development that can continue throughout the
316 lifespan (Sanders et al. 2009).

317 We noticed that there are some limitations to this study. i) the small number (26 out of 140) in the
318 low exposed group ($< 10 \mu\text{g dL}^{-1}$), ii) heavy metal levels were measured at a single point blood
319 sample at which the half-life of heavy metals in the blood is short (especially Pb) compared to
320 bone, and iii) no epidemiological or clinical data for the study subjects. All these factors might
321 inevitably alter determining a true effect incidence and weakened the statistical power on the
322 outcome result. This study, however, still contributes to the understanding of the GSTs'
323 involvement in the development of childhood Pb toxicity in children. Second, in this study, we
324 measured the levels of other heavy metals to highlight their burden on the children residing near
325 the abandoned Pb mine site.

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331 4. Conclusions

332 Our results stress the significance of genetic factors in determining individual differences in
333 blood heavy metal concentrations. We found a direct association between *GSTT1* null genotype
334 and blood Pb and Cd concentrations. Furthermore, *GSTT1* null genotype and *GSTP1* risk
335 genotypes (*Ile/Val* and *Val/Val*) play more significant roles in modifying the blood Pb and Cd and,
336 as a result, be risk factors associated with Pb susceptibility. Owing to GST's potential role
337 in detoxifying toxic metals, children with more copies of the *GSTP1 Val* allele and gene deletion
338 of *GSTT1* are at higher risk of Pb toxicity. These results suggest that the exposure of heavy metals,
339 especially Pb, may be influenced by genotypic variants of the GST genes. As a preliminary study,
340 the results generated will use as baseline data for future studies. However, replication of this
341 finding using population-based research is warranted.

342

343 **Ethical approval:** Study protocol approval and permission to conduct the research were obtained
344 from the University of Zambia Research Ethics Committee (UNZAREC; REF. No. 012-04-16)
345 and the Ministry of Health Zambia, respectively. Material transfer agreement (MTA, Approval No.
346 E00417) has been issued from the Ministry of Health, Zambia for transporting frozen samples to
347 Japan.

348

349 **Consent to participate:** Participation in this study was voluntary and participants enrolled after
350 getting signed informed consent from the parents.

351

352 **Consent for publication:** Children's' parent signed informed consent before participation and be
353 aware regarding publishing the data. All the coauthors have read the manuscript and agreed for
354 publication.

355

356 **Competing interests:** The authors have declared that no competing interests exist.

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Authors' contribution

YBY: responsible for conceptualization, data analysis, investigation, methodology, and writing both drafting the original paper and reviewing.

SMMN: responsible for conceptualization, analysis, funding acquisition, investigation, and writing, and corresponding author.

JY, HN, HT, AK and KM: participate in sample collection, material preparation, and reviewing the manuscript.

YK: Funding acquisition, and supervision.

KC: Supervision.

MI: responsible for project administration, funding acquisition, supervision, and reviewing the manuscript and corresponding author.

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409 **References**

410 Amrhein V, Greenland S, McShane B (2019) Scientists rise up against statistical significance.
411 *Nature* 567:305–307. <http://doi: 10.1038/d41586-019-00857-9>

412 Centers for Disease Control and Prevention (CDC) (2012) Update on BLLs in children. Available
413 online at http://www.cdc.gov/nceh/lead/ACCLPP/blood_lead_levels.htm (accessed on June 01,
414 2020)

415 Eum KD, Seal RM, Taylor KM, Grespin M, Umbach DM, Hu H, Sandler DP, Kamel F, Weisskopf
416 MG (2015) Modification of the association between lead exposure and amyotrophic lateral
417 sclerosis by iron and oxidative stress related gene polymorphisms. *Amyotroph Lateral Scler*
418 *Frontotemporal Degener* 16(1-2):72–79. <http://doi: 10.3109/21678421.2014.964259>

419 Eum KD, Wang FT, Schwartz J, Hersh CP, Kelsey K, Wright RO, Spiro A, Sparrow D, Hu H,
420 Weisskopf MG (2013) Modifying roles of glutathione S-transferase polymorphisms on the
421 association between cumulative lead exposure and cognitive function. *Neurotoxicology* 39:65–71.
422 <http://doi: 10.1016/j.neuro.2013.08.002>

423 Godwill EA, Ferdinand PU, Nwalo FN, Unachukwu MN (2019) Mechanism and health effects of
424 heavy metal toxicity in humans. In: *Poisoning in the modern world - new tricks for an old dog?*
425 Intechopen. <https://doi.org/10.5772/intechopen.82511>

426 Hayes JD, Flanagan JU, Jowsey IR (2005) Glutathione transferases. *Annu Rev Pharmacol Toxicol*
427 45:51–88. <http://doi: 10.1146/annurev.pharmtox.45.120403.095857>

428 Hays SM, Nordberg M, Yager JW, Aylward LL (2008) Biomonitoring equivalents (BE) dossier
429 for cadmium (Cd) (CAS No. 7440-43-9). *Regul Toxicol Pharmacol* 51(3):S49–56. [http://doi:](http://doi:10.1016/j.yrtph.2008.05.008)
430 [10.1016/j.yrtph.2008.05.008](http://doi:10.1016/j.yrtph.2008.05.008)

431 Hollman AL, Tchounwou PB, Huang HC (2016) The association between gene-environment
432 interactions and diseases involving the human GST superfamily with SNP variants. *Int J Environ*
433 *Res Public Health* 13(4):379. [http://doi: 10.3390/ijerph13040379](http://doi:10.3390/ijerph13040379)

434 Hunaiti AA, Soud M (2000) Effect of lead concentration on the level of glutathione, glutathione
435 S-transferase, reductase and peroxidase in human blood. *Sci Total Environ* 248:45–50. [http://doi:](http://doi:10.1016/s0048-9697(99)00548-3)
436 [10.1016/s0048-9697\(99\)00548-3](http://doi:10.1016/s0048-9697(99)00548-3)

437 International Agency for Research on Cancer (IARC) (2021) IARC monographs on the
438 identification of carcinogenic hazards to humans. List of classifications. Agents classified by the
439 IARC Monographs, Volumes 1–129 <https://monographs.iarc.who.int/list-of-classifications>

440 Jaishankar M, Tseten T, Anbalagan N, Mathew BB, Berregowda NK (2014) Toxicity, mechanism
441 and health effects of some heavy metals. *Interdiscip Toxicol* 7(2):60–72. [http://doi: 10.2478/intox-](http://doi:10.2478/intox-2014-0009)
442 [2014-0009](http://doi:10.2478/intox-2014-0009)

443 Jan AT, Azam M, Siddiqui K, Ali A, Choi I, Rizwanul Haq QM (2015) Heavy metals and human
444 health: mechanistic insight into toxicity and counter defense system of antioxidants. *Int J Mol Sci*
445 16(12):29592–29630. [http://doi: 10.3390/ijms161226183](http://doi:10.3390/ijms161226183)

446 Jozefczak M, Remans T, Vangronsveld J, Cuypers A (2012) Glutathione is a key player in metal-
447 induced oxidative stress defenses. *Int J Mol Sci* 13(3):3145–3175. [http://doi:](http://doi:10.3390/ijms13033145)
448 [10.3390/ijms13033145](http://doi:10.3390/ijms13033145)

449 Kasperczyk S, Kasperczyk A, Ostalowska A, Dziwisz M, Birkner E (2004) Activity of glutathione
450 peroxidase, glutathione reductase, and lipid peroxidation in erythrocytes in workers exposed to
451 lead. *Biol Trace Elem Res* 102:61–72. [http://doi: 10.1385/bter:102:1-3:061](http://doi:10.1385/bter:102:1-3:061)

452 Khansakorn N, Wongwit W, Tharnpoophasiam P, Hengprasith B, Suwannathon L,
453 Chanprasertyothin S, Sura T, Kaojarern S, Sritara P, Sirivarasai J (2012) Genetic variations of
454 glutathione S-transferase influence on blood cadmium concentration. *J Toxicol* 2012:356126.
455 [http://doi: 10.1155/2012/356126](http://doi:10.1155/2012/356126)

456 Khansakorn N, Wongwit W, Tharnpoophasiam P, Hengprasith B, Suwannathon L, Pethchpoung
457 K, Yoovathaworn K, Chanprasertyothin S, Sura T, Kaojarern S, Sritara P, Sirivarasai J (2011)
458 Impact of *GSTM1*, *GSTT1*, *GSTP1* polymorphism and environmental lead exposure on oxidative
459 stress biomarkers. *Sci Res Essays* 6(31):6540–6547. [http://doi: 10.5897/SRE11.1519](http://doi:10.5897/SRE11.1519)

460 Lamichhane DK, Leem JH, Park CS, Ha M, Ha EH, Kim HC, Lee JY, Ko JK, Kim Y, Hong YC
461 (2018) Associations between prenatal lead exposure and birth outcomes: modification by sex and
462 *GSTM1/GSTT1* polymorphism. *Sci Total Environ* 619–620:176–184. [http://doi:](http://doi:10.1016/j.scitotenv.2017.09.159)
463 [10.1016/j.scitotenv.2017.09.159](http://doi:10.1016/j.scitotenv.2017.09.159)

464 Lee BK, Lee SJ, Joo JS, Cho KS, Kim NS, Kim HJ (2012) Association of glutathione S-transferase
465 genes (*GSTM1* and *GSTT1*) polymorphisms with hypertension in lead-exposed workers. *Mol Cell*
466 *Toxicol* 8:203–208. <https://doi.org/10.1007/s13273-012-0025-5>

467 Palma-Cano LE, Córdova EJ, Orozco L, Martínez-Hernández A, Cid M, Leal-Berumen I, Licón-
468 Trillo A, Lechuga-Valles R, González-Ponce M, González-Rodríguez E, Moreno-Brito V (2017)
469 *GSTT1* and *GSTM1* null variants in Mestizo and Amerindian populations from northwestern

470 Mexico and a literature review. *Genet Mol Biol* 40(4):727–735. <http://doi: 10.1590/1678-4685->
471 GMB-2016-0142

472 Poddalgoda D, Macey K, Hancock S (2019) Derivation of biomonitoring equivalents (BE values)
473 for zinc. *Regul Toxicol Pharmacol* 106:178-176. <https://doi.org/10.1016/j.yrtph.2019.04.018>

474 Saitou M, Ishida T (2015) Distributions of the *GSTMI* and *GSTTI* null genotypes worldwide are
475 characterized by latitudinal clines. *Asian Pac J Cancer Prev* 16:355–361. <http://doi:>
476 10.7314/apjcp.2015.16.1.355

477 Sanders T, Liu Y, Buchner V, Tchounwou PB (2009) Neurotoxic effects and biomarkers of lead
478 exposure-Review. *Rev Environ Health* 24(1):15–45. <http://doi: 10.1515/reveh.2009.24.1.15>

479 Sharma A, Pandey A, Sharma S, Chatterjee I, Mehrotra R, Sehgal A, Sharma JK (2014) Genetic
480 polymorphism of glutathione S-transferase P1 (*GSTP1*) in Delhi population and comparison with
481 other global populations. *Meta Gene* 2:134–142. <https://doi.org/10.1016/j.mgene.2013.12.003>

482 Sirivarasai J, Wananukul W, Kaojarern S, Chanprasertyothin S, Thongmung N, Ratanachaiwong
483 W, Sura T, Sritara P (2013) Association between inflammatory marker, environmental lead
484 exposure, and glutathione S-transferase gene. *BioMed Res Int* 2013:474963. <http://doi:>
485 10.1155/2013/474963

486 Tchounwou PB, Yedjou CG, Patlolla AK, Sutton DJ (2012) Heavy metals toxicity and the
487 environment. *EXS* 101:133–164. http://doi: 10.1007/978-3-7643-8340-4_6

488 Yabe J, Nakayama SMM, Nakata H, Toyomaki H, Yohannes YB, Muzandu M, Kataba A, Zyambo
489 G, Hiwatari M, Narita D, Yamada D, Hangoma P, Munyinda NS, Mufune T, Ikenaka Y, Choongo
490 K, Ishizuka M (2020) Current trends of blood lead levels, distribution patterns and exposure

491 variations among household members in Kabwe, Zambia. *Chemosphere* 243:125412.
492 <https://doi.org/10.1016/j.chemosphere.2019.125412>

493 Yamada D, Hiwatari M, Hangoma P, Narita D, Mphuka C, Chitah B, Yabe J, Nakayama SMM,
494 Nakata H, Choongo K, Ishizuka M (2020) Assessing the population-wide exposure to lead
495 pollution in Kabwe, Zambia: an econometric estimation based on survey data. *Sci Rep* 10:15092.
496 <http://doi: 10.1038/s41598-020-71998-5>

497 Yohannes YB, Nakayama SMM, Yabe J, Nakata H, Toyomaki H, Kataba A, Muzandu K, Ikenaka
498 Y, Choongo K, Ishizuka M (2020) Bleed lead levels and aberrant DNA methylation of the ALAD
499 and p16 gene promoters in children exposed to environmental-lead. *Environ Res* 188:109759.
500 <https://doi.org/10.1016/j.envres.2020.109759>

501 Yohannes YB, Nakayama SMM, Yabe J, Toyomaki H, Kataba A, Nakata H, Muzandu K, Ikenaka
502 Y, Choongo K, Ishizuka M (2021) Delta-aminolevulinic acid dehydratase (ALAD) and vitamin D
503 receptor (VDR) genes polymorphisms in children residing in an abandoned lead-zinc mine area in
504 Kabwe, Zambia. *Meta Gene* 27:100838. <https://doi.org/10.1016/j.mgene.2020.100838>

505 Yohannes YB, Yoshinori Y, Nakayama SMM, Mizukawa H, Ishizuka M (2017) Trace Element
506 Contamination in Tissues of Four Bird Species from the Rift Valley Region, Ethiopia. *Bull*
507 *Environ Contam Toxicol* 98:172–177. <http://doi: 10.1007/s00128-016-2011-4>

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Tables and Figures

Figure captions

Fig. 1. Gel electrophoresis for representative PCR-RFLP analysis of the VDR SNPs.

Fig 2. Blood lead levels for all genotypes of VDR gene polymorphic variants (*BsmI*, *FokI*, *ApaI* and *TaqI*) in environmental lead exposed children (* = $p < 0.05$; ** = $p < 0.01$; nonparametric Wilcoxon / Kruskal-Wallis tests between aa genotype and other genotypes)

Table 1

Genotype and allele frequency of ALAD and VDR polymorphisms (*BsmI*, *FokI*, *Apal* and *TaqI*) in children from Kabwe, Zambia in comparison with quartiles of BLLs

Gene	Genotype	N (%)	HWE equation ^a	Quartiles of BLL ^b				<i>p</i> -Value *
				χ^2 , <i>p</i> -value	Q1, N (%)	Q2, N (%)	Q3, N (%)	
ALAD	ALAD 1-1	139 (100)		31 (100)	35 (100)	31 (100)	42 (100)	
	ALAD 1-2	0						
	ALAD 2-2	0						
	G	1						
	C	0						
<i>BsmI</i> G > A	BB (AA)	3 (2.2)	1.70, 0.23	1 (3.2)	1 (2.9)	0	1 (2.2)	0.91
	Bb (AG)	49 (35.3)		9 (29.1)	13 (37.1)	13 (41.9)	14 (33.3)	
	bb (GG)	87 (62.6)		21 (67.7)	21 (60.0)	18 (58.1)	27 (64.3)	
	B (A)	0.2						
	b (G)	0.8						
<i>FokI</i> T > C	FF (CC)	92 (66.2)	0.01, 0.89	20 (64.5)	24 (68.6)	20 (64.5)	28 (66.7)	0.89
	Ff (CT)	42 (30.2)		10 (32.3)	11 (31.4)	9 (29.0)	12 (28.6)	
	ff (TT)	5 (3.6)		1 (3.2)	0	2 (6.5)	2 (4.8)	
	F (C)	0.81						
	f (T)	0.19						
<i>Apal</i> G > T	AA (TT)	63 (45.3)	0.04, 0.92	12 (38.7)	16 (45.7)	17 (54.8)	18 (42.9)	0.03*
	Aa (TG)	62 (44.6)		19 (61.3)	17 (48.6)	11 (35.5)	15 (35.7)	
	aa (GG)	14 (10.1)		0	2 (5.7)	3 (9.7)	9 (21.4)	
	A (T)	0.68						
	a (G)	0.32						
<i>TaqI</i> T > C	TT (TT)	75 (54.0)	2.62, 0.14	14 (45.2)	20 (57.1)	12 (38.7)	29 (69.1)	0.18
	Tt (TC)	59 (42.4)		16 (51.6)	13 (37.1)	18 (58.1)	12 (28.6)	
	tt (CC)	5 (3.6)		1 (3.2)	2 (5.7)	1 (3.2)	1 (3.6)	
	T (T)	0.75						
	t (C)	0.25						

N: number; HWE: Hardy-Weinberg equilibrium

* Chi-square test

^a HWE equation: χ^2 (chi-squared test) < 3.841 and/or *p*-value > 0.05 indicates no deviation from HWE

^b BLL quartiles (Q1: $x \leq 10 \mu\text{g/dL}$; Q2: $10 < x \leq 19 \mu\text{g/dL}$; Q3: $19 < x \leq 25 \mu\text{g/dL}$; Q4: $x > 25 \mu\text{g/dL}$)

Table 2

Comparison of VDR SNPs allele frequencies between Zambian children and different populations

Population	<i>FokI</i> (rs2228570)	<i>BsmI</i> (rs1544410)	<i>ApaI</i> (rs7975232)	<i>TaqI</i> (rs731236)	Reference
	C/T	G/A	T/G	T/C	
Zambian children	0.81/0.19	0.8/0.20	0.68/0.32	0.75/0.25	This study
African ancestry	0.83/0.17	0.71/0.29	0.62/0.38	0.75/0.25	Lins et al., 2011
European ancestry	0.52/0.48***	0.52/0.48***	0.57/0.43	0.52/0.48**	"
Asian ancestry	0.64/0.36**	0.92/0.08**	0.35/0.65***	0.93/0.07***	"
Brazil	0.67/0.33*	0.60/0.40**	0.54/0.46*	0.62/0.38	"
Black South Africans	0.84/0.16	0.74/0.26	0.76/0.24	0.66/0.34	Meyer et al., 2017
White South Africans	0.54/0.46***	0.62/0.38**	0.54/0.46*	0.61/0.39*	"
African American	0.77/0.23	0.63/0.37**	0.70/0.30	0.73/0.27	Sarkissyan et al., 2014
China	0.60/0.40**	0.91/0.08**	0.35/0.65***	0.90/0.10***	Yu et al., 2017
* = p < 0.05; ** = p < 0.01; *** = p < 0.001; Two-tailed Fisher's exact test and chi square test was done using Zambian children as reference value					

Table 3

Blood lead levels in environmentally exposed children with different genotypic combinations of VDR gene

Group	Genotype combination	BLL ($\mu\text{g/dL}$)			
		N	Mean \pm SD	<i>p</i> value	Range
G1	bbFfAATT	5	37.8 \pm 14.9	Ref	24.1 – 60.8
G2	bbFFAATT	6	23.4 \pm 10.5	NS	10.5 – 35.1
G3	bbFFAATt	12	19.8 \pm 10.2	*	4.7 – 42.2
G4	bbFFAaTT	20	18.0 \pm 11.3	*	4.5 – 41.8
G5	bbFFaaTT	11	24.6 \pm 7.9	NS	11.1 – 34.8
G6	bbFFAaTt	11	12.7 \pm 6.3	**	4.9 – 24
G7	bbFfAaTT	8	15.2 \pm 8.5	**	3.4 – 26.2
G8	bbFfAATt	4	14.2 \pm 10.1	*	3.9 – 24.5
G9	bbFfAaTt	4	18.8 \pm 18.4	NS	5.1 – 45.8
G10	BbFFAATT	4	22.9 \pm 13.4	NS	4.4 – 33.5
G11	BbFFAaTT	6	21.0 \pm 11.4	NS	10.7 – 40.4
G12	BbFFAATt	9	17.6 \pm 6.4	**	5.9 – 23.2
G13	BbFFAaTt	5	20.9 \pm 9.1	NS	8.8 – 30.4
G14	BbFFAAtt	5	17.9 \pm 10.7	NS	4.7 – 34.1
G15	BbFfAATT	6	13.9 \pm 2.9 ^b	**	10.8 – 19.2
G16	BbFfAaTT	3	16.7 \pm 12.8	NS	1.65 – 25.1
G17	BbFfAATt	6	17.2 \pm 12.6	NS	1.65 – 33.5
G18	BbFfAaTt	3	17.3 \pm 6.1	*	10.2 – 21.2
G19	bbFfaaTT	2	30.3 \pm 6.6		25.6, 34.9
G20	BbffAATt	2	29.8 \pm 6.4		25.3, 34.3
G21	bbffAaTT	1	6.7		
G22	bbffAATt	1	24.8		
G23	bbffAaTt	1	19.3		
G24	bbFFaaTT	1	32.4		
G25	BBFFAATT	1	7.6		
G26	BBFFAATt	1	26.6		
G27	BBFfAATT	1	17		

N: Number of children; BLL: Blood lead level; Ref: Reference; NS: Not significant
* = $p < 0.05$; ** = $p < 0.01$; Nonparametric Wilcoxon analysis for comparison for each pair for sample numbers with greater than three

Table 4

Distribution of VDR genotypes in low vs high BLL subjects and associations with lead toxicity

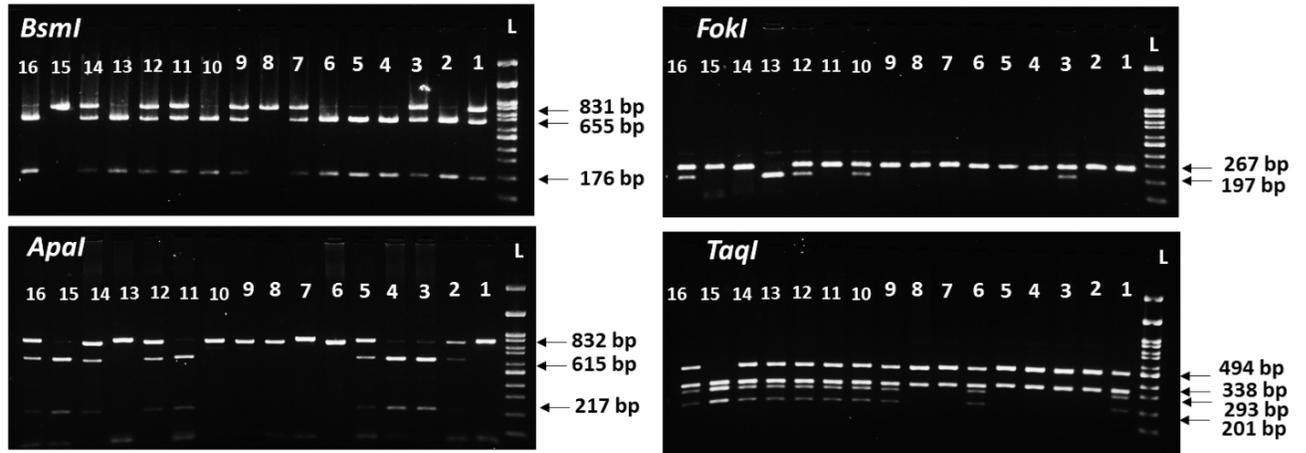
SNP	Genotype / Allele	Low BLL children [#] (N = 31)		High BLL children [#] (N = 108)		HWE equation χ^2, p	OR [95% CI], p -value
		N	%	N	%		
BsmI	BB	1	3.2	2	1.9	0.81, 0.66	(BB + Bb) vs bb
	Bb	9	29	40	37		1.33 [0.57 – 3.12], 0.50
	bb	21	67.7	66	61.1		
	B	11	18	44	20	1.19 [0.57 – 2.46], 0.64	
	b	51	82	172	80	1	
FokI	FF	20	64.5	72	66.7	0.08, 0.95	FF vs (Ff + ff)
	Ff	10	32.3	32	29.6		1.1 [0.47 – 2.54], 0.82
	ff	1	3.2	4	3.7		
	F	50	81	176	82	1.05 [0.52 – 2.16], 0.88	
	f	12	19	40	18	1	
ApaI	AA	12	38.7	51	47.2	6.89, 0.03*	AA vs (Aa + aa)
	Aa	19	61.3	43	39.8		1.42 [0.63 – 3.20], 0.40
	aa	0	0	14	13		
	A	43	69	145	67	0.90 [0.49 – 1.66], 0.74	
	a	19	31	71	33	1	
TaqI	TT	14	45.2	61	56.5	1.37, 0.50	TT vs (Tt + tt)
	Tt	16	51.6	43	39.8		1.58 [0.71 – 3.52], 0.26
	tt	1	3.2	4	3.7		
	T	44	71	165	76	1.32 [0.70 – 2.49], 0.38	
	t	18	29	51	24	1	

N: Number of children; %: Frequency; BLL: Blood lead level; HWE: Hardy-Weinberg equilibrium
OR: Odds ratios; 95% CI: 95% Confidence interval
[#] Low BLL children: BLL ≤ 10 µg/dL; High BLL children: BLL > 10 µg/dL
* Result is significant at $p < 0.05$

Table 5

Genotype frequencies of VDR SNPs and association with gender

SNP/ Genotype	All children		HWE equation χ^2, p	OR [95% CI], <i>p</i> -value Female vs Male
	Female N (%)	Male N (%)		
<i>BsmI</i>	BB	1 (1.5)	2 (3)	0.33, 0.84 (BB + Bb) vs bb 1.08 [0.54 – 2.16], 0.81
	Bb	24 (37)	25 (34)	
	bb	40 (61.5)	47 (63)	
<i>FokI</i>	FF	44 (68)	48 (65)	2.39, 0.30 FF Vs (Ff + ff) 1.13 [0.56 – 2.29], 0.72
	Ff	21 (32)	21 (28)	
	ff	0 (0)	5 (7)	
<i>Apal</i>	AA	29 (45)	34 (46)	0.16, 0.92 AA vs (Aa + aa) 0.94 [0.48 – 1.85], 0.87
	Aa	30 (46)	32 (43)	
	aa	6 (9)	8 (11)	
<i>TaqI</i>	TT	41 (63)	34 (46)	10.1, 0.006** TT vs (Tt + tt) 2.01 [1.02 – 3.97], 0.04*
	Tt	19 (29)	40 (54)	
	tt	5 (8)	0 (0)	
N: Number of children; %: Frequency; HWE: Hardy-Weinberg equilibrium OR: Odds ratios; 95% CI: 95% Confidence interval * Result is significant at $p < 0.05$				



L, 100 bp DNA ladder; 1-16, sample #.

For *BsmI*: #8 and 15 (831 bp) are BB homozygotes; #1, 3, 7, 9, 11, 12 & 14 (831, 655 and 176 bp) are Bb heterozygotes; #2, 4-6, 10, 13 & 16 (655 and 176 bp) are bb homozygotes. For *FokI*: #1, 2, 4-9, 11, 14 & 15 (267 bp) are FF homozygotes; #3, 10, 12 & 16 (267 and 197 bp) are Ff heterozygotes; #13 (197 and 70 bp) is ff homozygote. For *ApaI*: #1, 6-10 & 13 (832 bp) are AA homozygotes; #2, 12, 14 & 16 (832, 615 and 217 bp) are Aa heterozygotes; #3, 4, 11 & 15 (615 and 217 bp) are aa homozygotes. For *TaqI*: #2-5, 7 & 8 (494 and 338 bp) are TT homozygotes; #1, 6, 9-14 & 16 (494, 338, 293 and 201 bp) are Tt heterozygotes; #15 (338, 293 and 201 bp) is tt homozygote.

Fig. 1.

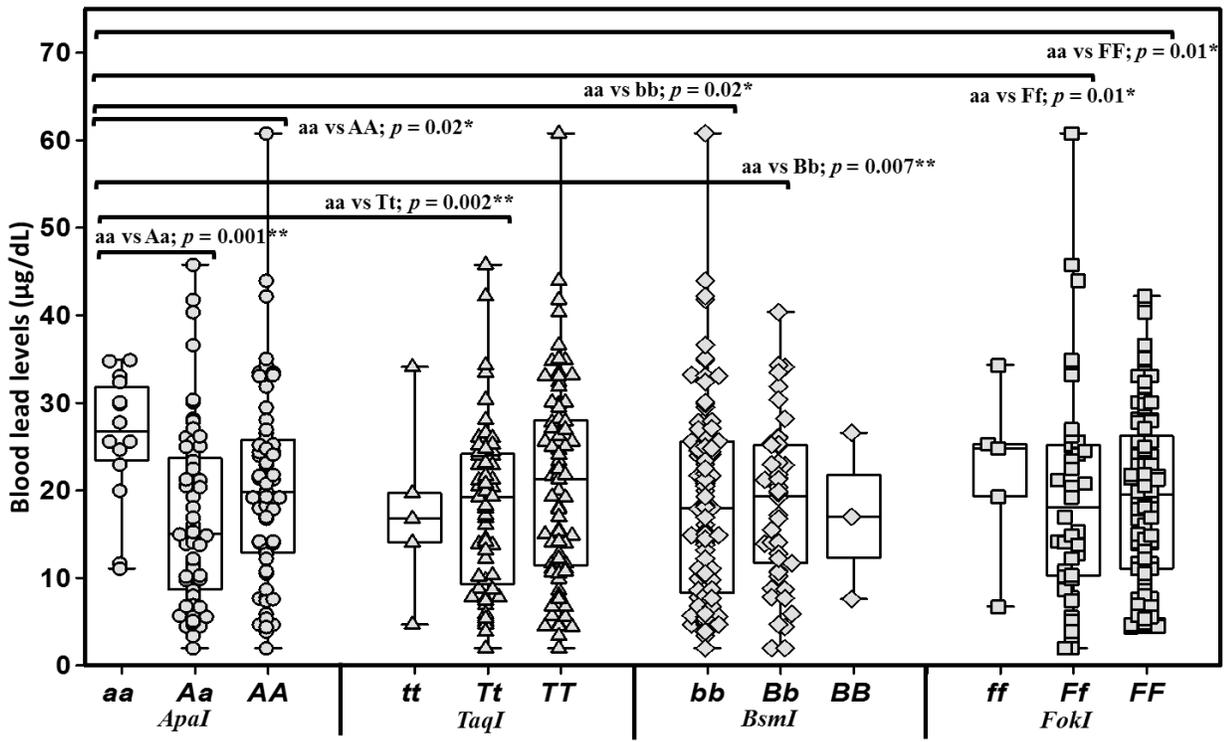


Fig 2.